

## CONTENTS

### Aquatic environment

- Performance and microbial diversity of an expanded granular sludge bed reactor for high sulfate and nitrate waste brine treatment  
Runhua Liao, Yan Li, Xuemin Yu, Peng Shi, Zhu Wang, Ke Shen, Qianqian Shi, Yu Miao, Wentao Li, Aimin Li ..... 717
- Pollutant removal from municipal wastewater employing baffled subsurface flow and integrated surface flow-floating treatment wetlands  
Tanveer Saeed, Abdullah Al-Muyeed, Rumana Afrin, Habibur Rahman, Guangzhi Sun ..... 726
- Removal of polycyclic aromatic hydrocarbons from aqueous solution by raw and modified plant residue materials as biosorbents  
Zemin Xi, Baoliang Chen ..... 737
- Hybrid constructed wetlands for highly polluted river water treatment and comparison of surface- and subsurface-flow cells  
Yucong Zheng, Xiaochang Wang, Jiaqing Xiong, Yongjun Liu, Yaqian Zhao ..... 749
- Minimization of methabenzthiazuron residues in leaching water using amended soils and photocatalytic treatment with TiO<sub>2</sub> and ZnO  
José Fenoll, Pilar Flores, Pilar Hellín, Joaquín Hernández, Simón Navarro ..... 757
- Enhanced struvite recovery from wastewater using a novel cone-inserted fluidized bed reactor  
Awoke Guadie, Siqing Xia, Wei Jiang, Lijie Zhou, Zhiqiang Zhang, Slawomir W. Hermanowicz, Xiaoyin Xu, Shuang Shen ..... 765
- Evaluating the effectiveness of marine actinobacterial extract and its mediated titanium dioxide nanoparticles in the degradation of azo dyes  
S Priyragini, S Veena, D Swetha, L Karthik, G Kumar, K V Bhaskara Rao ..... 775
- Effect of ozone on the performance of a hybrid ceramic membrane-biological activated carbon process  
Jianning Guo, Jiangyong Hu, Yi Tao, Jia Zhu, Xihui Zhang ..... 783
- Removal of perchlorate from aqueous solution by cross-linked Fe(III)-chitosan complex  
Long Lv, Yanhua Xie, Guoming Liu, Guo Liu, Jing Yu ..... 792

### Atmospheric environment

- Origin of major ions in monthly rainfall events at the Bamenda Highlands, NorthWest Cameroon  
Mengnjo J. Wirmvem, Takeshi Ohba, Wilson Y. Fantong, Samuel N. Ayonghe, Jonathan N. Hogarh, Justice Y. Suila, Asobo Nkengmatia E. Asaah, Seigo Ooki, Gregory Tanyileke, Joseph V. Hell ..... 801
- Ionic composition of submicron particles (PM<sub>1.0</sub>) during the long-lasting haze period in January 2013 in Wuhan, central China  
Hairong Cheng, Wei Gong, Zuwu Wang, Fan Zhang, Xinming Wang, Xiaopu Lv, Jia Liu, Xiaoxin Fu, Gan Zhang ..... 810
- Understanding the sources and composition of the incremental excess of fine particles across multiple sampling locations in one air shed  
Jerome E. McGinnis, Jongbae Heo, Michael R. Olson, Andrew P. Rutter, James J. Schauer ..... 818
- Characterization of particle size distribution of mainstream cigarette smoke generated by smoking machine with an electrical low pressure impactor  
Xiang Li, Haohui Kong, Xinying Zhang, Bin Peng, Cong Nie, Guanglin Shen, Huimin Liu ..... 827

### Terrestrial environment

- Differential responses of short-term soil respiration dynamics to the experimental addition of nitrogen and water in the temperate semi-arid steppe of Inner Mongolia, China  
Yuchun Qi, Xinchao Liu, Yunshe Dong, Qin Peng, Yating He, Liangjie Sun, Junqiang Jia, Congcong Cao ..... 834
- Effects of bile salts and divalent cations on the adsorption of norfloxacin by agricultural soils  
Xuesong Kong, Shixiang Feng, Xu Zhang, Yan Li ..... 846
- Tannic acid and saponin for removing arsenic from brownfield soils: Mobilization, distribution and speciation  
Zygmunt Mariusz Gusiatin ..... 855

### Environmental biology

- Molecular analysis of long-term biofilm formation on PVC and cast iron surfaces in drinking water distribution system  
Ruyin Liu, Junge Zhu, Zhisheng Yu, DevRaj Joshi, Hongxun Zhang, Wenfang Lin, Min Yang ..... 865
- Effect of a high strength chemical industry wastewater on microbial community dynamics and mesophilic methane generation  
Harish Venkatakishnan, Youming Tan, Maszenan bin Abdul Majid, Santosh Pathak, Antonius Yudi Sendjaja, Dongzhe Li, Jerry Jian Lin Liu, Yan Zhou, Wun Jern Ng ..... 875
- Effects of cathode potentials and nitrate concentrations on dissimilatory nitrate reductions by *Pseudomonas alcaliphila* in bioelectrochemical systems  
Wenjie Zhang, Yao Zhang, Wentao Su, Yong Jiang, Min Su, Ping Gao, Daping Li ..... 885
- Arsenic dynamics in the rhizosphere and its sequestration on rice roots as affected by root oxidation  
Weisong Pan, Chuan Wu, Shengguo Xue, William Hartley ..... 892

---

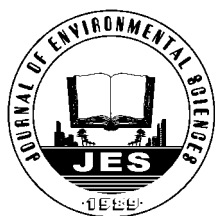
## Environmental health and toxicology

Alterations of endogenous metabolites in urine of rats exposed to decabromodiphenyl ether using metabonomic approaches Weijin Yang, Jianjie Fu, Thanh Wang, Hanxia Liu, Yawei Wang, Qunfang Zhou, Guibin Jiang .....	900
Integrated biomarkers in wild crucian carp for early warning of water quality in Hun River, North China Binghui Zheng, Kun Lei, Ruizhi Liu, Shuangshuang Song, Lihui An .....	909
T-2 toxin induces developmental toxicity and apoptosis in zebrafish embryos Guogang Yuan, Yimei Wang, Xiaoyan Yuan, Tingfen Zhang, Jun Zhao, Liuyu Huang, Shuangqing Peng .....	917

## Environmental analytical methods

Determining short chain fatty acids in sewage sludge hydrolysate: A comparison of three analytical methods and investigation of sample storage effects Victor Ibrahim, Tobias Hey, Karin Jönsson .....	926
--	-----

Serial parameter: CN 11-2629/X\*1989\*m\*217\*en\*P\*24\*2014-4

Available online at [www.sciencedirect.com](http://www.sciencedirect.com)

Journal of Environmental Sciences

[www.jesc.ac.cn](http://www.jesc.ac.cn)

## Effects of bile salts and divalent cations on the adsorption of norfloxacin by agricultural soils

Xuesong Kong, Shixiang Feng, Xu Zhang\*, Yan Li

School of Resources and Environmental Science, Wuhan University, Wuhan 430079, China. E-mail: [xuesongk@gmail.com](mailto:xuesongk@gmail.com)

### ARTICLE INFO

#### Article history:

Received 16 May 2013

revised 29 July 2013

accepted 03 September 2013

#### Keywords:

adsorption

agricultural soil

bile salts

divalent cation

norfloxacin

DOI: 10.1016/S1001-0742(13)60480-5

### ABSTRACT

The effects of bile salts (sodium cholate and sodium deoxycholate, 0–20 mmol/L), divalent cations ( $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ , 0–20 mmol/L) or pH (3.0–10.0) on the adsorption of norfloxacin by three selected soils (Paddy\_H, Paddy\_G and Red\_J) were systematically studied. Soil adsorption of norfloxacin follows a pseudo second-order kinetics model, and the maximum adsorption capacity has been determined from the nonlinear fit of the Langmuir isotherm model to be 88.8, 88.1 and 63.0  $\mu\text{mol/g}$  for the adsorption onto Paddy\_H, Paddy\_G and Red\_J, respectively. The results indicate that norfloxacin has a high adsorption affinity for the agricultural soils tested and that the organic content of these soils have at least a slight influence on this adsorption. The adsorption of norfloxacin to soils was strongly dependent on pH and exhibited a maximum at approximately pH 6. The presence of divalent cations prominently suppressed the adsorption of norfloxacin by paddy soils, which followed an order of  $\text{Cu}^{2+} > \text{Mg}^{2+} > \text{Ca}^{2+} > \text{Zn}^{2+}$ , and by red soil, which followed an order of  $\text{Cu}^{2+} > \text{Zn}^{2+} > \text{Ca}^{2+} > \text{Mg}^{2+}$ . The adsorption of norfloxacin (by the soils studied) sharply decreased as the amount of bile salts was increased. For uncharged norfloxacin at environmentally relevant pH values, such factors as soil type, exogenous divalent cations and macromolecules significantly altered the environmental fate and transport of norfloxacin between aquatic and soil interfaces.

## Introduction

Soil residues of pharmaceuticals and personal care products (PPCPs), including antibiotics and steroidal hormones, have originated from the irrigation and fertilization of agricultural land using contaminated excrement (Tolls, 2001; Thiele-Bruhn, 2003). Once released into the environment, in addition to biodegradation and chemical reactions, sorption of these residues onto various solids is an important transport process, particularly for PPCPs (Khetan and Collins, 2007). The adsorbents' properties, such as hydrophobicity/hydrophilicity and pH values, have been shown to affect the sorption process significantly.

Fluoroquinolone antibiotics are widely used as chemotherapeutic antibacterial agents and are widely

found in the environment (Sukul and Spiteller, 2007). Norfloxacin shows a high adsorption affinity for clay minerals, natural aquifer material and soils (Hari et al., 2005; Lorphensri et al., 2006; Liu et al., 2008; Zhang et al., 2009; Conkle et al., 2010; Pei et al., 2011). The factors influencing norfloxacin's adsorption have been systematically studied. The effect of aqueous pH on the adsorption of norfloxacin appears to depend on the particular adsorbent. Acidic conditions ( $\text{pH} < 5.0$ ) favor the adsorption of norfloxacin by Canadian River alluvium and montmorillonite (Hari et al., 2005; Pei et al., 2011), while maximum adsorption by alumina and silica occurs at neutral pH and around pH 8.5 (Lorphensri et al., 2006). Low-molecular-weight (LMW) organic acids have been shown to retard the adsorption of norfloxacin by Chinese Ferrallic soils (Zhang and Dong, 2008). Norfloxacin can form complexes with various types of metal ions (Khallo and Al-Assaf, 2011), and metal ions are usually

\* Corresponding authors. E-mail: [xuzhangwu@gmail.com](mailto:xuzhangwu@gmail.com)



present in wastewaters and enter the environment with PPCPs through wastewater discharge. Therefore, it is worthwhile to study the effects of coexisting metal ions on the adsorption of norfloxacin by soils. The influence of  $\text{Cu}^{2+}$  on the adsorption of norfloxacin depends on soil type and  $\text{Cu}^{2+}$  concentration (Zhang et al., 2009). However, the influence of other common aquatic divalent cations, including  $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$  and  $\text{Zn}^{2+}$ , has not been investigated and compared.

Host molecules, such as cyclodextrins, surfactant micelles, bile salts (sodium cholate, sodium deoxycholate, etc.) and macromolecules, are widely used in pharmaceutical products to improve the delivery and the effectiveness of drugs (Szejtli, 2004; Uekama et al., 2006; Zhou et al., 2010; Jazkewitsch et al., 2011; Yhaya et al., 2011). The surfactants cetylpyridinium chloride (CPC) and Tergitol NP9 have been shown to have no influence on the adsorption of norfloxacin by aquifer materials over a wide pH region (5.5–12). On the other hand, CPC can cause a considerable increase in the adsorption of nalidixic acid at higher pH values, increasing the adsorption coefficient 4-fold, from approximately 0.5 to 2 mL/g (Hari et al., 2005). The co-effect of humic acid and methyl- $\beta$ -cyclodextrin is similar to the solo effect of humic acid on the adsorption of norfloxacin by Hombikat UV-100, while the co-effect shows a greater retardation effect (compared to the solo effect of humic acid) on the adsorption of norfloxacin by anatase  $\text{TiO}_2$ . Methyl- $\beta$ -cyclodextrin and humic acid have exhibited a synergetic depressed effect on the adsorption of norfloxacin by both types of  $\text{TiO}_2$  nanoparticles (Peng et al., 2012). These findings indicate that host molecules with or without natural aquatic compositions (like humic acid) can alter the adsorption of pharmaceuticals.

The objective of the present work was to determine how norfloxacin adsorption is affected by the properties of soils from different provinces in China, as well as the presence of bio-macromolecules and exogenous divalent cations. To this end, the influences of two bile salts (sodium cholate (NaC) and sodium deoxycholate (NaDC)) at varying concentrations (0 to 50 mmol/L), the pH (3.0 to 10.0), the presence of exogenous divalent cations ( $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{Cu}^{2+}$  or  $\text{Zn}^{2+}$  from 0 to 10 mmol/L) and the soil type (paddy soil and red soil) on the adsorption of norfloxacin to Chinese agriculture soils were systemically investigated.

## 1 Materials and methods

### 1.1 Chemicals

Norfloxacin (98%) was purchased from Tokyo Chemical Industry and used as received. NaC (98%) and NaDC (98%) were commercial products from Aladdin-reagent Corporation (Shanghai, China). Paddy soils from Hubei Province (GBW07415a (ASA-4a)) and Guangdong Province (GBW07417a (ASA-6a)) and Red soils from Jiangxi Province (GBW07416a (ASA-5a)) were purchased from National Standard Reference Materials Center (Beijing, China) and are referred to as Paddy\_H, Paddy\_G and Red\_J throughout this study. The mechanical compositions and basic chemical properties of soils are listed in **Table 1**. To study the effect of soil organic matter on the adsorption of norfloxacin, organic C-free soils were obtained by treatment with  $\text{H}_2\text{O}_2$  as described in previous work (Zhang et al., 2009). Analytical-grade calcium chloride dihydrate ( $\text{CaCl}_2 \cdot 2\text{H}_2\text{O}$ ), magnesium chloride ( $\text{MgCl}_2$ ), copper chloride dihydrate ( $\text{CuCl}_2 \cdot 2\text{H}_2\text{O}$ ) and zinc chloride ( $\text{ZnCl}_2$ ), commercial products of Sinopharm Chemical Reagent Co. Ltd. (Shanghai, China), were used as precursors to the corresponding divalent cations. Deionized water (resistivity > 18.0  $\text{M}\Omega \cdot \text{cm}$ ) was used for sample preparation.

Norfloxacin (100  $\mu\text{mol/L}$ ), NaC (100  $\mu\text{mol/L}$ ) and NaDC (100  $\mu\text{mol/L}$ ) stock solutions were prepared by dissolving solid compounds in water. All low-concentration solutions were prepared by dilution. Soil dispersions (0.5 g/L) were prepared by directly adding 0.05 g of the corresponding soil into 100-mL solutions containing the analyte of interest.

### 1.2 Equipment and methods

Batch adsorption studies were performed in the dark using aqueous suspensions prepared in 150-mL glass vessels containing norfloxacin at different initial concentrations ( $c_0$  from 20 to 80  $\mu\text{mol/L}$ ) and 0.5 g/L of the soil under investigation. Each 100-mL suspension was continuously stirred for 12 hr, using a constant-temperature magnetic stirrer. After reaching equilibrium, 5-mL aliquots of these suspensions were withdrawn to determine the equilibrium concentration  $c_t$ . The effects of NaC, NaDC and metal ions on norfloxacin adsorption were investigated in the

**Table 1** Physicochemical properties of the three test soils

Soil	OM (%)	$f_{\text{OC}}$ (%)	CEC (cmol/kg)	WSS (%)	pH*	Main clay minerals
Paddy_H	3.3 $\pm$ 0.1	2.2 $\pm$ 0.3	19.0 $\pm$ 1.0	0.11	6.1	Hydromica, smectite, vermiculite, kaolinite
Paddy_G	3.8 $\pm$ 0.1	2.4 $\pm$ 0.2	19.7 $\pm$ 1.1	0.14	6.8	Kaolinite, hydromica, vermiculite, smectite
Red_J	0.73 $\pm$ 0.05	0.41 $\pm$ 0.3	10.0 $\pm$ 0.6	0.034	4.7	Kaolinite, hydromica, vermiculite

CEC: cation exchange capacity; WSS: water-soluble salt.

\*pH of the extraction solution of the soil dispersion (water ( $\text{CO}_2$  free):soil = 1:2.5, V/m).

same manner. The pH of the suspended solutions was adjusted by addition of dilute HClO<sub>4</sub> or NaOH solutions (0.1 mol/L).

Norfloxacin concentrations were determined using reversed-phase HPLC (Peng et al., 2012), and experiments were performed using a Waters 484 HPLC (Waters, America) with an Agilent Zorbax SB-C18 column (5 μm, 4.6 × 250 mm). The mobile phase consisted of a mixed solution of 0.025 mol/L phosphoric acid-acetonitrile (80/20, V/V), for which the pH was adjusted with triethylamine (to pH 3.0), and the flow rate was 1.0 mL/min. The UV-detection wavelength was set at 278 nm. The retention time for norfloxacin was 3.8 min, and its concentration was determined via the working curve method (from 0.1 to 100 μmol/L).

### 1.3 FT-IR study of norfloxacin adsorption by soils

Norfloxacin adsorption (by the soils of interest) was characterized using FT-IR spectroscopy. Norfloxacin (50 μmol/L) was added to each aqueous soil dispersion (1 g/L), and the resulting solutions were stirred for 12 hr. Norfloxacin-adsorbed soils were recovered by centrifuging each suspension (10,000 r/min for 30 min). The resulting solids were rinsed with deionized water until no norfloxacin was detected in the supernatant and were then dried in a vacuum oven at 60°C overnight. Fourier-transform infrared spectroscopy (FT-IR) studies of KBr pellets containing the samples were performed using a Nicolet 5700 infrared spectrometer (Thermo Electron Corporation, American). All spectra were scanned within the range 400–4000 cm<sup>-1</sup> with a resolution of 4 cm<sup>-1</sup>. Each sample was scanned 64 times.

### 1.4 Data analysis

Adsorption ratios (*R*, %) for norfloxacin adsorption by the soils studied were calculated using Eq. (1):

$$R = \frac{c_t}{c_0} \times 100\% \quad (1)$$

where, *c*<sub>0</sub> and *c*<sub>*t*</sub> represent the initial concentration and the concentration at different time intervals, respectively.

The reversible adsorption of norfloxacin between the bulk phase and soil at fixed initial concentration is represented by the adsorption coefficient (*K*<sub>d</sub>), which is defined as the ratio of the concentration of norfloxacin in the solid phase (*c*<sub>s</sub>) to that in the water (*c*<sub>eq</sub>) at equilibrium (Eq. (2)):

$$K_d = \frac{c_s}{c_{eq}} \quad (2)$$

*K*<sub>oc</sub> is the adsorption coefficient normalized to the percentage of organic carbon (*f*<sub>oc</sub>) present in the soil, which was calculated according to Eq. (3):

$$K_{oc} = \frac{K_d}{f_{oc}} \times 100\% \quad (3)$$

Both Langmuir and Freundlich isotherm models were used to quantify the adsorption of norfloxacin by the selected soils. Adsorption parameters were determined from nonlinear regression fits of the adsorption equations (Eqs. (4) and (5)) below.

$$q_e = \frac{q_{max} K_L c_{eq}}{1 + K_L c_{eq}} \quad (4)$$

where, *q*<sub>e</sub> (μmol/g) is the amount of solute adsorbed per gram of soil, *K*<sub>L</sub> (L/mol) is the Langmuir equilibrium constant, *q*<sub>max</sub> is the maximum adsorption capacity of the solute/soils and *c*<sub>eq</sub> (μmol/L) is the solute's equilibrium concentration.

$$q_e = K_F c_{eq}^{1/n} \quad (5)$$

where, *K*<sub>F</sub> and 1/*n* represent the Freundlich adsorption constant and the unitless linearity parameter, respectively.

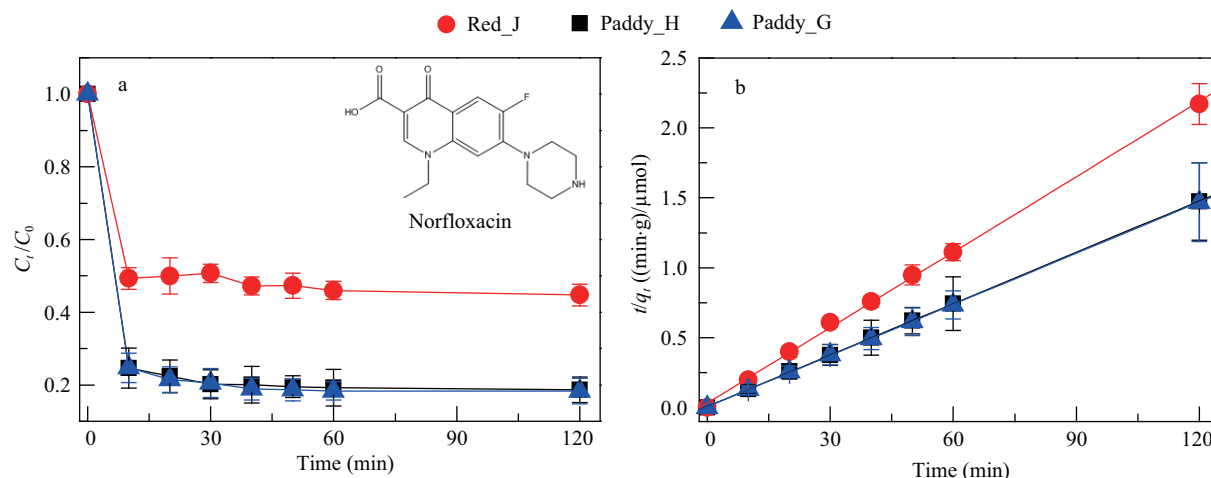
All experiments were performed in triplicate, and data are represented as means ± standard deviations. Regression analyses of the experimental data were performed using Origin 7.5 software.

## 2 Results and discussion

### 2.1 Norfloxacin adsorption equilibrium for the soils studied

As summarized in **Table 1**, each paddy soil studied had much higher organic content than the red soil studied. On the other hand, the cation-exchange capacity (CEC) and water-soluble salt (WSS) contents were determined to be approximately 1.9 and 3.2 times higher for the paddy soils than for the red soil. **Figure 1a** illustrates the effects of contact time on norfloxacin adsorption by soils. For all the test soils, the bulk concentration of norfloxacin decreased quickly with time up to 30 min and reached equilibrium in 60 min. After achieving adsorption equilibrium, the adsorption rate of norfloxacin onto both paddy soils was approximately 85%, much higher than that obtained for the red soil (55%).

To investigate the effects of soil organic matter on the adsorption of norfloxacin, adsorption experiments were also conducted using the corresponding organic C-free soils as adsorbents. After H<sub>2</sub>O<sub>2</sub> treatment, the organic C contents of Paddy\_H, Paddy\_G and Red\_J were found to be 0.2%, 0.3% and 0.1%, respectively. The corresponding adsorption rates were 70%, 70% and 45%. The results indicate that organic C plays a minor role in norfloxacin adsorption by the three soils studied. Similar results have been found regarding norfloxacin adsorption by other agricultural soils (Zhang et al., 2009).



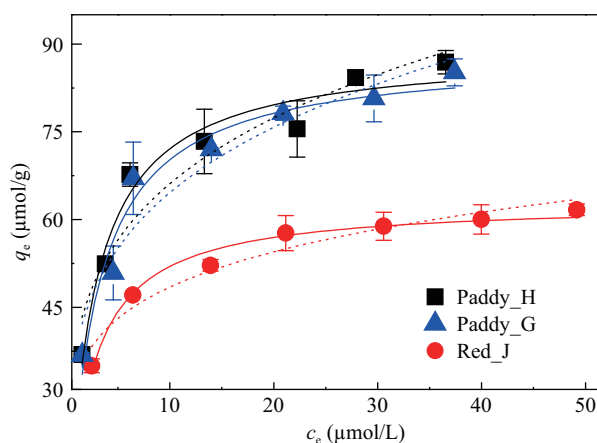
**Fig. 1** (a) Dependence of norfloxacin bulk concentration ( $C_t$ ) on contact time (in 0.5 g/L soil suspensions); inset shows the structural formula of norfloxacin, (b) adsorption kinetics of norfloxacin onto soils; the lines correspond to data fit to the pseudo second-order kinetic model. Norfloxacin conc.: 50 mol/L.

The pseudo second-order kinetics model has been widely applied to the sorption of antibiotics from aqueous solution onto a solid phase (Ho and McKay, 2003; Basha et al., 2011; Wang et al., 2011; Gao et al., 2012; Peng et al., 2012). The linear form of the pseudo second-order equation is represented by Eq. (6):

$$\frac{t}{q_t} = \frac{1}{kq_e^2} + \frac{t}{q_e} \quad (6)$$

where,  $k$  (g/(μmol·min)) is the rate constant of adsorption,  $q_e$  (μmol/g) is the amount of norfloxacin adsorbed at equilibrium,  $q_t$  (μmol/g) is the amount of norfloxacin adsorbed onto soils at reaction time  $t$ , and  $kq_e^2$  (μmol/(g·min)) is the initial rate. As shown in **Fig. 1b**, the correlation coefficient ( $r^2$ ) for the pseudo second-order kinetic model was higher than 0.999. The calculated adsorption-capacity ( $q_e$ ) values recovered from the pseudo second-order model were 81.7, 82.1, and 55.6 μmol/g for the adsorption of norfloxacin (at an initial concentration of 50 mol/L) onto Paddy\_H, Paddy\_G and Red\_J, respectively, and were consistent with the experimental results (81.4, 81.7, and 55.2 μmol/g, respectively). These results demonstrate that the pseudo second-order model is a good approach for describing the process of norfloxacin adsorption by different soils.

The  $K_d$  values reported for other fluoroquinolone antibiotics, such as ciprofloxacin, enrofloxacin and ofloxacin, varied from hundreds to thousands (Thiele-Bruhn, 2003). In the present work, the calculated  $K_d$  for Paddy\_H, Paddy\_G and Red\_J was found to be  $8.7 \times 10^3$ ,  $8.9 \times 10^3$  and  $2.5 \times 10^3$  L/kg, respectively, which is within the range of values ( $10^2$ – $10^4$  L/kg) reported in previous work (Zhang and Dong, 2007; Zhang et al., 2009). The corresponding  $K_{oc}$  for Paddy\_H, Paddy\_G and Red\_J is  $4.0 \times 10^5$ ,  $3.7 \times 10^5$  and  $6.1 \times 10^5$  L/kg, respectively.



**Fig. 2** Dependence of the adsorbed norfloxacin per gram of soil ( $q_e$ ) on the equilibrium concentration ( $c_e$ ). The solid lines and dotted lines correspond to fits of the data to Langmuir and Freundlich adsorption isotherm models, respectively. The regression coefficients were larger than 0.98.

## 2.2 Isotherms for norfloxacin adsorption by soils

The dependence of the norfloxacin equilibrium concentration on adsorption amount is shown in **Fig. 2**. When the same initial concentration was used, the norfloxacin adsorbed by paddy soils increased more sharply than that adsorbed by red soil. **Figure 2** shows plots of experimental adsorption data fitted with the corresponding Langmuir adsorption isotherm and Freundlich adsorption isotherm. The fitted parameters for the adsorption equations are listed in **Table 2**.

The  $K_L$  (L/mol) values obtained for norfloxacin adsorption by Paddy\_H, Paddy\_G and Red\_J are  $(4.3 \pm 0.6) \times 10^5$ ,  $(3.9 \pm 0.6) \times 10^5$  and  $(3.4 \pm 0.4) \times 10^5$ , respectively. These  $K_L$  values are within the range of values ( $10^4$ – $10^5$  L/mol) reported for norfloxacin adsorption to other adsorbents (Hari et al., 2005; Liu et al., 2008; Zhang et al., 2009; Yang

**Table 2** Calculated Langmuir and Freundlich isotherm parameters, and the correlation coefficients ( $R^2$ ) for the adsorption of norfloxacin onto the three soils

Soil	Freundlich isotherm			Langmuir isotherm		
	$K_F$ ( $\mu\text{mol}^{(1-1/n)} \times \text{L}^{1/n}/\text{g}$ )	$1/n$	$R^2$	$q_{\text{max}}$ ( $\mu\text{mol}/\text{g}$ )	$K_L$ ( $\times 10^5$ L/mol)	$R^2$
Paddy_H	$39.3 \pm 3.6$	$0.23 \pm 0.03$	0.928	$88.8 \pm 2.7$	$4.3 \pm 0.6$	0.965
Paddy_G	$38.1 \pm 3.5$	$0.23 \pm 0.03$	0.929	$88.1 \pm 2.9$	$3.9 \pm 0.6$	0.961
Red_J	$32.8 \pm 2.0$	$0.17 \pm 0.02$	0.948	$52.5 \pm 1.0$	$3.4 \pm 0.4$	0.967

et al., 2012). On the other hand, the values obtained in this work are much smaller than the values for norfloxacin adsorption onto a  $\text{TiO}_2$  surface ( $10^6$  L/mol) (Peng et al., 2012).

### 2.3 Effect of pH on norfloxacin adsorption

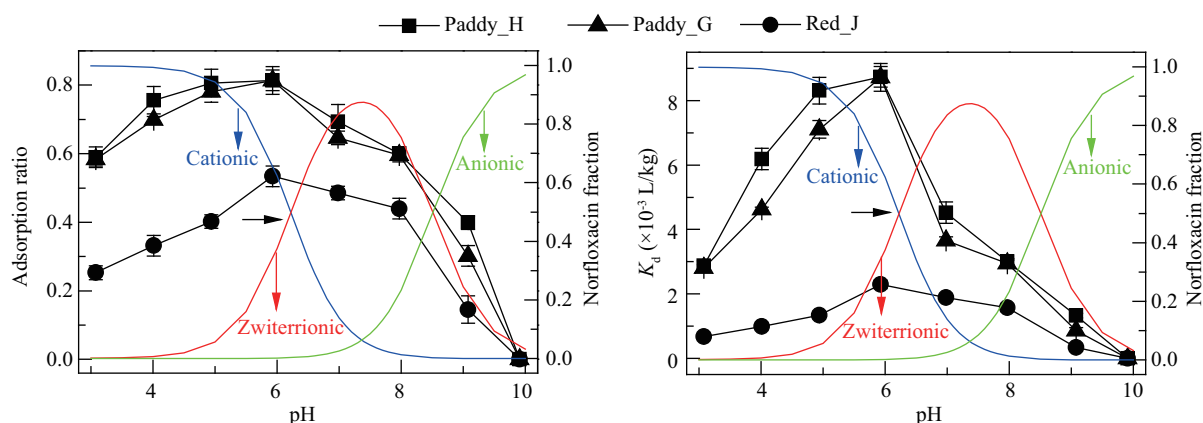
The pH of the aquatic environment can drastically change both the existing form of norfloxacin and the surface charge of the soil and can, therefore, influence norfloxacin adsorption. As shown in Fig. 3, the norfloxacin cation is the dominant species at pH values below 6.2, the zwitterionic form is dominant in the pH range of 6.3 ( $\text{p}K_{\text{a1}}$ ) to 8.4 ( $\text{p}K_{\text{a2}}$ ) (Hari et al., 2005), and the anionic form only exists when the pH is higher than 6.5 (according to its acid-base equilibrium constant). To evaluate the effect of pH on norfloxacin adsorption, adsorption experiments were performed with an initial concentration of  $50 \mu\text{mol}/\text{L}$  to determine adsorption ratios and adsorption coefficients. Norfloxacin adsorption studies of the selected soils were conducted for pH values ranging from 3.0 to 10.0, and the results are shown in Fig. 3.

Norfloxacin adsorption by paddy soils was higher than that by red soils at all the pH values studied. The adsorption rates and adsorption coefficients (for norfloxacin onto the three soils studied) were relatively smaller in acidic solution and increased quickly as the pH increased. Adsorption onto all three soil types studied reached a maximum at pH 6.0, which is within norfloxacin's cationic region. The adsorption of norfloxacin onto all soils decreased sharply

when the pH ranged from 6.0 to 10.0, where the anionic form of norfloxacin becomes the dominant species. The effect of pH on norfloxacin adsorption by the soils studied herein is consistent with the effect of pH on norfloxacin adsorption onto other types of soils and natural aquifer materials (Hari et al., 2005; Zhang et al., 2009). It has been suggested that the bulk of norfloxacin adsorption onto natural aquifer materials likely results from electrostatic attraction between the cationic and zwitterionic species of norfloxacin and the negatively charged quartz sand surface (Hari et al., 2005). Therefore, the electrostatic repulsion between anionic norfloxacin and a negatively charged soil surface may increase with an increase in pH, causing a corresponding decrease in adsorption.

### 2.4 Effects of NaC and NaDC on norfloxacin adsorption by the soils studied

Drug carriers, including liposomes, cyclodextrins, bile salts, surfactants and synthetic polymers, are used in sundry drug delivery systems to prolong *in vivo* drug actions, decrease drug metabolism, and reduce drug toxicity. Although these compounds are likely to be present at low concentrations in the environment and are usually considered to be non-toxic, their effects on the fate of PPCPs and other pollutants should also be taken in account when considering the design and synthesis of a new drug carrier because many studies have found that the fate and transport of the host-guest supramolecule is different from that of free guests, (Sortino et al., 2001; Hapiot et al., 2002;



**Fig. 3** Dependence of the norfloxacin adsorption ratio ( $R$ ) and adsorption coefficient ( $K_d$ ) on the soil, and norfloxacin's predominant form as a function of solution pH.



Yamada et al., 2002; Cosa and Scaiano, 2004) and that the adsorption/desorption process of pollutants is also changed in the presence of drug carriers as well as dissolved organic matter (Guo et al., 2010; Peng et al., 2012).

Sodium cholate (NaC) and sodium deoxycholate (NaDC) are the most common bile salts and, excepting that NaDC lacks a 7-hydroxyl group, have the same steroidal skeleton (Ćirin et al., 2012). NaC, NaDC and other bile salts are used in pharmaceutical formulations, where they solubilize poorly soluble molecules and can improve the permeation of various drugs across biological membranes (Martin et al., 2005; Guo et al., 2010; Mahajan and Mahajan, 2012). Therefore, it is worthwhile to study the effect of bile salts on the adsorption of drugs.

The effect of NaC and NaDC on norfloxacin adsorption by soils is shown in Fig. 4. Unlike the lack of apparent effects by other host molecules, including cyclodextrins, Tergitol NP9 and ethoxylated nonylphenol surfactants, on norfloxacin adsorption to a solid phase (Hari et al., 2005; Peng et al., 2012), both NaC and NaDC demonstrated a significant retardation effect on norfloxacin adsorption by all the soils tested in the present study. In the presence of 20 mmol/L NaC, the  $K_d$  values for norfloxacin adsorption onto Paddy\_H, Paddy\_G and Red\_J were  $(8.1 \pm 0.09) \times 10^2$ ,  $(7.1 \pm 0.06) \times 10^2$  and  $(6.4 \pm 0.08) \times 10^2$  L/kg, respectively. These values are approximately 10.7, 12.5 and 3.9 times smaller than their corresponding values obtained in the absence of NaC. A similar experimental phenomenon occurred for the adsorption of norfloxacin onto soils in the presence of 20 mmol/L NaDC.

Various guests, with different structures and properties, can form host-guest complexes with NaC and NaDC aggregates. Thus, it has been suggested that hydrophobic guests can bind to primary or primary/secondary aggregates, while hydrophilic guests are only incorporated into secondary aggregates (Amundson et al., 2008). Norfloxacin's point of maximum hydrophobicity occurs around neutral pH (Takács-Novák et al., 1992). Therefore,

it is reasonable to propose that, at neutral pH, norfloxacin could incorporate with both the primary and secondary aggregates of bile salts. The bile-salt concentration in the bulk phase should continuously increase with the as the bile-salt dosage is increased, so that norfloxacin binding to aggregates in the bulk phase could decrease norfloxacin adsorption onto the soil surface.

However, it is noteworthy that the concentrations of norfloxacin and bile salts used in this work are excessively higher than their concentrations in natural aquatic environments. Batch adsorption studies performed at lower concentrations, to reveal to what extent different adsorption properties are observed, are ongoing.

## 2.5 Effect of divalent cations on norfloxacin adsorption by the soils studied

Additional experiments investigated the effects of divalent cations ( $\text{Ca}^{2+}$ ,  $\text{Mg}^{2+}$ ,  $\text{Cu}^{2+}$  and  $\text{Zn}^{2+}$ ) on norfloxacin adsorption by the three test soils. The results, which are depicted in Fig. 5, clearly show that norfloxacin adsorption decreases as the concentration of divalent cations increases and that  $\text{Cu}^{2+}$  cause the greatest inhibition of norfloxacin adsorption. Close inspection of the experimental data reveals different effects of metal ions on norfloxacin adsorption by paddy soils and red soil. The effect of metal ions on norfloxacin adsorption by the two paddy soils follows the order of  $\text{Cu}^{2+} > \text{Mg}^{2+} > \text{Ca}^{2+} > \text{Zn}^{2+}$ , while the effect of metal ions on norfloxacin adsorption onto red soil follows the order of  $\text{Cu}^{2+} > \text{Zn}^{2+} > \text{Ca}^{2+} > \text{Mg}^{2+}$ . These results indicate that, neglecting differences in the ability of norfloxacin to complex with different metal ions, the different effects of the four divalent cations on norfloxacin adsorption should be significantly affected by differences of metal-ion adsorption by soils as well as by the metal ions that enter solution from the soils. Unfortunately, at the present stage, we cannot conclude which is the dominant factor.

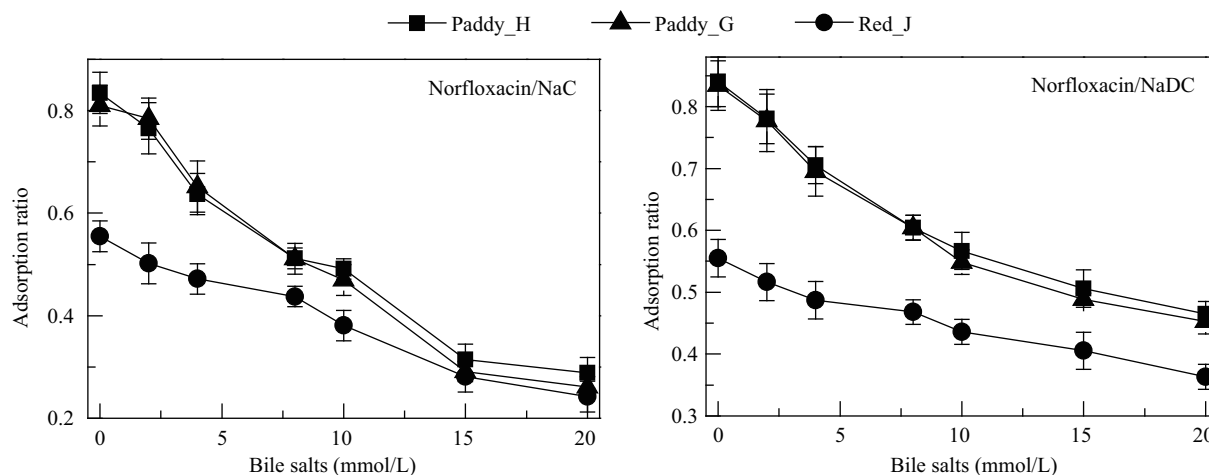


Fig. 4 Effect of NaC and NaDC on the adsorption ratio (calculated using Eq. (1)) for norfloxacin (50  $\mu\text{mol/L}$ ) adsorption onto soils (0.5 g/L).

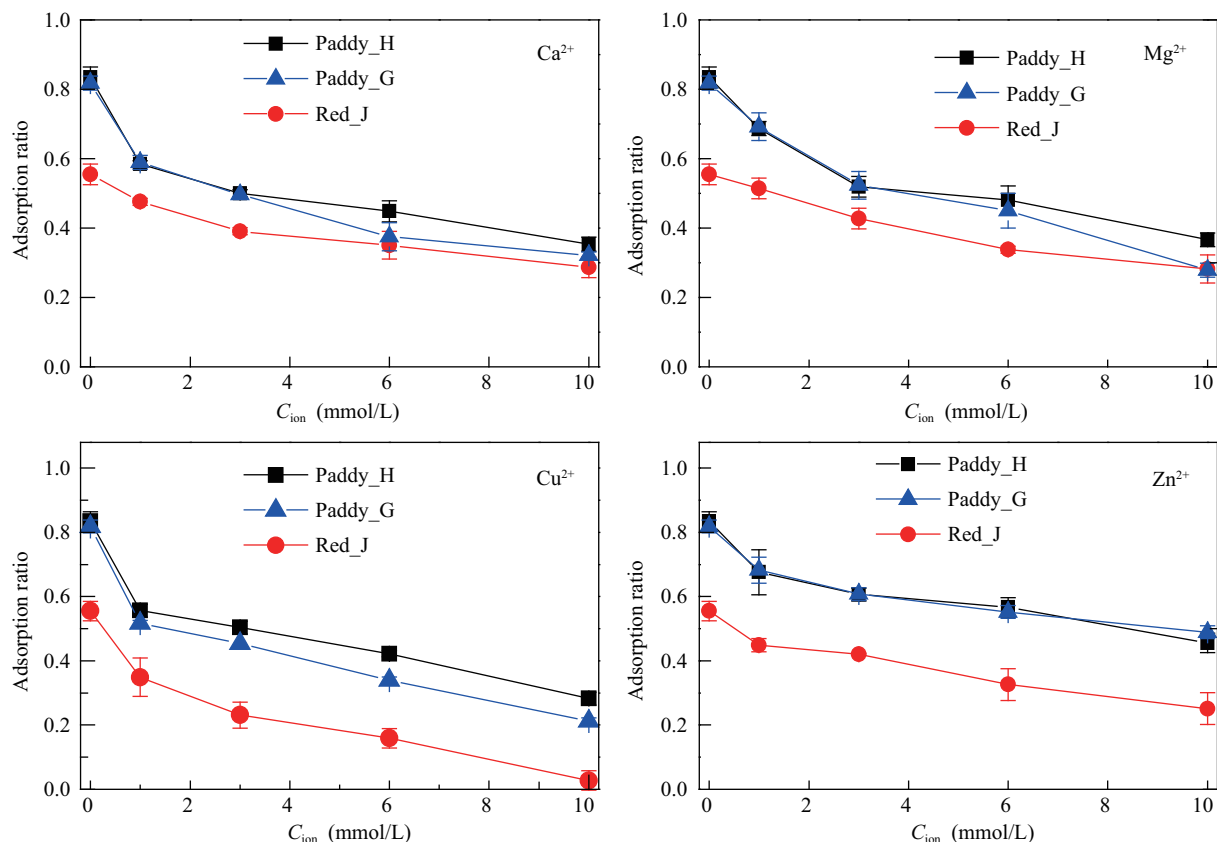


Fig. 5 Effect of divalent cations on the adsorption ratio (calculated using Eq. (1)) for norfloxacin (50  $\mu\text{mol/L}$ ) onto soils (0.5 g/L).

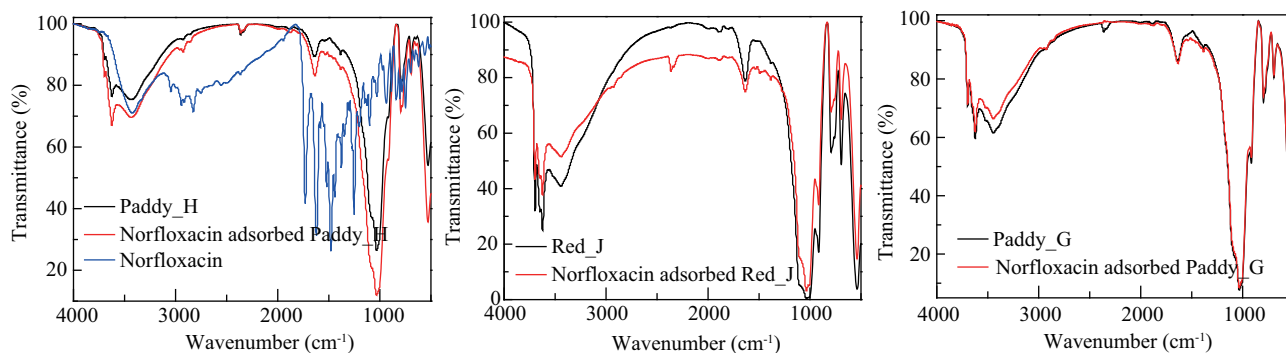


Fig. 6 FT-IR spectra of free norfloxacin, soils and norfloxacin-adsorbed soils.

## 2.6 FT-IR study of norfloxacin adsorption by soils

FT-IR analysis was performed to study the preliminary adsorption mode of norfloxacin onto soils, and the results are shown in Fig. 6. FT-IR spectra of norfloxacin and metal-norfloxacin complexes have been well assigned (Al-Mustafa, 2002; Sadeek, 2005). In the spectra of norfloxacin-adsorbed soils presented herein, it is reasonable to assign the new peaks at  $1492\text{ cm}^{-1}$  and  $1456\text{ cm}^{-1}$  to the vibration of ligated norfloxacin. However, the  $\nu(\text{C}=\text{O})$  stretching vibration of the carboxylic group, observed at  $1730\text{ cm}^{-1}$  for bulk norfloxacin, was not observed for adsorbed norfloxacin, indicating that the car-

boxyl group is involved in its interaction with soils. These results are similar to those obtained for several norfloxacin-metal-ion complexes (Al-Mustafa, 2002; Sadeek, 2005). These FT-IR results indicate that soils adsorb norfloxacin primarily through binding with metal ions located on soil surfaces.

## 3 Conclusions

The results obtained in the present study indicate that sodium cholate and sodium deoxycholate, divalent cations

(Ca<sup>2+</sup>, Mg<sup>2+</sup>, Cu<sup>2+</sup> and Zn<sup>2+</sup>) and solution pH play important roles in norfloxacin adsorption onto soils with different physicochemical properties. The adsorption of norfloxacin onto soil surfaces is highly pH-dependent. Both bile salts greatly inhibit the adsorption of norfloxacin onto paddy soils and red soil. Soils adsorb norfloxacin primarily via metal ions on the soil surface, rather than via norfloxacin penetration of layers or interlayers of soil clay minerals. Thus, the present study implies that the release of biomolecular hosts into the environment could greatly affect the fate and transport of norfloxacin molecules that exhibit high adsorption affinity.

### Acknowledgments

This work was supported by the National Natural Science Foundation of China (No. 21207104), the Natural Science Foundation of Hubei Province (No. 2011CDB274), the Youth Chenguang Project of Science and Technology of Wuhan City (No. 2013070104010009), the Fundamental Research Funds for the Central Universities (No. 121095) and the Postdoctoral Science Foundation of China (No. 2012 M511675).

### REFERENCES

- Al-Mustafa, J., 2002. Magnesium, calcium and barium perchlorate complexes of ciprofloxacin and norfloxacin. *Acta Chim. Slov.* 49(3), 457–466.
- Amundson, L.L., Li, R., Bohne, C., 2008. Effect of the guest size and shape on its binding dynamics with sodium cholate aggregates. *Langmuir* 24(16), 8491–8500.
- Basha, S., Barr, C., Keane, D., Nolan, K., Morrissey, A., Oelgemoller, M. et al., 2011. On the adsorption/photodegradation of amoxicillin in aqueous solutions by an integrated photocatalytic adsorbent (IPCA): experimental studies and kinetics analysis. *Photochem. Photobiol. Sci.* 10(6), 1014–1022.
- Ćirin, D.M., Poša, M.M., Krstonošić, V.S., 2012. Interactions between sodium cholate or sodium deoxycholate and nonionic surfactant (Tween 20 or Tween 60) in aqueous solution. *Indust. Eng. Chem. Res.* 51(9), 3670–3676.
- Conkle, J.L., Lattao, C., White, J.R., Cook, R.L., 2010. Competitive sorption and desorption behavior for three fluoroquinolone antibiotics in a wastewater treatment wetland soil. *Chemosphere* 80(11), 1353–1359.
- Cosa, G., Scaiano, J.C., 2004. Laser techniques in the study of drug photochemistry. *Photochem. Photobiol.* 80(2), 159–174.
- Gao, Y., Li, Y., Zhang, L., Huang, H., Hu, J., Shah, S.M., et al., 2012. Adsorption and removal of tetracycline antibiotics from aqueous solution by graphene oxide. *J. Coll. Inter. Sci.* 368(1), 540–546.
- Guo, H., Zhang, J., Liu, Z., Yang, S., Sun, C., 2010. Effect of Tween 80 and  $\beta$ -cyclodextrin on the distribution of herbicide mefenacet in soil-water system. *J. Hazard. Mater.* 177, 1039–1045.
- Hapiot, P., Lagrost, C., Aeiyaich, S., Jouini, M., Lacroix, J.C., 2002. Oxidative coupling of small oligothiophenes and oligopyrroles in water in the presence of cyclodextrin flash photolysis investigations. *J. Phys. Chem. B*, 106(14), 3622–3628.
- Hari, A.C., Paruchuri, R.A., Sabatini, D.A., Kibbey, T.C.G., 2005. Effects of pH and cationic and nonionic surfactants on the adsorption of pharmaceuticals to a natural aquifer material. *Environ. Sci. Technol.* 39(8), 2592–2598.
- Ho, Y.S., McKay, G., 2003. Sorption of dyes and copper ions onto biosorbents. *Process Biochem.* 38(7), 1047–1061.
- Jazkewitsch, O., Mondrzyk, A., Staffel, R., Ritter, H., 2011. Cyclodextrin-modified polyesters from lactones and from bacteria: an approach to new drug carrier systems. *Macromolecules* 44(6), 1365–1371.
- Khallow, K.I., Al-Assaf, A.Y.R., 2011. Synthesis and characterization of oxozirconium(IV), dioxomolybdenum(VI) and dioxotungsten(VI) with ciprofloxacin and Norfloxacin Complexes. *E-J. Chem.* 8(2), 576–580.
- Khetan, S.K., Collins, T.J., 2007. Human pharmaceuticals in the aquatic environment: a challenge to green chemistry. *Chem. Rev.* 107(6), 2319–2364.
- Lorphensri, O., Intravijit, J., Sabatini, D.A., Kibbey, T.C.G., Osathaphan, K., Saiwan, C., 2006. Sorption of acetaminophen, 17 $\alpha$ -ethynyl estradiol, nalidixic acid, and norfloxacin to silica, alumina, and a hydrophobic medium. *Water Res.* 40(7), 1481–1491.
- Mahajan, S., Mahajan, R.K., 2012. Interactions of phenothiazine drugs with bile salts: Micellization and binding studies. *J. Coll. Interf. Sci.* 387(1), 194–204.
- Martin, C., Thongborisute, J., Takeuchi, H., Yamamoto, H., Kawashima, Y., Alpar, H.O., 2005. Cholesterol-bile salt vesicles as potential delivery vehicles for drug and vaccine delivery. *Inter. J. Pharm.* 298(2), 339–343.
- Pei, Z.G., Shan, X.Q., Zhang, S.Z., Kong, J.J., Wen, B., Zhang, J. et al., 2011. Insight to ternary complexes of co-adsorption of norfloxacin and Cu(II) onto montmorillonite at different pH using EXAFS. *J. Hazard. Mater.* 186(1), 842–848.
- Peng, H., Feng, S.X., Zhang, X., Li, Y., Zhang, X.Y., 2012. Adsorption of norfloxacin onto titanium oxide: Effect of drug carrier and dissolved humic acid. *Sci. Total Environ.* 438, 66–71.
- Sadeek, S.A., 2005. Synthesis, thermogravimetric analysis, infrared, electronic and mass spectra of Mn(II), Co(II) and Fe(III) norfloxacin complexes. *J. Molec. Struct.* 753(1-3), 1–12.
- Sortino, S., Giuffrida, S., Fazio, S., Monti, S., 2001. Spectroscopic characterization and photochemical behavior of host-guest complexes between  $\beta$ -cyclodextrin and drugs containing a biphenyl-like chromophore. *New J. Chem.* 25(5), 707–713.
- Sukul, P., Spitteller, M., 2007. Fluoroquinolone antibiotics in the environment. In: *Reviews of Environmental Contamination and Toxicology*. 191. Springer, New York, pp. 131–162.
- Szejtli, J., 2004. Past, present and future of cyclodextrin research. *Pure Appl. Chem.* 76(10), 1825–1845.
- Takács-Novák, K., Józán, M., Hermecz, I., Szász, G., 1992. Lipophilicity of antibacterial fluoroquinolones. *Internat. J. Pharm.* 79(1-3), 89–96.
- Thiele-Bruhn, S., 2003. Pharmaceutical antibiotic compounds in soils – a review. *J. Plant Nutrit. Soil Sci.* 166(2), 145–167.
- Tolls, J., 2001. Sorption of Veterinary pharmaceuticals in soils: a review. *Environ. Sci. Technol.* 35(17), 3397–3406.
- Uekama, K., Fumitoshi, H., Arima, H., 2006. Pharmaceutical applications of cyclodextrins and their derivatives. In: Uekama, K., (Ed.). *Cyclodextrins and Their Complexes*. WILEY-VCH, Weinheim, pp. 381–422.

- Wang, C.J., Li, Z.H., Jiang, W.T., 2011. Adsorption of ciprofloxacin on 2:1 dioctahedral clay minerals. *Appl. Clay Sci.* 53(4), 723–728.
- Yamada, M., Kato, K., Nomizu, M., Ohkawa, K., Yamamoto, H., Nishi, N., 2002. UV-irradiated DNA matrixes selectively bind endocrine disruptors with a planar structure. *Environ. Sci. Technol.* 36(5), 949–954.
- Yang, W., Lu, Y., Zheng, F., Xue, X., Li, N., Liu, D., 2012. Adsorption behavior and mechanisms of norfloxacin onto porous resins and carbon nanotube. *Chem. Eng. J.* 179, 112–118.
- Yhaya, F., Lim, J., Kim, Y., Liang, M., Gregory, A.M., Stenzel, M.H., 2011. Development of micellar novel drug carrier utilizing temperature-sensitive block copolymers containing cyclodextrin moieties. *Macromolecules* 44(21), 8433–8445.
- Zhang, J.Q., Dong, Y.H., 2008. Effect of low-molecular-weight organic acids on the adsorption of norfloxacin in typical variable charge soils of China. *J. Hazard. Mater.* 151(2-3), 833–839.
- Zhang, J.Q., Dong, Y.H., 2007. Adsorption and desorption of norfloxacin on four typical soils in China. *Chin. J. Environ. Sci.* 28(9), 2134–2140.
- Zhang, J., Li, Z., Ge, G., Sun, W., Liang, Y., Wu, L., 2009. Impacts of soil organic matter, pH and exogenous copper on sorption behavior of norfloxacin in three soils. *J. Environ. Sci.* 21(5), 632–640.
- Zhou, H., Yu, W., Guo, X., Liu, X., Li, N., Zhang, Y. et al., 2010. Synthesis and characterization of amphiphilic glycidol-chitosan-deoxycholic acid nanoparticles as a drug carrier for doxorubicin. *Biomacromolecules* 11(12), 3480–3486.



## Editorial Board of Journal of Environmental Sciences

### Editor-in-Chief

**Hongxiao Tang** Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences, China

### Associate Editors-in-Chief

**Jiuhui Qu** Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences, China

**Shu Tao** Peking University, China

**Nigel Bell** Imperial College London, United Kingdom

**Po-Keung Wong** The Chinese University of Hong Kong, Hong Kong, China

### Editorial Board

#### Aquatic environment

**Baoyu Gao**

Shandong University, China

**Maohong Fan**

University of Wyoming, USA

**Chihpin Huang**

National Chiao Tung University

Taiwan, China

**Ng Wun Jern**

Nanyang Environment &  
Water Research Institute, Singapore

**Clark C. K. Liu**

University of Hawaii at Manoa, USA

**Hokyong Shon**

University of Technology, Sydney, Australia

**Zijian Wang**

Research Center for Eco-Environmental Sciences,  
Chinese Academy of Sciences, China

**Zhiwu Wang**

The Ohio State University, USA

**Yuxiang Wang**

Queen's University, Canada

**Min Yang**

Research Center for Eco-Environmental Sciences,  
Chinese Academy of Sciences, China

**Zhifeng Yang**

Beijing Normal University, China

**Han-Qing Yu**

University of Science & Technology of China

#### Terrestrial environment

**Christopher Anderson**

Massey University, New Zealand

**Zucong Cai**

Nanjing Normal University, China

**Xinbin Feng**

Institute of Geochemistry,  
Chinese Academy of Sciences, China

**Hongqing Hu**

Huazhong Agricultural University, China

**Kin-Che Lam**

The Chinese University of Hong Kong

Hong Kong, China

**Erwin Klumpp**

Research Centre Juelich, Agrosphere Institute  
Germany

**Peijun Li**

Institute of Applied Ecology,  
Chinese Academy of Sciences, China

**Michael Schloter**

German Research Center for Environmental Health  
Germany

**Xuejun Wang**

Peking University, China

**Lizhong Zhu**

Zhejiang University, China

#### Atmospheric environment

**Jianmin Chen**

Fudan University, China

**Abdelwahid Mellouki**

Centre National de la Recherche Scientifique  
France

**Yujing Mu**

Research Center for Eco-Environmental Sciences,  
Chinese Academy of Sciences, China

**Min Shao**

Peking University, China

**James Jay Schauer**

University of Wisconsin-Madison, USA

**Yuesi Wang**

Institute of Atmospheric Physics,  
Chinese Academy of Sciences, China

**Xin Yang**

University of Cambridge, UK

#### Environmental biology

**Yong Cai**

Florida International University, USA

**Henner Hollert**

RWTH Aachen University, Germany

**Jaе-Seong Lee**

Sungkyunkwan University, South Korea

**Christopher Rensing**

University of Copenhagen, Denmark

**Bojan Sedmak**

National Institute of Biology, Ljubljana

**Lirong Song**

Institute of Hydrobiology,  
the Chinese Academy of Sciences, China

**Chunxia Wang**

National Natural Science Foundation of China

**Gehong Wei**

Northwest A & F University, China

**Daqiang Yin**

Tongji University, China

**Zhongtang Yu**

The Ohio State University, USA

#### Environmental toxicology and health

**Jingwen Chen**

Dalian University of Technology, China

**Jiaying Hu**

Peking University, China

**Guibin Jiang**

Research Center for Eco-Environmental Sciences,  
Chinese Academy of Sciences, China

**Sijin Liu**

Research Center for Eco-Environmental Sciences,  
Chinese Academy of Sciences, China

**Tsuyoshi Nakanishi**

Gifu Pharmaceutical University, Japan

**Willie Peijnenburg**

University of Leiden, The Netherlands

**Bingsheng Zhou**

Institute of Hydrobiology,  
Chinese Academy of Sciences, China

#### Environmental catalysis and materials

**Hong He**

Research Center for Eco-Environmental Sciences,  
Chinese Academy of Sciences, China

**Junhua Li**

Tsinghua University, China

**Wenfeng Shangguan**

Shanghai Jiao Tong University, China

**Yasutake Teraoka**

Kyushu University, Japan

**Ralph T. Yang**

University of Michigan, USA

#### Environmental analysis and method

**Zongwei Cai**

Hong Kong Baptist University,  
Hong Kong, China

**Jiping Chen**

Dalian Institute of Chemical Physics,  
Chinese Academy of Sciences, China

**Minghui Zheng**

Research Center for Eco-Environmental Sciences,  
Chinese Academy of Sciences, China

#### Municipal solid waste and green chemistry

**Pinjing He**

Tongji University, China

#### Environmental ecology

**Rusong Wang**

Research Center for Eco-Environmental Sciences,  
Chinese Academy of Sciences, China

### Editorial office staff

**Managing editor** Qingcai Feng

**Editors** Zixuan Wang Suqin Liu Zhengang Mao

**English editor** Catherine Rice (USA)



# JOURNAL OF ENVIRONMENTAL SCIENCES

环境科学学报(英文版)  
(<http://www.jesc.ac.cn>)

## Aims and scope

*Journal of Environmental Sciences* is an international academic journal supervised by Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences. The journal publishes original, peer-reviewed innovative research and valuable findings in environmental sciences. The types of articles published are research article, critical review, rapid communications, and special issues.

The scope of the journal embraces the treatment processes for natural groundwater, municipal, agricultural and industrial water and wastewaters; physical and chemical methods for limitation of pollutants emission into the atmospheric environment; chemical and biological and phytoremediation of contaminated soil; fate and transport of pollutants in environments; toxicological effects of terrorist chemical release on the natural environment and human health; development of environmental catalysts and materials.

## For subscription to electronic edition

Elsevier is responsible for subscription of the journal. Please subscribe to the journal via <http://www.elsevier.com/locate/jes>.

## For subscription to print edition

China: Please contact the customer service, Science Press, 16 Donghuangchenggen North Street, Beijing 100717, China. Tel: +86-10-64017032; E-mail: [journal@mail.sciencep.com](mailto:journal@mail.sciencep.com), or the local post office throughout China (domestic postcode: 2-580).

Outside China: Please order the journal from the Elsevier Customer Service Department at the Regional Sales Office nearest you.

## Submission declaration

Submission of an article implies that the work described has not been published previously (except in the form of an abstract or as part of a published lecture or academic thesis), that it is not under consideration for publication elsewhere. The submission should be approved by all authors and tacitly or explicitly by the responsible authorities where the work was carried out. If the manuscript accepted, it will not be published elsewhere in the same form, in English or in any other language, including electronically without the written consent of the copyright-holder.

## Submission declaration

Submission of the work described has not been published previously (except in the form of an abstract or as part of a published lecture or academic thesis), that it is not under consideration for publication elsewhere. The publication should be approved by all authors and tacitly or explicitly by the responsible authorities where the work was carried out. If the manuscript accepted, it will not be published elsewhere in the same form, in English or in any other language, including electronically without the written consent of the copyright-holder.

## Editorial

Authors should submit manuscript online at <http://www.jesc.ac.cn>. In case of queries, please contact editorial office, Tel: +86-10-62920553, E-mail: [jesc@263.net](mailto:jesc@263.net), [jesc@rcees.ac.cn](mailto:jesc@rcees.ac.cn). Instruction to authors is available at <http://www.jesc.ac.cn>.

## Journal of Environmental Sciences (Established in 1989)

Vol. 26 No. 4 2014

<b>Supervised by</b>	Chinese Academy of Sciences	<b>Published by</b>	Science Press, Beijing, China
<b>Sponsored by</b>	Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences	<b>Distributed by</b>	Elsevier Limited, The Netherlands
<b>Edited by</b>	Editorial Office of Journal of Environmental Sciences P. O. Box 2871, Beijing 100085, China Tel: 86-10-62920553; <a href="http://www.jesc.ac.cn">http://www.jesc.ac.cn</a> E-mail: <a href="mailto:jesc@263.net">jesc@263.net</a> , <a href="mailto:jesc@rcees.ac.cn">jesc@rcees.ac.cn</a>	<b>Domestic</b>	Science Press, 16 Donghuangchenggen North Street, Beijing 100717, China Local Post Offices through China
<b>Editor-in-chief</b>	Hongxiao Tang	<b>Foreign</b>	Elsevier Limited <a href="http://www.elsevier.com/locate/jes">http://www.elsevier.com/locate/jes</a>
<b>CN 11-2629/X</b>	<b>Domestic postcode: 2-580</b>	<b>Printed by</b>	Beijing Beilin Printing House, 100083, China
		<b>Domestic price per issue</b>	<b>RMB ¥ 110.00</b>

ISSN 1001-0742

