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# Contribution of winter fluxes to the annual CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O emissions from freshwater marshes in the Sanjiang Plain

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**Abstract:** Wetlands at the interface of the terrestrial and aquatic ecosystems are intensive sites for mineralization of organic matter, but the contribution of winter season fluxes of CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O from wetland ecosystems to annual budgets is poorly known. By using the static opaque chamber and GC techniques, fluxes of CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O at two freshwater marshes in the Sanjiang Plain were measured during the winter seasons of 2002/2003 and 2003/2004 with contrasting snow conditions and flooding regimes. The results showed that there were significant interannual and spatial differences in CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O fluxes. The *Carex lasiocarpa* marsh emitted more CH<sub>4</sub> and CO<sub>2</sub> while absorbed less N<sub>2</sub>O than the *Deyeuxia angustifolia* marsh during the winter seasons. Over the winter season, emissions of CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O ranged from 0.42 to 2.41 gC/m<sup>2</sup>, from 24.13 to 50.16 gC/m<sup>2</sup>, and from -25.20 to -148.96 mgN/m<sup>2</sup>, respectively. The contributions of winter season CH<sub>4</sub> and CO<sub>2</sub> emission to the annual budgets were 2.32%—4.62% and 22.17%—27.97%, respectively. Marshes uptake N<sub>2</sub>O during the freezing period, while release N<sub>2</sub>O during the thawing period. The winter uptake equaled to 13.70%—86.69% of the growing-season loss. We conclude that gas exchange between soil/snow and the atmosphere in the winter season contributed greatly to the annual budgets and cannot be ignored in a cool temperate freshwater marsh in Northeast China.

**Keywords:** freshwater marsh; winter flux; greenhouse gas emission; Sanjiang Plain

## Introduction

Carbon dioxide, methane and nitrous oxide play important roles in the radiation balance of the earth contributing to the greenhouse effect (Rodhe, 1990). N<sub>2</sub>O also takes part in the destruction of stratospheric ozone (Wang, 1999). Natural wetlands are globally important sources for atmospheric CH<sub>4</sub> and sinks for atmospheric CO<sub>2</sub> (Tyler, 1991). In the past decades, the processes of soil freezing and thawing have been identified as important sources of trace gas emissions from soils (Panikov and Dedysh, 2000; Teepe *et al.*, 2001). Some findings show that carbon sequestration could be overestimated when the winter fluxes are not included in budget calculations (Charlotte and Nigel, 2003). It is therefore crucial to take the trace gas emissions during the cold season into account when one tries to accurately estimate the emissions of CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O from wetlands.

Wetlands are the interface between terrestrial and aquatic components of the landscape. These areas have unique characteristics due to the mixture of terrestrial and aquatic factors that influence their structure and function (Yang, 2002). Wetland ecosystems have received attention in several contexts including biodiversity, water quality maintenance and habitat protection (Lu, 2002). The area of mires in the Sanjiang Plain is about  $1.04 \times 10^4$  km<sup>2</sup>, being second only to the Qinghai-Tibet Plateau in China (Zhao, 1999). Marshes in this area are mainly divided into

four types according to plant type, viz. *Carex lasiocarpa*, *Carex pseudocuraica*, *Carex meyeriana* and *Deyeuxia angustifolia* (Zhao, 1999). Over the last several years, scientists have made considerable efforts on quantifying the emissions of CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O in the plant-growing seasons (Ding *et al.*, 2004; Song *et al.*, 2003), while few studies have been dedicated to the emissions during winter season and the periods of freezing and thawing. The objective of this paper is to determine the magnitude and the significance of the winter CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O fluxes to the net annual emissions from marsh in the Sanjiang Plain, Northeast China.

## 1 Materials and methods

### 1.1 Study sites

Filed observations were made in the Sanjiang Mire Wetland Experimental Station (47°35' N, 133°31' E), Chinese Academy of Sciences, located in the eastern part of Heilongjiang province, China. More detailed information of the site was described by Ding *et al.* (2004).

From the edge to the center of the billabong, the types of vegetation vary from *D. angustifolia* to *C. lasiocarpa* as standing water depth increases. Vegetations of *C. lasiocarpa* and *D. angustifolia* were selected, which are continuously flooded and seasonal flooded marshes, respectively. Characteristics of the marsh soils in the experimental plots are listed in Table 1.

**Table 1** Characteristics of the marsh soil in the experimental plots

Vegetation	Depth, cm	Organic carbon, g/kg	Total N, g/kg	Bulk density, g/cm <sup>3</sup>
<i>C. lasiocarpa</i>	0—5	321.8 ± 19.1	14.7 ± 2.3	0.35 ± 0.08
	5—10	266.4 ± 38.4	12.5 ± 1.7	0.54 ± 0.13
<i>D. angustifolia</i>	0—5	148.1 ± 51.0	9.5 ± 2.5	0.56 ± 0.21
	5—10	76.3 ± 7.2	8.2 ± 2.5	0.79 ± 0.17

## 1.2 Measurements of CH<sub>4</sub>, N<sub>2</sub>O and CO<sub>2</sub> fluxes

Three neighboring plots at each marsh were subjectively established to permanently measure CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O emission rates. Wooden boardwalks were installed for minimizing the disturbance during sampling. Three stainless steel bases (0.5 m × 0.5 m × 0.2 m) with a water groove on top were installed in the soil prior to sampling and were kept throughout the experimental period. Gas samples were taken between 0900 and 1100 LST. During sampling, an open-bottom stainless steel chamber (0.5 m × 0.5 m × 0.5 m, equipped with two fans) was placed over the base and filled with water in the groove to ensure tightness. Air sample inside the chamber was taken for every 10 min over a 30-min period by using 60 ml plastic syringes (total of four samples). Air temperature inside the chamber and atmospheric pressure were measured during taking samples. Concentrations of CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O were simultaneously analyzed in the laboratory (within a period of at most 12 h) using a gas chromatograph (Wang and Wang, 2003). The rates of CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O emission were calculated from a linear regression of the changes in the concentration. Corrections were made for air temperature and pressure. The data deviating significantly from linearity ( $r^2 < 0.95$  for CH<sub>4</sub> and CO<sub>2</sub>,  $r^2 < 0.90$  for N<sub>2</sub>O) were discarded.

The gas sample was taken every other day during the thawing period (April—June), twice a week during the growing season (May—October) and once or twice a month in the winter (November—March next year). Daily fluxes were estimated by a linear interpolation for the days without measurements. The total amounts of annual and winter-season CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O emissions from both sites were computed from the daily fluxes.

Data analyses were performed with SPSS 11.0 and Microsoft Excel for Windows 2000. The tested factors and the hypothesized relationship were considered statistically significant when  $p < 0.05$ .

## 2 Results and discussion

### 2.1 Temperature and precipitation

In the Sanjiang Plain, the marshes are generally ice covered from early November to late March next year. We define the ice-covered period as winter

season. The mean air temperatures in the winter seasons of 2002/2003 and 2003/2004 were -15.1 and -14.2 °C, respectively. One-way ANOVA indicated that there was no significant difference in the air temperatures between the two winter seasons ( $p = 0.82$ ). The average air temperature in January was the lowest, -20.3 °C and -21.0 °C in 2003 and in 2004, respectively. The air temperature from April onward was above zero and the frozen soil began to thaw.

About 10% of the annual precipitation falls as snow in this region. In the winter season of 2002/2003, the precipitation was only 20.0 mm and less than the long-term average. In contrast, the precipitation in the winter season of 2003/2004 was 91.7 mm and much higher than the long-term average. The difference in precipitation between the two winter seasons was significant ( $p = 0.03$ ). Winter season of 2003/2004 was warmer and wetter by contrast with that of 2002/2003. During the soil thawing period, the monthly mean air temperature in 2004 was lower while total amount of monthly precipitation was higher than that in 2003.

### 2.2 CH<sub>4</sub> emission

CH<sub>4</sub> emission from the soil/snow to the atmosphere during the entire winter period is given in Fig. 1. At the *C. lasiocarpa* site, CH<sub>4</sub> fluxes were slightly lower in the winter season of 2002/2003 than those in 2003/2004. Average CH<sub>4</sub> fluxes in these two seasons were 0.32 and 0.89 mg/(m<sup>2</sup>·h), respectively. On the contrary, CH<sub>4</sub> fluxes in the winter season of 2002/2003 at the *D. angustifolia* site were slightly higher than those in the winter season of 2003/2004. Average CH<sub>4</sub> fluxes were 0.25 and 0.15 mg/(m<sup>2</sup>·h), respectively. The spatial averages ± standard deviations of methane emission flux were  $0.29 \pm 0.18$  and  $0.52 \pm 0.29$  mg/(m<sup>2</sup>·h) during the two respective winter seasons in the Sanjiang Plain. These values fall among the results reported from other wetland ecosystems. Northern Minnesota bogs and fens emitted 0.19—1.29 mg/(m<sup>2</sup>·h) during this time of

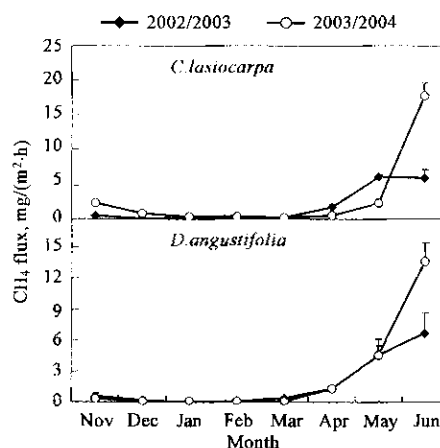


Fig. 1 CH<sub>4</sub> flux from *C. lasiocarpa* and *D. angustifolia* marshes during freezing and thawing periods

winter (Dise, 1993). In West Siberia (Tomsk), a smaller winter CH<sub>4</sub> flux ( $0.21 \pm 0.15 \text{ mg}/(\text{m}^2 \cdot \text{h})$ ) was found and the average CH<sub>4</sub> fluxes for five winter seasons varied between 0.83 and 2.50 mg/(m<sup>2</sup>·h) in New Hampshire (Panikov and Dedysh, 2000; Melloh and Crill, 1996).

The cumulative CH<sub>4</sub> emissions from the *C. lasiocarpa* site were 0.87 and 2.41 gC/m<sup>2</sup> in the two winter seasons, respectively, and those from the *D. angustifolia* site were 0.70 and 0.42 gC/m<sup>2</sup>, respectively. The winter season CH<sub>4</sub> flux was about 2.32%—4.62% of the annual flux in both marshes, which is similar to the lower results reported from other wetland ecosystems. For example, winter season CH<sub>4</sub> emission was 4%—21% of annual fluxes in Minnesota, 2%—10% in New Hampshire and 3.5%—11% in West Siberia (Dise, 1993; Panikov and Dedysh, 2000; Melloh and Crill, 1996).

CH<sub>4</sub> fluxes increased rapidly during the thawing period. There was a significant interannual difference in spring (through April to May) CH<sub>4</sub> release at the *C. lasiocarpa* site. The CH<sub>4</sub> released in 2004 was only about 37% of that in 2003 spring season. A study performed in boreal lakes during three consecutive winters showed that much greater snowfall in the winter reduced enormously the spring CH<sub>4</sub> release (Tuula *et al.*, 2004). The results of our study are consistent with their report.

### 2.3 CO<sub>2</sub> emission

Winter season CO<sub>2</sub> emission (Fig. 2) was higher than CH<sub>4</sub>. Significant difference in CO<sub>2</sub> flux ( $p=0.014$ ) from the *C. lasiocarpa* site existed between the two winter seasons. The average CO<sub>2</sub> flux from the *C. lasiocarpa* marsh was  $10.55 \pm 1.71 \text{ mg}/(\text{m}^2 \cdot \text{h})$  in the winter season of 2002/2003 and  $44.66 \pm 5.99 \text{ mg}/(\text{m}^2 \cdot \text{h})$  in the 2003/2004, and  $25.63 \pm 4.11$  and  $26.52 \pm 3.76 \text{ mg}/(\text{m}^2 \cdot \text{h})$  from the *D. angustifolia* marsh in the two respective winter seasons. Note that most of the CO<sub>2</sub> fluxes during the winter season of 2002/2003 from the *C. lasiocarpa* marsh were below the detection limit of the gas chromatography system (Wang, 2003), we suspect that the values were erroneous and problematic, so they were not used in the determination of the annual carbon budget. The spatial averages  $\pm$  standard deviations of CO<sub>2</sub> flux were  $25.63 \pm 4.11$  and  $35.59 \pm 4.88 \text{ mg}/(\text{m}^2 \cdot \text{h})$  during the two winter seasons in the Sanjiang Plain. For comparison, the CO<sub>2</sub> flux in winter season was 22.92 mg/(m<sup>2</sup>·h) in the Kolyma basin, Northeast Siberia (Zimov *et al.*, 1993), 45.83 and 12.22 mg/(m<sup>2</sup>·h) from moist tussock tundra and coastal wet sedge ecosystems in the Alaskan Arctic respectively (Oechel *et al.*, 1997, 2000), and  $43.2 \pm 10.8 \text{ mg}/(\text{m}^2 \cdot \text{h})$  in the Mirabel peatland, Quebec, Canada (Charlotte and Nigel, 2003).

Some findings show that carbon sequestration

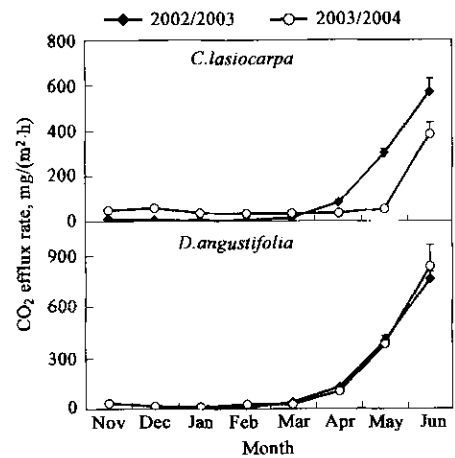


Fig. 2 CO<sub>2</sub> efflux rate from *C. lasiocarpa* and *D. angustifolia* marshes during freezing and thawing periods

could be overestimated if the cold season fluxes are not included in budget calculations (Charlotte and Nigel, 2003). In our study, the measured CO<sub>2</sub> emission using a dark chamber represents the ecosystem respiration rate, including nocturnal respiration of plant shoots and whole-day metabolism of all below-ground organisms, roots, microbes, and animals, as well as any other soil reactions degrading organic matter to CO<sub>2</sub>. In winter, the CO<sub>2</sub> emission rate is equal to CO<sub>2</sub> flux owing to the lack of plant photosynthesis. Indeed, the net annual CO<sub>2</sub> flux is accounted for as a difference:

$$\text{Net C} - \text{flux} = \text{RH} - \text{NPP}$$

where RH and NPP stand for soil heterotrophic respiration and net primary production of the ecosystem, respectively. In the years of 2002—2004, we measured the CO<sub>2</sub> emission rates of bare soil (RH) and the plant biomass during the growing season (NPP) in the two marshes for each year (data not shown). The results of winter CO<sub>2</sub> flux, total annual respiration and net annual CO<sub>2</sub> flux are shown in Table 2.

The results demonstrated that winter CO<sub>2</sub> efflux comprised 2.60%—8.51% of the annual respiration and 22.17%—27.97% of the annual C budget. Current estimates of the winter CO<sub>2</sub> contribution to the annual C budget, based on nonconcurrent measurements of summer and winter exchange, are 21% in a Finnish mixed mire (Alm *et al.*, 1999), 25% in an Alaskan forest (Sommerfeld *et al.*, 1996) and 17% in an alpine wetland (Mast *et al.*, 1998).

*C. lasiocarpa* marsh emits more CH<sub>4</sub> and CO<sub>2</sub> in the winter than *D. angustifolia* marsh. Perhaps this results from the deeper upper organic layers of *C. lasiocarpa* marsh, which delays the time of soil freezing and reduces the depth of frost penetration.

**Table 2 Contribution of winter to the annual CO<sub>2</sub> emissions and net annual C budget in the marshes, northeast China**

Vegetation	Winter season	C emission, gC/m <sup>2</sup>	Annual total respiration, gC/m <sup>2</sup> (winter of annual, %)	Annual net C budget, gC/m <sup>2</sup> (winter of annual, %)
<i>C. lasiocarpa</i>	2002/2003	ND	506.97	-278.26
	2003/2004	50.16	589.79 (8.51)	-226.27 (22.17)
<i>D. angustifolia</i>	2002/2003	24.13	745.22 (3.24)	ND
	2003/2004	26.05	1000.12 (2.60)	-93.12 (27.97)

Note: Minus indicates net uptake of CO<sub>2</sub>; ND, no data

In general, there are two major sources of C emission during wintertime, the CO<sub>2</sub> and CH<sub>4</sub> produced *de novo* in winter by psychrophilic microorganisms (Flanagan and Bunnell, 1980), and the stored methane and carbon dioxide formed by mesophilic microbes during the warm period (Dise, 1993). In the study by Stadler (1996), 8%—20% of the soil water was not frozen although the soil temperature was -5°C for several days. In the liquid water of frozen soils, the availability of labile C may be high as a consequence of microorganisms being killed by freezing or hygroscopic effects and organic matter in broken aggregates (Christensen and Tiedje, 1990; Christensen and Christensen, 1991). This creates a favorable condition for the activity of microorganisms and part of the trapped gas may escape through the frost-induced cracks. Panikov *et al.* (2000) studied winter CO<sub>2</sub> and CH<sub>4</sub> emissions in West Siberia and found that cold-resistant microorganisms located within the frozen peat layers were still active at -16°C and could provide the observed instant winter CO<sub>2</sub> emission. In the Sanjiang Plain, according to the *in situ* observation for three years, the lowest temperatures of the marsh topsoil and subsoil (10—30 cm) were about -13°C and -4 — -7°C, respectively. So we suggest that at least part of the C emission during the winter season in the Sanjiang Plain might be attributed to microorganism activities.

We found a significantly positive correlation between the emission rates of CH<sub>4</sub> and CO<sub>2</sub> ( $p=0.01$ ) during the winter season. This provides an indication that there should be common mechanisms controlling emissions of both gas species under the cold non-growing conditions. Such mechanisms may include the diffusion of gases through snow (Sommerfeld *et al.*, 1996).

Contrary to the winter period, CH<sub>4</sub> fluxes and CO<sub>2</sub> emission rates both correlated with air temperatures during the thawing period. But the coefficients in the Equations are somewhat different in different marshes (Table 3). This is perhaps due to different hydrological conditions. On the basis of experiments and observations in wetland ecosystems, many researchers have pointed out that the relationship between CH<sub>4</sub> flux and soil temperature can be described by the Arrhenius equation (Shangguan *et al.*, 1993; Huang *et al.*, 2001) as

follows:

$$F=A \cdot e^{-Ea/(R \cdot T)}$$

where  $F$  is CH<sub>4</sub> flux,  $Ea$  is apparent activation energy,  $R$  is the gas constant,  $T$  is the soil temperature (K), and  $A$  is an empirical coefficient. The change of CH<sub>4</sub> flux with temperature can also be expressed by the term of  $Q_{10}$  that denotes the factor of CH<sub>4</sub> flux increase with a soil temperature increase of 10°C. Generally, the  $Q_{10}$  of CH<sub>4</sub> flux is 2—3 (Khalil *et al.*, 1998). But during the thawing period, the very steep temperature response of microbial metabolism, which is especially high for methanogenic bacteria (the apparent values of  $Q_{10}$  are 5.79 and 9.12 in the two marshes, respectively) means that temperature not only stimulates microorganism activity, but also greatly improves the gas emission in soil. During the thawing period, the relationship between CO<sub>2</sub> release rates and air temperatures can also be described by the Arrhenius equation. The  $Q_{10}$  values are 5.21 and 5.44 in the two marshes respectively, while it is only 1.43 in the summer season and 3.93 in the autumn season. Obviously, the increase in air temperature stimulates more CO<sub>2</sub> production and emission during the thawing period. But compared to CH<sub>4</sub>, CO<sub>2</sub> is less dependent upon temperature. This agrees with the results from a laboratory study that shows that spring thaw should considerably activate gas production, especially for the anaerobic formation of CH<sub>4</sub> (Panikov and Dedysh, 2000).

**Table 3 Relationship between CH<sub>4</sub> fluxes, CO<sub>2</sub> release rates and air temperature during thawing periods**

Vegetation	CH <sub>4</sub>	CO <sub>2</sub>
<i>C. lasiocarpa</i>	$Y = 0.26e^{0.18x}$ ( $p < 0.001$ )	$Y = 10.73e^{0.15x}$ ( $p < 0.001$ )
	$r^2 = 0.65, Q_{10} = 5.79$	$r^2 = 0.78, Q_{10} = 5.21$
<i>D. angustifolia</i>	$Y = 0.43e^{0.22x}$ ( $p < 0.001$ )	$Y = 61.66e^{0.17x}$ ( $p < 0.001$ )
	$r^2 = 0.54, Q_{10} = 9.12$	$r^2 = 0.66, Q_{10} = 5.44$

Notes:  $Y$ , CH<sub>4</sub> flux or CO<sub>2</sub> emission rate;  $x$ , air or soil temperature

There is a distinct difference between the CO<sub>2</sub> emissions from the two marshes in the freezing and thawing periods. During the freezing periods, the difference is very small ( $p > 0.05$ ), but during the thawing periods, CO<sub>2</sub> emission from the *D. angustifolia* marsh is significantly higher than that from the *C. lasiocarpa* marsh ( $p < 0.01$ ). This is mainly

attributed to the differences of the hydrological conditions and plant growth.

#### 2.4 N<sub>2</sub>O emission

Higher N<sub>2</sub>O emission was found during the soil freezing period in some studies (Papen and Butterbach, 1999; Teepe *et al.*, 2000). However, the measurements of this study indicated that the two marshes absorbed atmospheric N<sub>2</sub>O in the two winter seasons (Fig.3). Perhaps this discrepancy in the research is due to the different ecosystems. The previous studies of winter N<sub>2</sub>O flux have mainly focused on agriculture and forest ecosystem and the researchers have held the opinion that soil freezing can create an anaerobic environment for N<sub>2</sub>O production (Papen *et al.*, 1999; Teepe *et al.*, 2000). The mechanism of N<sub>2</sub>O absorption by marsh in the two winter seasons probably resulted from frozen soil absorbing atmospheric N<sub>2</sub>O through frost-induced cracks and perhaps the anaerobic environment was too strong to reduce N<sub>2</sub>O to N<sub>2</sub> by reductase enzymes. In the winter seasons of 2002/2003 and 2003/2004, the cumulative N<sub>2</sub>O emissions were -29.49 and -25.20 mgN/m<sup>2</sup> from the *C. lasiocarpa* site, and -148.96 and -34.33 mgN/m<sup>2</sup> from the *D. angustifolia* site, while the marshes became a source of N<sub>2</sub>O during the plant-growing season. Correspondingly, the contributions of winter uptake to the growing-season emission were 37.09%, 13.83%, 86.69% and 13.70%, respectively. The N<sub>2</sub>O flux at the *D. angustifolia* site exhibited remarkable variability between the two winter seasons ( $p < 0.001$ ). This large interannual difference may be ascribed to the difference of precipitation. Precipitation in the winter season of 2003/2004 was significantly higher than that in the 2002/2003, so the snow cover on the soil during the former was much deeper than that during the latter. Deeper snow might baffle the absorption of atmospheric N<sub>2</sub>O and then reduced the magnitude of the soil sink for atmospheric N<sub>2</sub>O.

Nitrous oxide flux in the winter season also showed an evident difference between the two marshes. The amount of N<sub>2</sub>O uptake at the *D. angustifolia* site was much higher than that at the *C. lasiocarpa* site ( $p = 0.005$ ). It is probable that N<sub>2</sub>O fluxes were related to the actual density of frost-induced cracks on the ice or frozen surface soil in the measurement plot because it were cracks that provided a route for gas entry. Some studies showed that there existed significant power correlation between the volume of crack and the daily N<sub>2</sub>O emission flux in rice field (Huang *et al.*, 2004). The shallower upper organic layer of the *D. angustifolia* site advanced the time of soil freezing and increased the area and density of cracks.

Both marshes were transformed from sinks into sources of atmospheric N<sub>2</sub>O as the temperature rise

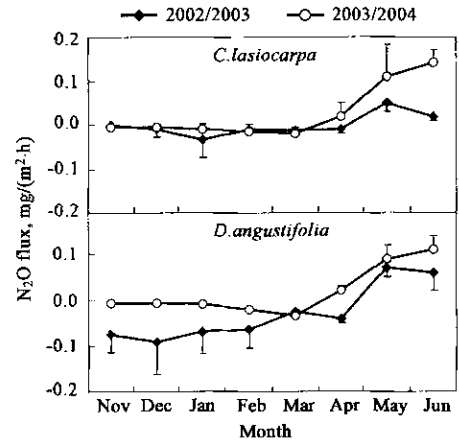


Fig.3 N<sub>2</sub>O flux from *C. lasiocarpa* and *D. angustifolia* marshes during soil freezing and thawing periods

during the thawing periods. These periods were very important for N<sub>2</sub>O emission and accounted for nearly half of the annual N<sub>2</sub>O emissions, which agrees with the results reported from other ecosystems (Teepe *et al.*, 2000). Several studies indicated that the thawing period was important for N<sub>2</sub>O production and emission because the process of soil thawing could provide amounts of labile C and N to microorganisms, and thus stimulate N<sub>2</sub>O substantive production in the soil (Sommerfeld *et al.*, 1993). Studies in rice fields have shown that either the soil moisture or the water table was a crucial factor for N<sub>2</sub>O emission and that a deeper water table would restrict N<sub>2</sub>O formation and emission during the flooding period (Zheng *et al.*, 2000). Surprisingly and interestingly, the period of flooding was longer and the water table was deeper in 2004 than that in 2003, while N<sub>2</sub>O fluxes were much higher in 2004 than that in 2003. The reason is still unknown and a further investigation is necessary.

Considerable investigations demonstrated that process of soil thawing enhanced CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O emissions (Oeclel *et al.*, 1997; Teepe *et al.*, 2001; Tuula *et al.*, 2004). The possible explanations for this phenomenon come from several aspects, including the release of the stored CH<sub>4</sub>, CO<sub>2</sub> and N<sub>2</sub>O in the soil below the frozen soil surface (Burton and Beauchamp, 1994), the increased C and N supply to microorganisms originating from organic matter in broken aggregates or from soil microbes killed by freeze/thaw cycles (Panikov and Dedysh, 2000), the creation of an anaerobic, water saturated topsoil by melting ice and snow in the frozen subsoil (Teepe *et al.*, 2001), and the temperature activation of dormant microbial cells and plant roots (Tuula *et al.*, 2004).

### 3 Conclusions

The results of this study indicated that the gas exchange between soil/snow and the atmosphere in the winter season contributed greatly to the annual budgets and cannot be ignored in a cool temperate

freshwater marsh in northeast China. Marshes uptake  $N_2O$  during the freezing period, while release  $N_2O$  during the thawing period. Significant temporal and spatial differences in the  $CH_4$ ,  $CO_2$  and  $N_2O$  fluxes existed in the winter season.

The mechanisms and processes of the production and emission of the trace gases during the winter season are complex in the freshwater marshes. To accurately quantify the contribution of winter season emissions of  $CH_4$ ,  $CO_2$  and  $N_2O$  to the annual budgets, it is crucial to pay more attention to the freezing and thawing actions.

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## References:

- Alm J, Saarnio S, Nykanen N *et al.*, 1999. Winter  $CO_2$ ,  $CH_4$  and  $N_2O$  fluxes on some natural and drained peatlands [J]. *Biogeochemistry*, 44: 163—186.
- Burton D L, Beauchamp E G, 1994. Profile nitrous oxide and carbon dioxide in a soil subject to freezing [J]. *Soil Science Society of America Journal*, 58: 115—122.
- Charlotte L R, Nigel T R, 2003. Seasonal contribution of  $CO_2$  fluxes in the annual C budget of a northern bog[J]. *Global Biogeochemical Cycles*, 17(1): 1029—1038.
- Christensen S, Christensen B T, 1991. Organic matter available for denitrification in different soil fractions: Effect of freeze/thaw cycles and straw disposal [J]. *Journal of Soil Science*, 42: 637—647.
- Christensen S, Tiedje J M, 1990. Brief and vigorous  $N_2O$  production by soil at spring thaw[J]. *Journal of Soil Science*, 41: 1—4.
- Ding W X, Cai Z C, Tsuruta H, 2004. Diel variation in methane emissions from the stands of *Carex lasiocarpa* and *Deyeuxia angustifolia* in a cool temperate freshwater marsh[J]. *Atmospheric Environment*, 38: 181—188.
- Dise N B, 1993. Methane emission from Minnesota peatlands: Spatial and seasonal variability [J]. *Global Biogeochemical Cycles*, 7(1): 123—142.
- Flanagan P W, Bunnell F L, 1980. Microfloral activities and decomposition, An Arctic Ecosystem [M]. In: *The coastal tundra of northern Alaska*(J. Brown *et al.* ed.). New York: van Nostrand Reinhold.
- Huang S H, Lu J, Zeng G H, 2004. The influence of generation of paddy soil cracks on  $N_2O$  emissions [J]. *China Environmental Science*, 24(4): 410—413.
- Huang Y, Jiang J Y, Zong L G *et al.*, 2001. Comparison of field measurements of  $CH_4$  emission from rice cultivation in Nanjing, China and in Texas, USA[J]. *Advances in Atmospheric Sciences*, 18(7): 1121—1130.
- Khalil M A K, Rasmussen R A, Shearer M J *et al.*, 1998. Factors affecting methane emissions from rice fields [J]. *Journal of Geophysical Research*, 103: 25219—25231.
- Lu X G, 2002. A review and prospect for wetland science[J]. *Journal of the Graduate School of the Chinese Academy of Sciences*, 3: 170—172.
- Mast M A, Wickland K P, Striegl R G, 1998. Winter fluxes of  $CO_2$  and  $CH_4$  from subalpine soils in the Rocky Mountain National Park, Colorado[J]. *Global Biogeochemical Cycles*, 12(4): 607—620.
- Melloh R A, Crill P M, 1996. Winter methane dynamics in a temperate peatland[J]. *Global Biogeochemical Cycles*, 10: 247—254.
- Oechel W C, Vourlitis G L, Hastings S J, 1997. Cold season  $CO_2$  emission from arctic soils[J]. *Global Biogeochemical Cycles*, 11: 163—172.
- Oechel W C, Vourlitis G L, Hastings S J *et al.*, 2000. Acclimation of ecosystem  $CO_2$  exchange in the Alaskan Arctic in response to decadal climate warming[J]. *Nature*, 406: 978—981.
- Panikov N S, Dedysh S N, 2000. Cold season  $CH_4$  and  $CO_2$  emissions from boreal peat bogs (West Siberia): Winter fluxes and thaw activation dynamics [J]. *Global Biogeochemical Cycles*, 14(4): 1071—1080.
- Papen H, Butterbach B K, 1999. 3-Year continuous record of N-trace gas fluxes from untreated and limed soil of a N-saturated spruce and beech forest in Germany: 1.  $N_2O$  emissions [J]. *Journal of Geophysical Research*, 104: 18487—18503.
- Rodhe H, 1990. A comparison of the contribution of the various gases to the greenhouse effect[J]. *Science*, 248: 1217—1219.
- Shangguan X J, Wang M X, Chen D Z *et al.*, 1993. Methane production in rice paddy fields[J]. *Advance in Earth Sciences*, 8: 1—12.
- Sommerfeld R A, Massman W J, Musselmann P J, 1996. Diffusional flux of  $CO_2$  through snow: Spatial and temporal variability among subalpine sites [J]. *Global Biogeochemical Cycles*, 10(3): 473—482.
- Sommerfeld R A, Mosier A R, Messelman R C, 1993.  $CO_2$ ,  $CH_4$  and  $N_2O$  flux through a Wyoming snowpack and implications for global budgets[J]. *Nature*, 361: 140—142.
- Song C C, Yan B X, Wang Y S, 2003. Fluxes of carbon and methane from swamp and impact factors in Sanjiang plain, China [J]. *Chinese Science Bulletin*, 48(24): 2749—2753.
- Stadler D, 1996. Water and solute dynamics in frozen forest soils—Measurements and modeling [J]. *Diss ETH Zürich*, 115: 74—79.
- Teepo R, Brumme R, Beese F, 2000. Nitrous oxide emissions from frozen soils under agricultural, fallow and forest land [J]. *Soil Biology & Biochemistry*, 32: 1807—1810.
- Teepo R, Brumme R, Beese F, 2001. Nitrous oxide emissions from soil during freezing and thawing periods [J]. *Soil Biology & Biochemistry*, 33: 1269—1275.
- Tuula L, Jukka A, Sari J *et al.*, 2004. Contribution of vegetated littoral zone to winter fluxes of carbon dioxide and methane from boreal lakes[J]. *Journal of Geophysical Research*, 109: 19102—19114.
- Tyler S C, 1991. The global methane budget [M]. In: *Microbial production and consumption of greenhouse gases: Methane, nitrous oxide and halomethanes* (Rogers J. E., Whitman W. B. ed.). Washington, DC: American Society for Microbiology.
- Wang M X, 1999. *Atmospheric chemistry* [M]. 2nd ed. Beijing: China Meteorological Press.
- Wang Y S, Wang Y H, 2003. Quick measurement of  $CH_4$ ,  $CO_2$  and  $N_2O$  emission from a short-plant ecosystem [J]. *Advances in Atmospheric Sciences*, 20(5): 842—844.
- Yang Y X, 2002. Main characteristics, progress and prospect of international wetland science research[J]. *Progress in Geography*, 21(2): 111—120.
- Zhao K Y, 1999. *Mires in China*[M]. Beijing: Science Press.
- Zheng X H, Wang M X, Wang Y S, 2000. Impacts of soil moisture on nitrous oxide emission from croplands: A case study on the rice-based agro-ecosystem in Southeast China [J]. *Chemosphere-Global Change Science*, 2: 207—224.
- Zimov S A, Zimova G M, Davidov S P *et al.*, 1993. Winter biotic activity and production of  $CO_2$  in Siberian soils: A factor in the greenhouse effect [J]. *Journal of Geophysical Research*, 98: 5017—5023.

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