



ISSN 1001-0742

CN 11-2629/X

2012

Volume **24**
Number **7**

JOURNAL OF
**ENVIRONMENTAL
SCIENCES**



Sponsored by
Research Center for Eco-Environmental Sciences
Chinese Academy of Sciences

JOURNAL OF ENVIRONMENTAL SCIENCES

(<http://www.jesc.ac.cn>)

Aims and scope

Journal of Environmental Sciences is an international academic journal supervised by Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences. The journal publishes original, peer-reviewed innovative research and valuable findings in environmental sciences. The types of articles published are research article, critical review, rapid communications, and special issues.

The scope of the journal embraces the treatment processes for natural groundwater, municipal, agricultural and industrial water and wastewaters; physical and chemical methods for limitation of pollutants emission into the atmospheric environment; chemical and biological and phytoremediation of contaminated soil; fate and transport of pollutants in environments; toxicological effects of terrorist chemical release on the natural environment and human health; development of environmental catalysts and materials.

For subscription to electronic edition

Elsevier is responsible for subscription of the journal. Please subscribe to the journal via <http://www.elsevier.com/locate/jes>.

For subscription to print edition

China: Please contact the customer service, Science Press, 16 Donghuangchenggen North Street, Beijing 100717, China. Tel: +86-10-64017032; E-mail: journal@mail.sciencep.com, or the local post office throughout China (domestic postcode: 2-580).

Outside China: Please order the journal from the Elsevier Customer Service Department at the Regional Sales Office nearest you.

Submission declaration

Submission of an article implies that the work described has not been published previously (except in the form of an abstract or as part of a published lecture or academic thesis), that it is not under consideration for publication elsewhere. The submission should be approved by all authors and tacitly or explicitly by the responsible authorities where the work was carried out. If the manuscript accepted, it will not be published elsewhere in the same form, in English or in any other language, including electronically without the written consent of the copyright-holder.

Submission declaration

Submission of the work described has not been published previously (except in the form of an abstract or as part of a published lecture or academic thesis), that it is not under consideration for publication elsewhere. The publication should be approved by all authors and tacitly or explicitly by the responsible authorities where the work was carried out. If the manuscript accepted, it will not be published elsewhere in the same form, in English or in any other language, including electronically without the written consent of the copyright-holder.

Editorial

Authors should submit manuscript online at <http://www.jesc.ac.cn>. In case of queries, please contact editorial office, Tel: +86-10-62920553, E-mail: jesc@263.net, jesc@rcees.ac.cn. Instruction to authors is available at <http://www.jesc.ac.cn>.

Copyright

© Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences. Published by Elsevier B.V. and Science Press. All rights reserved.

CONTENTS

Aquatic environment

Investigation of the hydrodynamic behavior of diatom aggregates using particle image velocimetry Feng Xiao, Xiaoyan Li, Kitming Lam, Dongsheng Wang.....	1157
Shellac-coated iron oxide nanoparticles for removal of cadmium(II) ions from aqueous solution Jilai Gong, Long Chen, Guangming Zeng, Fei Long, Jiuhua Deng, Qiuya Niu, Xun He.....	1165
Prediction of DOM removal of low specific UV absorbance surface waters using HPSEC combined with peak fitting Linan Xing, Rolando Fabris, Christopher W. K. Chow, John van Leeuwen, Mary Drikas, Dongsheng Wang	1174
Photo-production of dissolved inorganic carbon from dissolved organic matter in contrasting coastal waters in the southwestern Taiwan Strait, China Weidong Guo, Liyang Yang, Xiangxiang Yu, Weidong Zhai, Huasheng Hong	1181
One century sedimentary record of lead and zinc pollution in Yangzong Lake, a highland lake in southwestern China Enlou Zhang, Enfeng Liu, Ji Shen, Yanmin Cao, Yanling Li	1189
Antimony(V) removal from water by iron-zirconium bimetal oxide: Performance and mechanism Xuehua Li, Xiaomin Dou, Junqing Li	1197
Carbonaceous and nitrogenous disinfection by-product formation in the surface and ground water treatment plants using Yellow River as water source Yukun Hou, Wenhai Chu, Meng Ma	1204
Water quality evaluation based on improved fuzzy matter-element method Dongjun Liu, Zhihong Zou	1210
Formation and cytotoxicity of a new disinfection by-product (DBP) phenazine by chloramination of water containing diphenylamine Wenjun Zhou, Linjie Lou, Lifang Zhu, Zhimin Li, Lizhong Zhu	1217

Atmospheric environment

Chemical compositions of PM _{2.5} aerosol during haze periods in the mountainous city of Yong'an, China Liqian Yin, Zhenchuan Niu, Xiaoqi Chen, Jinsheng Chen, Lingling Xu, Fuwang Zhang.....	1225
Decomposition of trifluoromethane in a dielectric barrier discharge non-thermal plasma reactor M. Sanjeeva Gandhi, Y. S. Mok	1234
Transverse approach between real world concentrations of SO ₂ , NO ₂ , BTEX, aldehyde emissions and corrosion in the Grand Mare tunnel I. Ameur-Bouddabbous, J. Kasperek, A. Barbier, F. Harel, B. Hannyoyer	1240
A land use regression model incorporating data on industrial point source pollution Li Chen, Yuming Wang, Peiwu Li, Yaqin Ji, Shaofei Kong, Zhiyong Li, Zhipeng Bai	1251

Terrestrial environment

Effect of vegetation of transgenic Bt rice lines and their straw amendment on soil enzymes, respiration, functional diversity and community structure of soil microorganisms under field conditions Hua Fang, Bin Dong, Hu Yan, Feifan Tang, Baichuan Wang, Yunlong Yu.....	1259
Enhanced flushing of polychlorinated biphenyls contaminated sands using surfactant foam: Effect of partition coefficient and sweep efficiency Hao Wang, Jiajun Chen	1270
Transpiration rates of urban trees, <i>Aesculus chinensis</i> Hua Wang, Xiaoke Wang, Ping Zhao, Hua Zheng, Yufen Ren, Fuyuan Gao, Zhiyun Ouyang	1278

Environmental biology

Methanogenic community dynamics in anaerobic co-digestion of fruit and vegetable waste and food waste Jia Lin, Jiane Zuo, Ruofan Ji, Xiaojie Chen, Fenglin Liu, Kaijun Wang, Yunfeng Yang	1288
Differential fate of metabolism of a disperse dye by microorganisms <i>Galactomyces geotrichum</i> and <i>Brevibacillus laterosporus</i> and their consortium GG-BL Tatoba R. Waghmode, Mayur B. Kurade, Anuradha N. Kagalkar, Sanjay P. Govindwar	1295

Environmental catalysis and materials

Effects of WO _x modification on the activity, adsorption and redox properties of CeO ₂ catalyst for NO _x reduction with ammonia Ziran Ma, Duan Weng, Xiaodong Wu, Zhichun Si	1305
Photocatalytic degradation of bisphenol A using an integrated system of a new gas-liquid-solid circulating fluidized bed reactor and micrometer Gd-doped TiO ₂ particles Zhiliang Cheng, Xuejun Quan, Jinxin Xiang, Yuming Huang, Yunlan Xu	1317
Effect of CeO ₂ and Al ₂ O ₃ on the activity of Pd/Co ₃ O ₄ /cordierite catalyst in the three-way catalysis reactions (CO/NO/C _n H _m) Sergiy O. Soloviev, Pavlo I. Kyriienko, Nataliia O. Popovych	1327

Environmental analytical methods

Development of indirect competitive fluorescence immunoassay for 2,2',4,4'-tetrabromodiphenyl ether using DNA/dye conjugate as antibody multiple labels Zi-Yan Fan, Young Soo Keum, Qing-Xiao Li, Weilin L. Shelver, Liang-Hong Guo	1334
A novel colorimetric method for field arsenic speciation analysis Shan Hu, Jinsuo Lu, Chuanyong Jing	1341
Aminobenzenesulfonamide functionalized SBA-15 nanoporous molecular sieve: A new and promising adsorbent for preconcentration of lead and copper ions Leila Hajiaghbabaei, Babak Ghasemi, Alireza Badieli, Hassan Goldooz, Mohammad Reza Ganjali, Ghodsi Mohammadi Ziarani	1347



Effects of WO_x modification on the activity, adsorption and redox properties of CeO_2 catalyst for NO_x reduction with ammonia

Ziran Ma¹, Duan Weng^{1,*}, Xiaodong Wu^{1,2}, Zhichun Si³

1. State Key Laboratory of New Ceramics & Fine Process, Department of Materials Science and Engineering, Tsinghua University, Beijing 100084, China

2. Yangtze Delta Region Institute of Tsinghua University, Zhejiang 314000, China

3. Advanced Materials Institute, Graduate School at Shenzhen, Tsinghua University, Shenzhen 518055, China

Received 30 September 2011; revised 05 December 2012; accepted 14 December 2011

Abstract

A series of WO_3/CeO_2 (WO_x/CeO_2) catalysts were synthesized by wet impregnation of ammonium metatungstate on a CeO_2 support. The resulting solid acid catalysts were characterized by X-ray diffraction (XRD), UV-Vis spectroscopy (UV-Vis), Raman spectroscopy (Raman), *in-situ* Fourier transform infrared spectroscopy (*in-situ* FT-IR) of ammonia adsorption, NH_3 -TPD, H_2 temperature-programmed reduction (H_2 -TPR), NH_3/NO oxidation and activity measurements for NO_x reduction by NH_3 (NH_3 -SCR). The results show that polytungstate (WO_x) species are the main species of tungsten oxide on the surface of ceria. The addition of tungsten oxide enhances the Brönsted acidity of ceria catalysts remarkably and decreases the amount of surface oxygen on ceria, with strong interaction between CeO_2 and WO_x . As a result, the N_2 selectivity of NH_3 oxidation and NH_3 -SCR at high temperatures ($> 300^\circ\text{C}$) is enhanced. Therefore, a wide working temperature window in which NO_x conversion exceeds 80% (NO_x conversion $> 80\%$) from 200 to 450°C , is achieved over 10 wt.% WO_x/CeO_2 catalyst. A tentative model of the NH_3 -SCR reaction route on WO_x/CeO_2 catalysts is presented.

Key words: NO_x reduction; WO_x/CeO_2 catalyst; redox; NH_3 adsorption

DOI: 10.1016/S1001-0742(11)60925-X

Introduction

In light of the increasingly stringent emission standards for vehicle exhaust, interest in reducing NO_x emissions has received much attention in recent years. The most efficient technology to remove NO_x from stationary sources is the selective catalytic reduction of NO_x by ammonia (NH_3 -SCR), and it is also currently the most favored technology for lean burn engine exhaust deNO_x (Busca et al., 2005). Reduction of NO_x emissions from a diesel engine requires a highly efficient catalyst which operates in a temperature range of $200\text{--}500^\circ\text{C}$ and is capable of high space velocity. To meet such a requirement, a typical commercial SCR catalyst $\text{V}_2\text{O}_5\text{--}\text{WO}_3(\text{MoO}_3)/\text{TiO}_2$ (VWTi) with about 1 wt.% V_2O_5 supported on 6–10 wt.% WO_3/MoO_3 containing titania (anatase) was used first for stationary sources, and was then introduced into application in diesel vehicles (Sjövall et al., 2009; Klukowski et al., 2009; Schwidder et al., 2008). VWTi oxides have high SCR activity in the temperature range $300\text{--}500^\circ\text{C}$. Nevertheless, the high activity for oxidation of SO_2 to SO_3 , rapid decrease in activity and selectivity at higher temperatures, low stability due to the phase transformation of anatase to rutile and the toxicity of the vanadia species limit their applications in diesel

deNO_x (Brandenberger et al., 2008). Metal-exchanged zeolites, especially ZSM-5, have been studied extensively for removing NO_x from diesel vehicles. However, some inevitable problems with this catalyst system still remain; for example, their relatively low stability in hydrothermal aging, causing dealumination, leads to deactivation of the zeolite catalysts. Coke transformed from hydrocarbons can poison the active sites or block their access, and the regeneration of zeolite catalysts, requiring the removal of coke at high temperatures causes detrimental effects like dealumination, degradation of the zeolite and sintering of supported metals, etc. (Brandenberger et al., 2008; Grossale et al., 2008; Guisnet and Magnoux, 1997).

Extensive efforts have been made over the past decade to develop non-vanadium catalysts for deNO_x , especially CeO_2 -based NH_3 -SCR catalysts, which have recently been the subject of extensive development. $\text{MnO}_x\text{--}\text{CeO}_2$ mixed oxide showed better activity at low temperature, and the addition of niobium oxides into the catalyst was found to significantly enhance its low-temperature activity (Qi et al., 2004; Casapu et al., 2009). A strong interaction between ceria and titania was found in $\text{CeO}_2\text{--}\text{TiO}_2$ (Xu et al., 2009; Gao et al., 2010) and $\text{V}_2\text{O}_5\text{--}\text{CeO}_2/\text{TiO}_2$ catalysts (Huang et al., 2008), where a high concentration of amorphous Ce^{3+} on the surface increased the chemisorption of weakly bound oxygen species and resulted in good

* Corresponding author. E-mail: duanweng@tsinghua.edu.cn

activity. In addition, synergistic effects between CeO₂ and other species in catalysts, which could generate active oxygen species with high reducibility and increase acid sites, were also considered to promote SCR reaction in many CeO₂-containing catalysts synthesized by different methods (Gao et al., 2010; Liu et al., 2010; Si et al., 2010; Zhu et al., 2010; Adamowska et al., 2008; Reddy et al., 2003). Besides these studies, in recent years, WO_x- and CeO₂ containing catalysts have gained much attention as a result of their good activity. Li et al. (Chen et al., 2010) also reported CeO₂/WO₃-TiO₂ or CeO₂-WO₃-TiO₂ novel catalysts with high activity for NH₃-SCR of NO_x in the temperature range 200–450°C at a space velocity of 28,000 hr⁻¹. WO₃ supported on CeO₂-ZrO₂ solid solutions has been reported to show high SCR activity within a high temperature range (300–500°C) at a space velocity of 90,000 hr⁻¹ in NH₃-SCR applications (Li et al., 2008). He et al. (Shan et al., 2011) reported a novel CeO₂-WO₃ mixed oxide catalyst prepared by a homogeneous precipitation method which achieved nearly 100% NO_x conversion in a wide temperature range from 250 to 425°C under an extremely high GHSV of 500,000 hr⁻¹. Their work all indicated that WO₃ crystallites can be finely dispersed over the catalyst surface to enhance the acidity of ceria catalysts, and some kind of synergetic effect between tungsten oxides and ceria promoted the SCR activity by ammonia. However, the specific synergistic effects between CeO₂ and WO_x in the mixed oxide catalysts and the corresponding SCR reaction pathways have not been discussed in depth.

In the present work, a series of WO₃/CeO₂ catalysts were prepared for the selective catalytic reduction of NO_x by NH₃. The structure of CeO₂-supported WO_x was characterized through XRD, UV-Vis and Raman spectra. By means of *in-situ* FT-IR tests of NH₃ adsorption, H₂-TPR, NH₃/NO oxidation and the activity for NO_x reduction by NH₃, the effects of tungsten oxide modification on the acidity, redox and activity of ceria catalysts were analyzed and discussed.

1 Experimental

1.1 Catalyst preparation

The catalysts were synthesized by a wet impregnation method. Corresponding amounts of ammonium paratungstate (Beijing Beihua, China), ceria power (American BASF, 310 m²/g) and oxalic acid (Beijing Beihua, China) were used as precursors. Ammonium paratungstate, oxalic acid and deionized H₂O were mixed by the molar ratio of 1:1.5:100 to obtain a clear solution, and then the corresponding amount of CeO₂ powders was added. The mixture was stirred at 100°C until all H₂O had evaporated. After that, the resulting powders were dried at 100°C overnight and subjected to calcination at 600°C for 4 hr in static air. The catalysts were denoted as (x)WO_x/CeO₂ (XWC), where *x* represented the mass percentage of WO₃. The pure ceria powders were also calcined at 600°C for 4 hr in static air for use as a reference.

1.2 Activity measurement

The activity measurements for NH₃-SCR were carried out in a quartz reactor with 200 mg of catalyst (diluted to 1 mL by silica) inside. The reaction gas mixture consisted of 500 ppm NO, 500 ppm NH₃, 5% O₂ and N₂ in balance. The total flow rate of the gas mixture was 500 mL/min at a gas hourly space velocity (GHSV) of 30,000 hr⁻¹. The SCR activities of the catalysts were measured under steady-state conditions at various temperatures. The concentrations of NO_x and NH₃ were measured by a Thermo Nicolet 380 Fourier transformed infrared spectroscopy (FT-IR) detector after the NH₃-SCR reaction reached a steady state. NO_x conversion (deNO_x) was defined as follows:

$$\text{NO}_x \text{ conversion} = \frac{C_{\text{NO-in}} - C_{\text{NO-out}} - C_{\text{NO}_2\text{-out}} - 2C_{\text{N}_2\text{O-out}}}{C_{\text{NO-in}}} \times 100\% \quad (1)$$

1.3 Catalysts characterization

X-ray diffraction (XRD) experiments were performed on a Japan SHIMADZU s-7000 diffractometer employing Cu K α radiation ($\lambda = 0.15418$ nm) operated at 40 kV and 120 mA. The X-ray diffractograms were recorded at 0.02° intervals in the range of 20° ≤ 2 θ ≤ 80° with a scanning velocity of 4°/min. The crystal phases were identified using JCPDS (Joint Committee on Powder Diffraction Standards) cards, and the lattice constants and mean crystallite sizes of ceria were calculated from Cohen's method and the Williamson-Hall equation.

Brunauer-Emmett-Teller (BET) surface areas were calculated from N₂ adsorption data obtained using an F-Sorb 3400 apparatus (Gold APP Instrument) at -196°C. Before the measurement, the sample was degassed in N₂ at 200°C for 2 hr.

The Raman spectra (Raman) were obtained with a LabRAM HR 800 (HORIBA Jobin Yvon, France) spectrometer at room temperature (RT) and atmospheric pressure. An argon ion laser beam with the wavelength of 488 nm was focused on a spot 1 μ m in diameter.

The UV-Vis spectra (UV-Vis) were measured in diffuse reflectance mode using a Shimadzu UV-2100S spectrometer with an internal integration sphere. A BaSO₄ pellet was used as a reference. The spectra were recorded at RT in the spectral range 200–800 nm.

H₂ temperature-programmed reduction (H₂-TPR) was performed in a fixed-bed reactor with the effluent gases monitored using a quadrupole mass spectrometer (MS) (Omnistar 200). Prior to the H₂-TPR experiment, the 50 mg sample was treated with 2 vol.% O₂/He with a total flow rate of 50 mL/min at 550°C for 30 min, then cooled down to RT in the same atmosphere, and subsequently flushed with 50 mL/min He for 30 min to remove the physically adsorbed molecules. Finally, the reactor temperature was raised to 900°C at a constant heating rate of 10°C/min in 5 vol.% H₂/He with a flow rate of 50 mL/min. H₂ consumption during the experiment was monitored by MS.

Ammonia temperature-programmed desorption (NH₃-TPD) experiments were carried out with a Nicolet-380 FT-IR spectroscopy detector. Before the experiment, 0.2 g catalyst was treated in 20% O₂/N₂ at 500°C for 30 min. After cooling down to RT, the sample was exposed to 1000 ppm NH₃/N₂ for 30 min and flushed with N₂ for 30 min. The experimental runs were recorded in N₂ while heating to 500°C at a heating rate of 10°C/min.

In-situ FT-IR spectra for NH₃ chemisorption were recorded using a thermo Nicolet 6700 FT-IR spectrometer. The sample was pretreated at 500°C in N₂ for 30 min. After cooling down to RT, the sample was flushed with N₂ for 30 min for background collection. Then a gas mixture containing 1000 ppm NH₃ and N₂ (100 mL/min) was passed through the sample for 1 hr. The IR spectra were collected after purging with N₂ flow for 30 min.

Ammonia oxidation experiments were carried out using a similar method to the NH₃-SCR measurement in an atmosphere of 500 ppm NH₃ and 5% O₂ in N₂. The gas concentrations were recorded after the oxidation reaction reached a steady state. Selectivity for the formation of N₂ during NH₃ oxidation (S_{N_2} , %) was defined by the following equation:

$$S_{N_2} = 1 - \frac{C_{NO_2-out} + C_{NO-out} + 2C_{N_2O-out}}{C_{NH_3-in} - C_{NH_3-out}} \times 100\% \quad (2)$$

NO oxidation experiments were carried out using a method similar to the NH₃-SCR measurement, with 500 ppm NO and 5% O₂ in N₂.

2 Results

2.1 SCR activity

The NH₃-SCR performance of WO_x/CeO₂ catalysts was measured as a function of temperature (Fig. 1a). Pure CeO₂ catalyst shows rather poor SCR activity over the whole temperature range. The SCR activities of the WO₃-modified catalysts are clearly promoted. Among the catalysts investigated, 10WC catalyst shows the highest SCR activity, obtaining over 80% NO_x conversion over a wide temperature interval of 200–450°C, which is compa-

rable to the performance of commercial V₂O₅/WO₃-TiO₂ catalyst. The N₂ selectivity of the catalysts follows the order 20WC > 10WC > 5WC > CeO₂, indicating that the modification with tungsten oxide could inhibit the formation of N₂O and NO₂ during the SCR process. The amount of NO₂, which is mainly the product of ammonia oxidation and NO oxidation, decreases with increasing tungsten oxide loading. No N₂O is observed over the WO_x-modified catalysts, while considerable N₂O is generated over the CeO₂ catalyst. All these results indicate that a synergistic effect on the NH₃-SCR reaction exists between the CeO₂ support and WO_x.

2.2 XRD

Figure 2 shows the XRD patterns of the catalysts with different WO₃ loading. The symmetrical peaks of all samples are consistent with the characteristic peaks of fluorite-like ceria. No obvious peaks of tungsten oxides are observed, indicating that tungsten oxides are finely dispersed on the surface of ceria. According to the variations in the average crystallite size of ceria as shown in Table 1, the WO₃ addition seems to inhibit the sintering of CeO₂. However, the blocking effect of the CeO₂ support pores by the impregnated tungsten oxide accounts for the reduced surface area of the modified catalysts.

The lattice constant of ceria was also calculated according to the Williamson-Hall equation and the results are summarized in Table 1. Generally, the lattice constant of ceria is associated with the ratio of Ce⁴⁺/Ce³⁺ and possible incorporation of Wⁿ⁺ into the ceria lattice. Considering the smaller ionic radii of W⁶⁺ (0.060 nm) and W⁵⁺ (0.062 nm) compared to those of Ce⁴⁺ (0.087 nm) and Ce³⁺ ions (0.110 nm), the incorporation of Wⁿ⁺ cations into the CeO₂ lattice may result in shrinkage of the CeO₂ lattice.

Table 1 Structural and textural properties of the catalysts

Sample	BET surface area (m ² /g)	Crystallite size of CeO ₂ (nm)	Lattice constant of CeO ₂ (nm)
CeO ₂	117	13.3	0.5415
5WC	92	10.8	0.5407
10WC	56	13.4	0.5406
20WC	53	10.4	0.5405

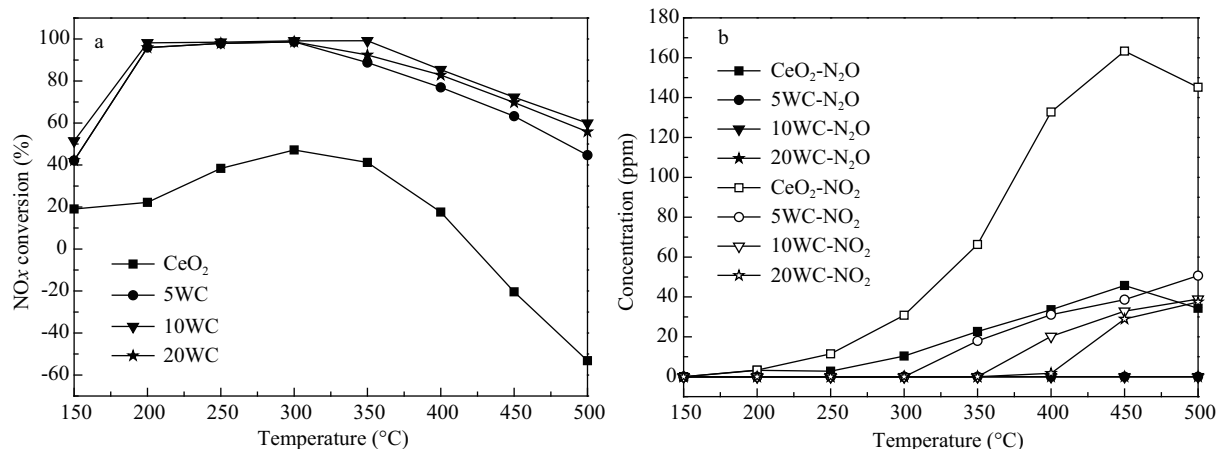


Fig. 1 SCR activity of CeO₂ and WO_x/CeO₂ catalysts. (a) NO_x conversion; (b) NO₂ and N₂O outlet concentration. Reaction conditions: [NO] = [NH₃] = 500 ppm, [O₂] = 5%, N₂ in balance, GHSV = 30,000 hr⁻¹. 5WC, 10WC, 20WC mean the mass percentages of WO₃ are 5%, 10%, 20%, respectively.

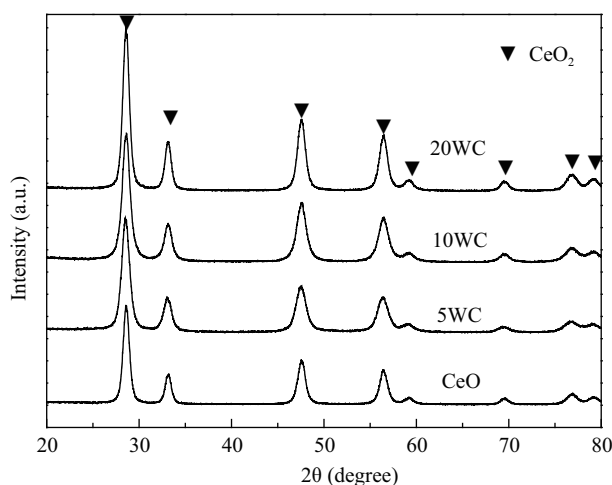


Fig. 2 XRD patterns of the catalysts.

Alternatively, the shrinkage of the ceria lattice could also be due to the decreased amount of Ce^{3+} ions caused by the interaction with tungsten oxide.

2.3 UV-Vis spectra

Figure 3 shows the UV-vis spectra of the catalysts. Ceria shows an absorption edge corresponding to the O_{2p} - Ce_{4f} transition with a band gap of 2.9 eV (Cormal et al., 2004; Tsunekawa et al., 2000). All the tungsten-containing catalysts show an absorption edge between 405 and 456 nm, with a band gap of 3.0, 2.8 and 2.5 eV for 5WC, 10WC and 20WC, respectively, which may correspond to the harmonics and combination absorption edges of the ligand-to-metal charge transfer (O_{2p} - W_{5d} - O_{2p}) of WO_x species and O_{2p} - Ce_{4f} charge transfer of CeO_2 . The absorption may be ascribed to defects resulting from a partial dissolution occurring at the interface of the two oxide phases. 5WC shows the lowest wavelength of the absorption edge at 405 nm. The blue shift of the absorption edge compared to CeO_2 , with a band gap of 3.0 eV in 5WC, mainly results from the decrease of the $\text{Ce}^{3+}/\text{Ce}^{4+}$ ratio in the

catalyst (Tsunekawa et al., 2000). The lattice relaxation of ceria by tungsten modification may induce the oxidation of Ce^{3+} to Ce^{4+} , decreasing the charge transfer gap between the O_{2p} and Ce_{4f} bands. Such an absorption edge shift depends on the extent of WO_3 loading, which is consistent with the literature (Singh and Singh, 2011; Yagoubi and Hogarth, 1993). The gradual increase in the surface density of tungsten oxide increases the concentrations of octahedral $[\text{WO}_6]$ and tetragonal $[\text{WO}_4]$ units, which causes the creation of a large number of donor centers which can accept electrons from O_{2p} and then donate them to Ce_{4f} . As a result, the impurity band becomes more extended into the main band gap. This development shifts the absorption edge towards lower energies, leading to a significant shrinkage of the band gap. This can be attributed to the strong electron interaction between CeO_2 and WO_x species. Moreover, the increased size of WO_x domains, which cannot be detected in XRD patterns, can also result in continuously decreasing E_g values (Lin et al., 2009; Barton et al., 1999; Onfroy et al., 2005).

2.4 Raman spectra

Figure 4 shows the Raman spectra of the catalysts. In contrast to the XRD patterns that give information related mainly to the cation sublattice, the Raman spectrum is dominated by the oxygen sublattice, the Raman spectrum is dominated by the oxygen lattice vibration of CeO_2 and the surface structure of WO_x on ceria (Wachs et al., 2006; Wang et al., 2007; Bigey et al., 2001). The band around 460 cm^{-1} is ascribed to the F_{2g} Raman-active mode of the cubic fluorite structure (Wang et al., 2007; Mineshige et al., 2000; Vidal et al., 2000). This band shifts from 455 cm^{-1} for CeO_2 to 461, 463 and then back to 460 cm^{-1} for 5WC, 10WC and 20WC, respectively. These shifts of peaks indicate a similar trend of variation in the crystallite size of ceria, which matches well with the XRD results in Table 1 (Reddy et al., 2002, 2003). It is known that the intensity of Raman band depends on several factors, including the grain size and morphology.

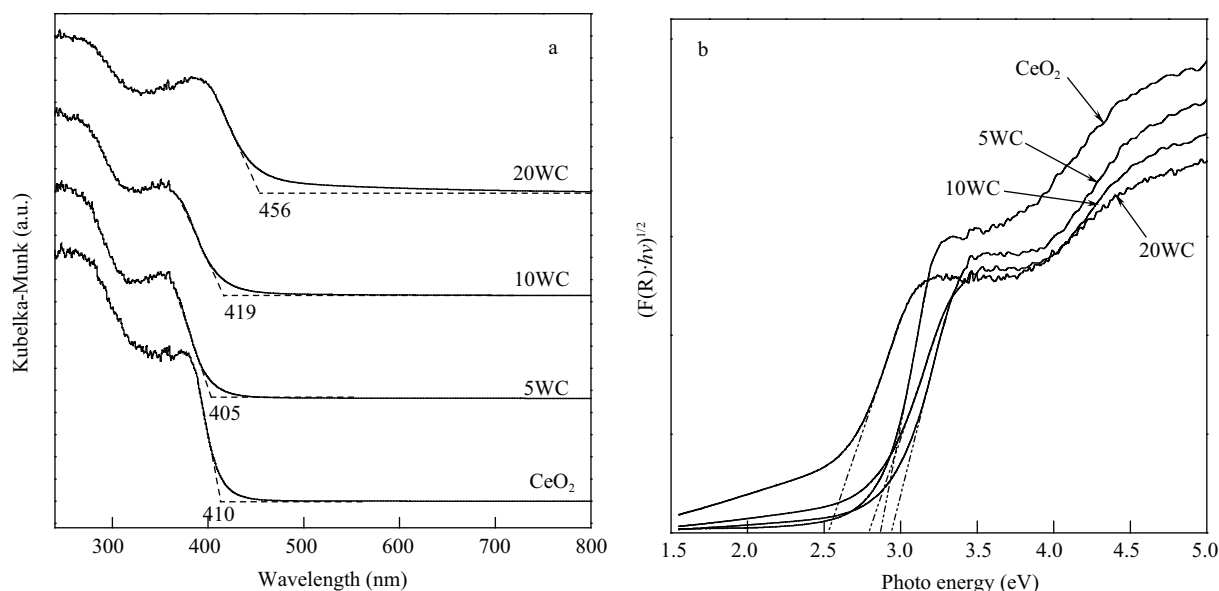


Fig. 3 UV-Vis spectra of WO_x/CeO_2 catalysts. (a) plotted as the Kubelka-Munk function vs. wavelength; (b) plotted as the $(F(R)h\nu)^{1/2}$ function vs. photon energy.

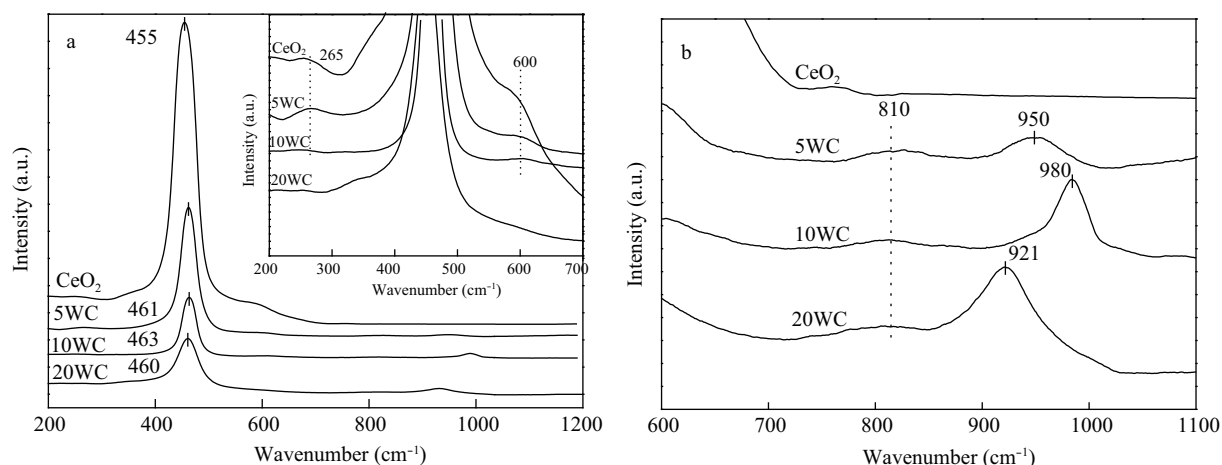


Fig. 4 Raman spectra of the WO_x/CeO₂ catalysts in the wavenumber range of (a) 200–1200 and (b) 600–1100 cm⁻¹.

However, all these F_{2g} bands are weakened and become less symmetrical with the increase of WO_x content, as seen in Fig. 4a. Considering the increasing size of ceria crystallites as reflected by the XRD results, it is reasonable to suggest that the surface content of ceria decreases as a consequence of WO_x coverage and a stronger interaction between CeO₂ and surface tungsten oxide species. The bands at 265 and 600 cm⁻¹ indicate the presence of some defects in CeO₂, which could lead to a relaxation of selection rules. In particular, these bands have been linked to oxygen vacancies in the CeO₂ lattice (Wang et al., 2007; Mineshige et al., 2000; Vidal et al., 2000) which disappear when the WO₃ loading exceeds 5%.

Raman spectroscopy gives additional structural information about tungsten species in catalysts. Tungsten-containing catalysts exhibit a band at 810 cm⁻¹ assigned to W–O stretching modes (A_{1g}) (Bigey et al., 2001; Mamede et al., 2004; Huang et al., 2007; Cortés-Jácome et al., 2007), which is a typical characteristic of polytungstate species possessing [WO₅/WO₆] coordination (Wachs et al., 2006; Bigey et al., 2001; Mamede et al., 2004; Huang et al., 2007). The bands at 921–1000 cm⁻¹ are ascribed to terminal W=O groups, which correspond to [WO₄] coordination in polytungstate species (Wachs et al., 2006; Bigey et al., 2001; Mamede et al., 2004; Huang et al., 2007). No bands of nano-crystallite WO₃ particles are observed, as indicated by the absence of bands at 714 cm⁻¹ assigned to W–O stretching modes (E_g) and 271 cm⁻¹ assigned to the W–O–W bending mode (F_{2g}) (Lin et al., 2009; Wachs et al., 2006; Bigey et al., 2001; Mamede et al., 2004). This indicates a high dispersion of WO_x on these catalysts in the form of surface polytungstate species, which correlates well with the XRD and UV-Vis results. Compared with 5WC, the band assigned to the W=O vibration for the 10WC sample shifts from 950 to 980 cm⁻¹. This W=O band becomes larger with increasing surface density of tungsten oxide, which is related to polymerization of the surface WO_x species (Wachs et al., 2006; Kim et al., 2007). However, the W=O vibration band shifts to lower wavenumber at 921 cm⁻¹ for 20WC, implying a strong interaction between ceria and tungsten in parallel with the decrease of surface CeO₂ content, which is found to be similar to a published report that the

typical band of a W=O vibration affected by Ce–O bonds is located near 920 cm⁻¹ (Mamede et al., 2007).

2.5 NH₃ adsorption

NH₃-derived species on the WO_x-CeO₂ catalysts were investigated by FT-IR spectroscopy, and the results are shown in Fig. 5. Bands at 1437, 1663, 2791, 3000 and 3196 cm⁻¹, which are only observed on the tungsten-containing catalysts, are assigned to NH₄⁺ species on Brönsted acid sites (Zhu et al., 2010; Ramis et al., 1999; Vargas et al., 2007). These bands of NH₄⁺ species increase with increasing WO_x loading, indicating that Brönsted acid sites mainly arise from polymerized WO_x. Bands at 1061–1241 and 3393 cm⁻¹ are attributed to NH₃ coordinated to Lewis acid sites (Mamede et al., 2004; Ramis et al., 1999; Vargas et al., 2007) which are found on both ceria and the WO_x-modified ceria, and their intensity decreased with increased WO_x loading. Bands at 1554, 1300 and 1136 cm⁻¹ are assigned to the NH₂ shear vibration, swing vibration and rocking vibration, respectively (Zhu et al., 2010; Ramis et al., 1999). These results clearly indicate that Lewis acid sites are much more prevalent on the CeO₂ catalyst, while WO_x-CeO₂ catalysts mainly possess Brönsted acidity. WO₃ has been reported as an important additive that increases the amount and strength of Brönsted acid sites for V₂O₅-WO₃/TiO₂ catalysts (Sjövall et al., 2009; Klukowski et al., 2009; Schwidder et al., 2008). The negative band at 1003 cm⁻¹ is typical for the fundamental overtone of the W=O stretching mode of surface polytungstate species (Onfroy et al., 2005; Vargas et al., 2007). Besides these, large amounts of NH₂ are present on pure ceria due to the de-protonation of ammonia coordinated on Lewis acid sites by active surface oxygen on ceria. No bands of NH₂ are observed when the loading amount of tungsten oxide exceeds 10%, suggesting that ammonia oxidation may be inhibited by tungsten oxide modification.

NH₃-TPD experiments were performed to evaluate the acidity of the catalysts. The low-temperature (100–200°C) peaks around 130°C are ascribed to the desorption of physically weakly-adsorbed ammonia, which is prone to desorb at temperatures above 200°C. Therefore, this kind of adsorbed ammonia cannot be used to evaluate the acidity of the catalysts. The high-temperature peaks are ascribed

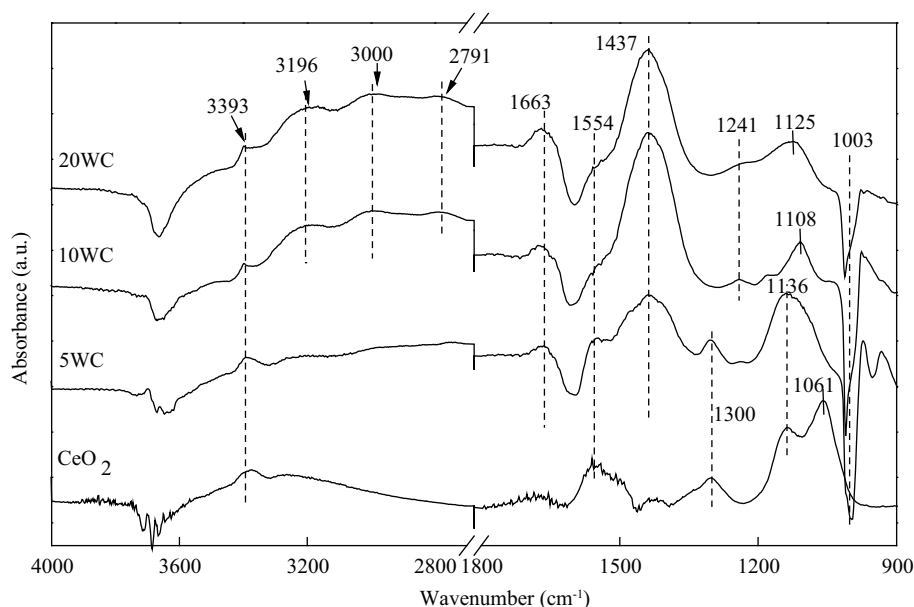


Fig. 5 FT-IR spectra of ammonia derived species on catalysts at room temperature.

to chemisorbed ammonia, which can reflect information on acid sites on catalysts. As shown in Table 2, compared with CeO_2 , the acid site density (total number of acid sites per unit BET area) of 5WC decreased from 0.88 to $0.70 \mu\text{mol}/\text{m}^2$. This is ascribed to the blocking of coordinately unsaturated Ce^{4+} (Lewis acid sites) by the supported tungsten oxides (Onfroy et al., 2005; Cortés-Jácome et al., 2007), which can explain the decreasing intensity of bands attributed to Lewis acid sites. For the WO_x -containing catalysts, with the increase of the tungsten content, more ammonia is desorbed, representing the increase of acidity of the catalysts. Combined with the results above, stronger acidity is mainly brought about by the greater number of Brönsted acid sites.

Table 2 Acidity of catalysts derived from NH_3 -TPD experiment

Catalyst	WO_3 loading (wt.%)	Acidity	
		Acid site density ($\mu\text{mol}/\text{m}^2$)	T_m ($^\circ\text{C}$)
CeO_2	0	0.88	192
5WC	5	0.70	227
10WC	10	0.80	251
20WC	20	0.98	235

2.6 In-situ DRIFTS of NH_3 adsorption

Figure 6 shows DRIFT spectra of NH_3 -derived species on CeO_2 and 10WC catalysts by N_2 purging at various temperatures. In Fig. 6a, bands at 1560, 1310 and 1138 cm^{-1} are ascribed to NH_2 (scissoring, wagging and rocking modes respectively) arising from ammonia oxidation on the surface of ceria. The band at 1067 cm^{-1} is contributed to NH_3 (σ_{as}) on Lewis acid sites (Zhu et al., 2010; Mamede et al., 2004; Ramis et al., 1995). The weak bands at 1435 and 1666 cm^{-1} are attributed to NH_4^+ (σ_{s}) and NH_4^+ (σ_{as}) on Brönsted acid sites (Mamede et al., 2004; Ramis et al., 1995). As indicated by the bands of NH_3 -derived species, N_2 purging at 300°C causes the overall desorption of NH_3 and NH_4^+ species on ceria.

Figure 6b shows the spectra of NH_3 -derived species on the 10WC catalyst. Both coordinated ammonia ($\text{NH}_3(\sigma_{\text{s}})$: 1254, 1178 and 1117 cm^{-1}) and ammonium ions (NH_4^+ (σ_{s}): $1417\text{--}1436 \text{ cm}^{-1}$; NH_4^+ (σ_{as}): 1672 cm^{-1}) are observed on the catalyst. Two bands for coordinated ammonia indicated that ammonia is coordinated on two different Lewis acid sites, i.e., Ce^{n+} and W^{n+} . Compared with CeO_2 in Fig. 6a, tungsten modification leads to the formation of more ammonium ions involving W-OH centers, as indicated by the intense band at 1436 cm^{-1} . In addition, no protonation of NH_3 (NH_2) is observed, indicating that ammonia oxidation on ceria is effectively inhibited by WO_x modification. The adsorbed NH_3 species could even exist stably at 400°C after N_2 purging, indicating that tungsten oxide modification can enhance the stability of adsorbed NH_3 on CeO_2 , which, to some extent, favors the SCR reaction at high temperatures.

2.7 H_2 -TPR

Figure 7 shows the H_2 -TPR profiles of the catalysts. Pure ceria presents a sharp reduction peak at 320°C and a broad peak at about 770°C . The low-temperature peak is associated with the reduction of surface active oxygen, and the high-temperature one is attributed to the reduction of bulk oxygen in ceria (Zhu et al., 2010). All WO_3/CeO_2 catalysts show two defined peaks in the temperature range of $500\text{--}600^\circ\text{C}$ and $700\text{--}800^\circ\text{C}$. Bulk WO_3 can be reduced by H_2 via multiple steps: $\text{WO}_3 \rightarrow \text{WO}_{2.9} \rightarrow \text{WO}_2 \rightarrow \text{W}$, with three reduction peaks at 430°C ($\text{WO}_3 \rightarrow \text{WO}_{2.9}$), 630°C ($\text{WO}_{2.9} \rightarrow \text{WO}_2$) and 730°C ($\text{WO}_2 \rightarrow \text{W}$) (Barton et al., 1999; Santiesteban et al., 1999). Accordingly, the peak at $500\text{--}585^\circ\text{C}$ is assigned to the WO_3 to WO_2 reduction process and the reduction of some surface ceria. The broad overlapping peak centered at 730°C is attributed to the reduction of WO_2 to W as well as bulk CeO_2 . Clearly, the reduction peak of surface CeO_2 decreases in intensity and shifts to higher temperatures from 320 to 500°C at the very start of the addition of tungsten oxide.

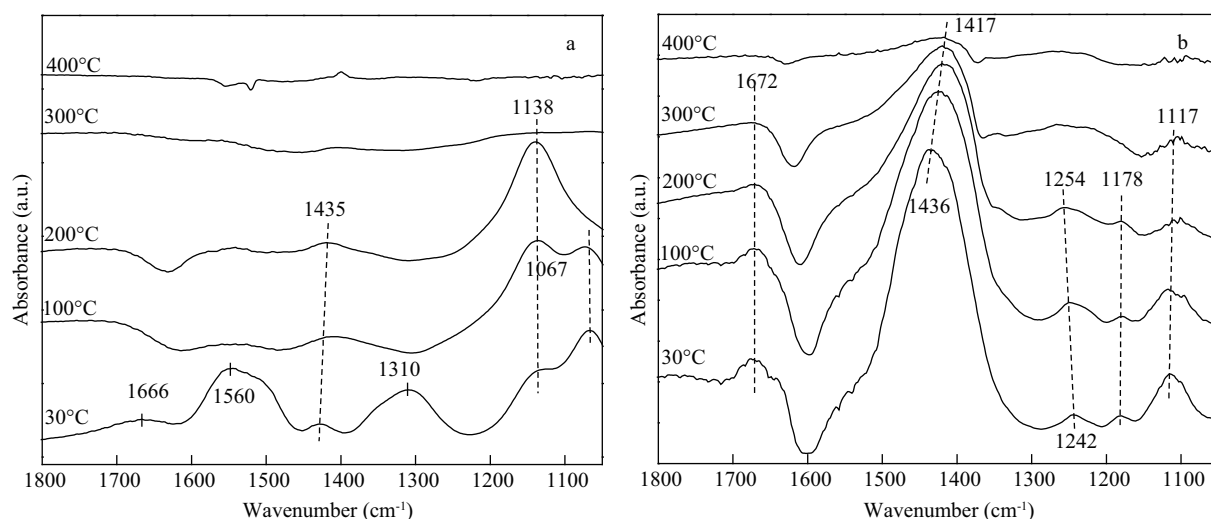


Fig. 6 DRIFT spectra of NH₃ derived species over CeO₂ (a) and 10WC (b) catalyst arising from contact of NH₃ at different temperatures.

indicating that the active surface oxygen on the CeO₂ surface is diminished by WO_x modification to a large extent. The first reduction peak of the WO_x-containing catalysts shifts to high temperatures with increased WO_x loading, revealing a strong interaction between CeO₂ and WO_x species.

Table 3 lists the H₂ reduction characteristics of the samples, of which the last column is the theoretical consumption of H₂ due to reduction of WO₃ to WO₂. The reduction of WO₃ to WO₂ is always considered to be sensitive to the size of WO_x domains: the reduction becomes easier with the increased size of WO_x domains on WO_x/ZrO₂, WO_x/SiO₂ and WO_x/Al₂O₃, with the reduction of WO₃ to WO₂ at about 630°C (Barton et al., 1999; Santiesteban et al., 1999). In this work, the reduction peak ranged from 500–585°C for all WO_x/CeO₂ catalysts, which is lower than the above results. The initial reduction temperature of these dispersed WO_x species is determined by the strength of the W–O bonds that contain oxygen shared with the support (Vargas et al., 2007). It should be noted that unlike WO_x loaded on irreducible supports, the active oxygen from W–O–Ce domains is easily reduced by H₂ due to the promotion effects of CeO₂ on W⁶⁺ reduction in WO_x/CeO₂ catalysts. Therefore, the lower reduction peak in WO_x/CeO₂ catalysts can be considered as a reduction peak of the bridging oxygen from W–O–W and W–O–Ce domains. For example, more H₂ (0.305 mmol/g) is consumed at 500°C for the low-WO_x-loading sample (5WC) than the theoretical H₂ consumption for reduction of WO₃ to WO₂ (0.216 mmol/g), demonstrating the above assumption. All these results show that WO_x modification increases the onset temperature of WO_x/CeO₂ catalysts

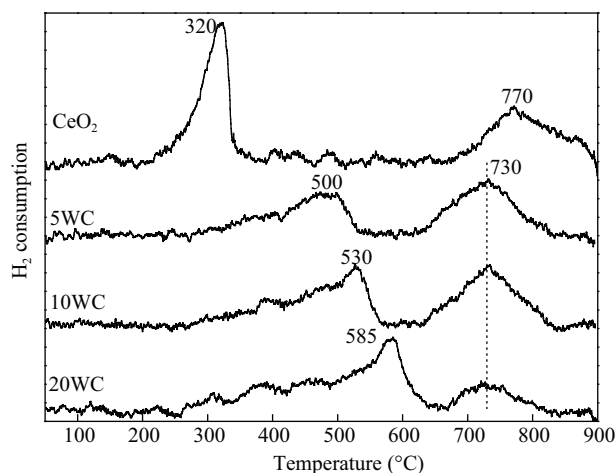


Fig. 7 H₂-TPR profiles of WO_x/CeO₂ catalysts.

during H₂-TPR, while more available bridging oxygens are generated in W–O–W and W–O–Ce domains. Therefore, the reduction peak of WO₃ to WO₂ shifts to higher temperatures and the H₂ consumption of this process increases.

2.8 NH₃ oxidation and NO oxidation

Generally, ammonia oxidation and ammonia consumption by NO reduction proceed in parallel on acid sites of catalysts during the SCR reaction. It is widely accepted that at low temperatures the reaction between ammonia and NO is dominant, whereas at high temperatures oxygen becomes a more powerful oxidant than NO for ammonia oxidation (Liu et al., 2006). It is clear in Fig. 8a that the NH₃ oxidation activity of the ceria catalyst is lowered by the addition of tungsten oxide. Among the catalysts

Table 3 Redox properties of the catalysts derived from H₂-TPR results

Sample	Low-temperature peak		High-temperature peak		Theoretical H ₂ consumption: reduction of WO ₃ to WO ₂ (mmol/g)
	Temperature (°C)	H ₂ consumption (mmol/g)	Temperature (°C)	H ₂ consumption (mmol/g)	
CeO ₂	320	0.39	770	0.62	–
5WC	500	0.31	730	0.31	0.22
10WC	530	0.42	733	0.35	0.43
20WC	585	0.59	727	0.14	0.86

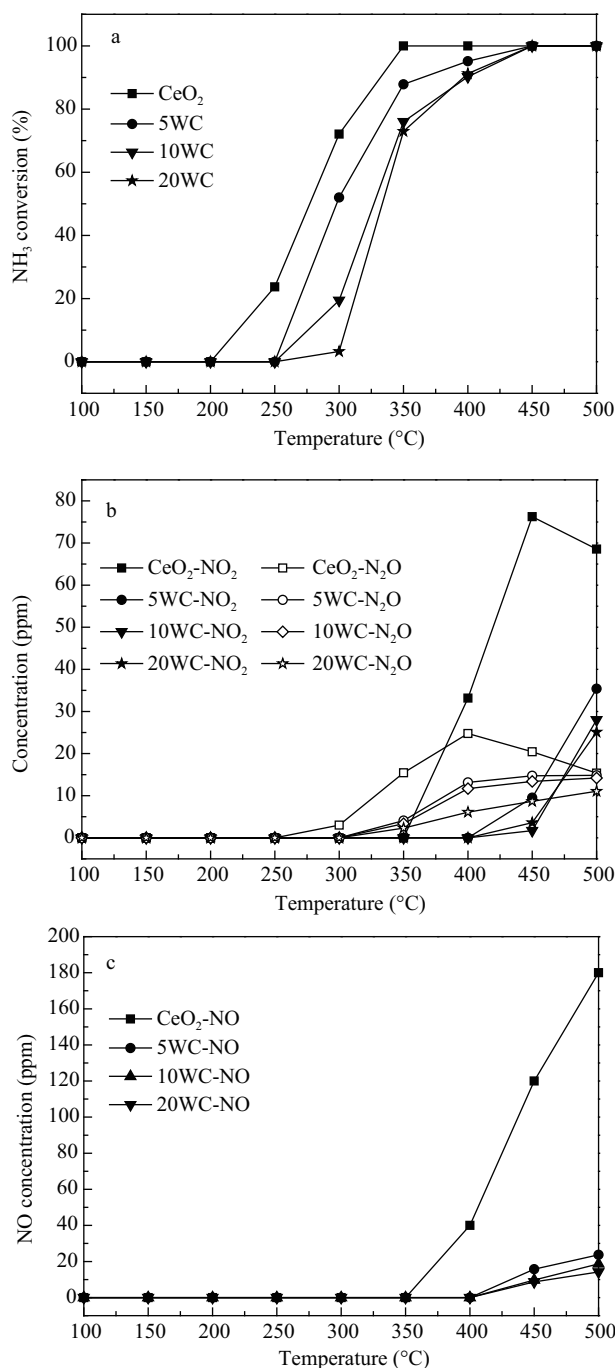


Fig. 8 NH₃ oxidation profiles of WO_x/CeO₂ catalysts. (a) NH₃ conversion; (b) NO₂ and N₂O outlet concentrations; (c) NO outlet concentration.

investigated, 20WC has the lowest ammonia oxidation activity, which correlates well with the highest N₂ selectivity at high temperatures. The lower activity of ammonia oxidation for WO_x/CeO₂ catalysts ensures the availability of ammonia in the SCR process at high temperatures. In Fig. 8b and c, it can be seen that ammonia could hardly be oxidized to NO, NO₂ and N₂O on all the WO_x-modified catalysts, while large amounts of over-oxidized products including NO, NO₂ and N₂O are formed on ceria. The higher selectivity for WO_x/CeO₂ catalysts in ammonia oxidation decreases the amount of unexpected outlet by-products.

Figure 9 shows the NO₂ concentration profiles during the NO-TPO process. NO₂ generation over ceria starts

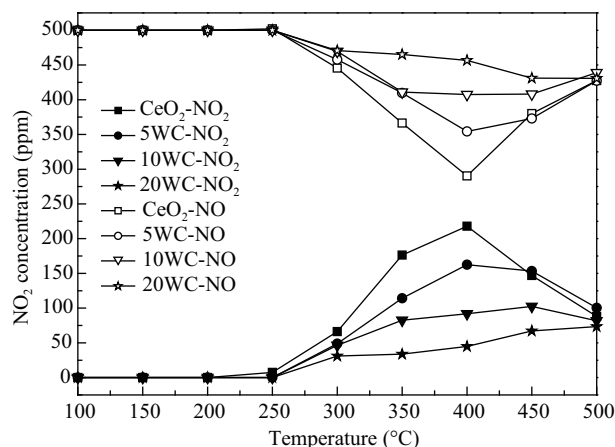


Fig. 9 NO₂ and NO outlet concentration of WO_x/CeO₂ catalysts in the NO oxidation experiment.

at 250°C and reaches a maximum (250 ppm) at about 400°C due to the strong oxidation of NO. With increasing tungsten oxide loading, the oxidation of NO is gradually inhibited. The amount of NO₂ derived from NO oxidation decreases significantly with tungsten oxide modification, which can be considered to be the reason for the high N₂ selectivity obtained with the W-containing catalysts in the SCR reaction.

3 Discussion

3.1 Structure of WO_x/CeO₂ catalysts

According to the WO₃ loading, the apparent surface densities of WO_x are in range of 1.4–9.8W atoms/nm² for the WO_x/CeO₂ catalysts synthesized by the impregnation method, regardless of the water used for impregnation and calcination temperatures, due to the similar BET surface area of catalysts. Some researchers found that a definite crystalline tungsten compound Ce₂(WO₄)₃ is preferentially formed when the surface density of W exceeds 10.8 atoms/nm² in WO_x/CeO₂ catalysts (Mamede et al., 2004). In the present work, the formation of neither crystalline WO₃ nor the tungsten compound Ce₂(WO₄)₃ was detected by XRD; however, the results of UV-Vis and Raman imply the strong interaction of CeO₂ and WO_x on the WC catalysts. For the given loading amount of WO₃, a stronger W-Ce interaction is created, along with more bridging oxygen in the W-Ce interface accompanied by the oxidation of Ce³⁺ to Ce⁴⁺.

It has been reported that if the surface coverage of WO_x on supports like ZrO₂ and Al₂O₃ is lower than a monolayer, the dominant WO_x species are monomeric or polymeric tetrahedrally coordinated tungsten oxide. Tetrahedral WO_x species transform to octahedrally coordinated WO_x when the monotungstate polymerizes to polytungstate species. With a further increase in the tungsten surface density, WO₃ nano-particles or even bulk WO₃ may form via the polymerization of surface WO_x species (Onfroy et al., 2005; Kim et al., 37; Gregorio and Keller, 2004). The XRD, UV-Vis and Raman results suggest that only highly dispersed polytungstate species formed on the WC catalysts, and no crystalline WO₃ was detected.

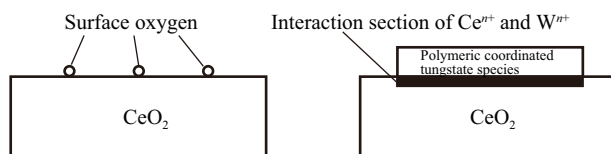


Fig. 10 Scheme of the structure evolution of CeO₂ to WO_x/CeO₂.

Figure 10 shows a scheme for the structural evolution of CeO₂ to WO_x-containing catalysts. WO_x coverage reduces the amount of surface active oxygen, and at the same time the stronger electron interaction between Ceⁿ⁺ and Wⁿ⁺ reduces the amount of oxygen vacancies and Ce³⁺. The surface WO_x polytungstate species, composed of polymeric octahedral [WO₅/WO₆] species connected by W–O bonds and tetrahedral [WO₄] species connected by W–O and W=O bonds, are highly dispersed on the catalyst surface. These highly dispersed WO_x species could increase the amount of active sites on the catalyst surface for reactant adsorption and reaction.

3.2 Acidity properties

In the NH₃ adsorption experiment, both Lewis and Brønsted acid sites were observed on the WO_x-CeO₂ catalysts. The acidity of catalysts increased with WO_x loading, and reached the maximal acidity for the 20WC catalyst. According to the literature reports (Onfroy et al., 2005; Gregorio and Keller, 2004; Martin et al., 1998), the Lewis acid sites on WO_x/CeO₂ arise from the electronically deficient unsaturated Ceⁿ⁺ cations and Wⁿ⁺ cations in the CeO₂ and dispersed octahedral polymeric [WO₆] species. The Brønsted acid sites arise from the partially hydrated tungsten species, like the hydration of bridging Ce–O–W and W–O–W in octahedral polymeric [WO₆] species and terminal W=O in tetrahedral [WO₄] species. The acid sites on WO_x-CeO₂ catalysts are able to undergo a facile transformation from Lewis to Brønsted acid sites when WO_x species interact with NH₃ or H₂O (Reactions (3) and (4)). The initial decrease in Lewis acid sites on 5WC compared with CeO₂ may arise from the blocking of coordinatively unsaturated Ceⁿ⁺ (Lewis acid sites) by the supported tungsten oxide. With further increase in the WO_x loading, the formation of octahedral polymeric [WO₆] domains contributes to the increase in the amounts of both Lewis and Brønsted acid sites.

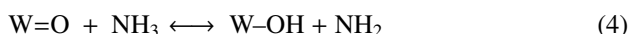
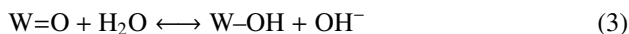


Figure 11 shows the relationship between the total acidity of the catalyst and the WO_x surface density. The acidity of the catalyst was calculated from the ammonia desorption per unit BET surface of catalyst in the NH₃-TPD experiment. The (Lewis and Brønsted) acidity of the catalyst behaves as a function of the WO_x surface density. According to the previous reports (Wachs et al., 2006; Kim et al., 2007), below monolayer coverage of WO_x on WO_x/ZrO₂ model catalysts, the surface acidity of the catalyst is mainly brought about by the polymerized WO_x species, and increases linearly with the surface density of tungsten oxide. Basically, our results demonstrate that 20

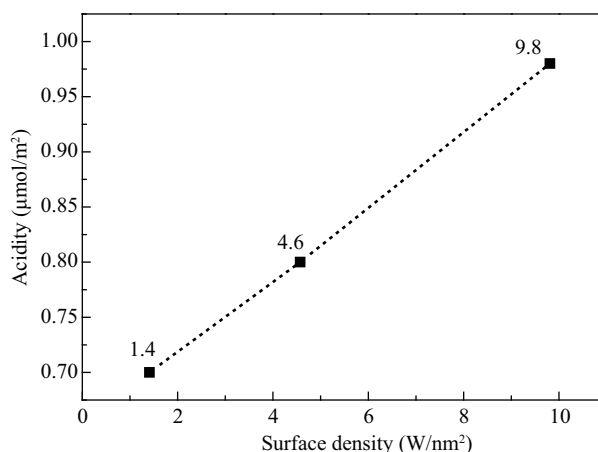


Fig. 11 The trend of the catalyst acidity as a function of tungsten oxide surface density.

wt.% WO_x content (9.8 W atoms/nm²) does not exceed the upper limitation of monolayer surface coverage of WO_x on the CeO₂ support.

3.3 Redox properties

It is seen from the H₂-TPR results that the amount of surface active oxygen on CeO₂ is reduced by WO_x addition. CeO₂ is considered to possess good OSC ability. The oxygen storage capacity is tightly related to the oxygen transfer process from bulk ceria to the surface in which oxygen vacancies play an important role (Vidal et al., 2000; Reddy et al., 2002, 2003; Schulz et al., 2003). The oxygen vacancies are formed through the process as Eq. (5):



According to the results of XRD, UV-Vis and Raman measurements, the impregnation of WO_x on the ceria surface causes a higher degree of ceria crystallization, a lower amount of Ce³⁺ and a lower amount of oxygen vacancies in ceria, which can be ascribed to the electron-withdrawing nature of W⁶⁺ atoms in polytungstates. In this way, the special electron interaction between ceria and tungsten oxides causes an electron transfer from Ceⁿ⁺ to Wⁿ⁺, stabilizing the ceria in a relatively high valence state, which leads to a decrease in Ce³⁺ and oxygen vacancies. The decreased amount of oxygen vacancies results in the inhibition of oxygen transfer from the bulk CeO₂ phase to the surface, and this thereby would decrease the reducibility of the catalysts.

3.4 Catalytic behavior

Overall, the N₂ selectivity and NH₃-SCR activity of catalysts are controlled by the redox properties and acidity of catalysts. The reduced redox properties of ceria due to tungsten oxide modification leads to smaller production of NO₂ and NO in the NH₃ oxidation experiment and NO₂ in the NO oxidation experiment. The enhanced Brønsted acidity of ceria caused by tungsten modification results in stronger ammonia adsorption on acid sites, and the adsorbed ammonia is more stable at high temperatures. Therefore, the activity and selectivity of WO_x-CeO₂ cat-

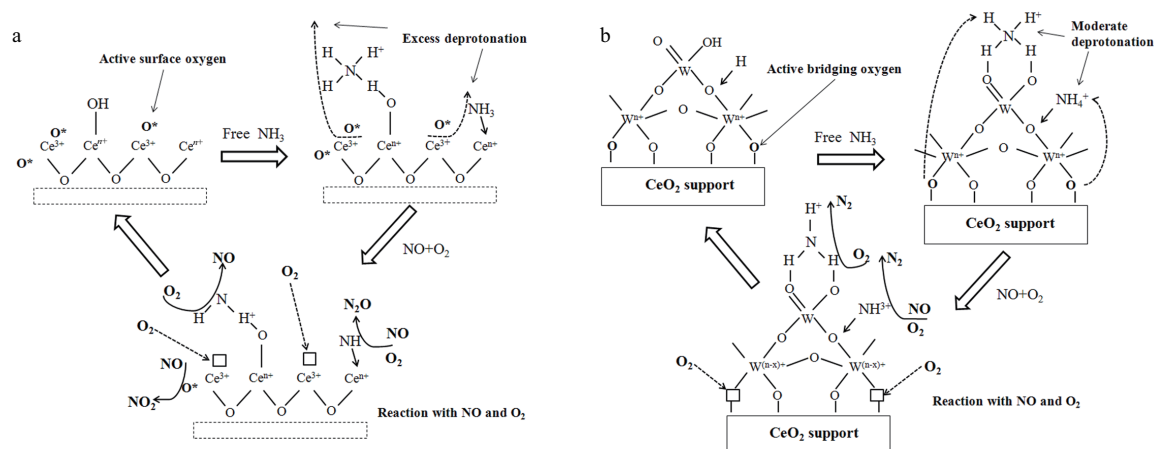


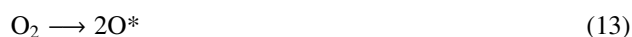
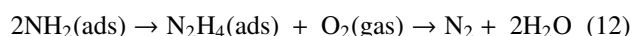
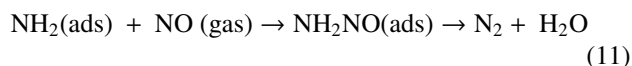
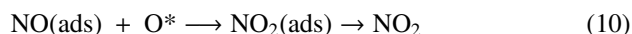
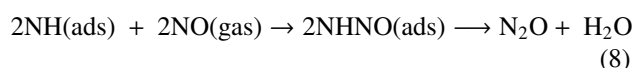
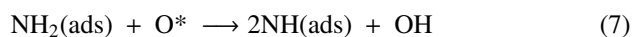
Fig. 12 Tentative models for the NH_3 -SCR reaction route on CeO_2 (a) and 10WC (b) catalysts.

alysts for the NH_3 -SCR reaction increase with the loading amount of WO_x . NH_3 adsorbed on acid sites tends to form amide species as a result of the de-protonation of NH_3 and NH_4^+ . On WO_x/CeO_2 catalysts, the NH_3 -SCR reaction proceeds between adsorbed amide species and gaseous nitric oxide, in which NH_2 and NH_3^+ act as important intermediates in SCR and ammonia oxidation reactions (Busca et al., 1998; Liu et al., 2006; Ramis et al., 1990; Dumesic et al., 1996). Moderate oxidation of ammonia to NH_2 species is an important step in the activation reaction, and then gaseous NO reacts with activated NH_2 to give nitrogen and water. Excessive oxidation of ammonia may generate NH or even N and subsequently N_2O and NO , leading to decreases in both SCR activity and N_2 selectivity.

The reduction of Ce^{4+} to Ce^{3+} can release active oxygen, facilitating the oxidative dehydrogenation of ammonia. The special electron interaction between ceria and tungsten oxides inhibited the reduction of Ce^{4+} . This effect is beneficial to form a WO_x/CeO_2 structure bearing moderate oxidation ability. Ceria exhibits the worst activity and N_2 selectivity, due to its strongest abilities for ammonia and NO oxidation. For WO_x/CeO_2 catalysts, the continuous decrease of reducibility caused by WO_x modification restrains both NO_x formation and ammonia oxidation. A similar mechanism exists in NO oxidation over all the tungsten-containing catalysts.

On the basis of the literature and our experiments, two tentative reaction routes for the SCR reaction over CeO_2 and WO_x/CeO_2 catalysts are presented in Fig. 12. For ceria, a large amount of ammonia is mainly adsorbed on Lewis acid sites. The active surface oxygen accelerates the successive H-abstraction process of NH_3 to form NH_2 and NH species. NH then reacts with NO and O_2 to form N_2O and the further oxidation of NO generates NO_2 (Reactions (6)–(10)). The consumed surface oxygen can be rapidly supplemented by gaseous oxygen. For WO_x -containing catalysts, the increased amount of acid sites due to tungsten modification promotes NH_3 adsorption. Meanwhile, the decreased amount of surface oxygen prevents the excessive de-protonation of NH_3 species to NH . More NH_2 species form from NH_3 deprotonated by bridging oxygen from W-O-Ce and W-O-W domains. NH_2 reacts with NO and O_2

to generate water and nitrogen. The decrease of surface oxygen inhibits the generation of NO_2 from NO . The oxygen vacancies left behind by the reaction of bridging oxygen are further oxidized by gaseous oxygen (Reaction (13)).



“ O^* ” represents surface active oxygen or bridging oxygen, “ads” represents adsorption species, and “gas” represents the gas phase species.

4 Conclusions

Both the acidity and redox properties of WO_x/CeO_2 catalysts are determined by the structure of WO_x species and the strong interaction between WO_x and ceria, and thereby the activity and selectivity of WO_x/CeO_2 catalysts is explained.

(1) $10\text{WO}_x/\text{CeO}_2$ calcined at 600°C shows the best SCR activity with a wide working temperature interval (NO_x conversion $> 80\%$) of $200\text{--}450^\circ\text{C}$.

(2) For all WO_x/CeO_2 samples, isolated surface polytungstate is the main dominant species on ceria, and no crystalline WO_3 nanoparticles were observed. With increasing tungsten oxide loading, the electron interaction of WO_x and CeO_2 increases, and the size of polytungstate domains grows.

(3) The modification by tungsten oxides also enhances the strength of the acid sites, which promotes the adsorption of ammonia and enhances the stability of adsorbed ammonia. The activity of surface oxygen is inhibited

with WO_x loading, which is mainly caused by the strong interaction between WO_x and CeO₂, leading to the lower oxidation ability of catalysts to inhibit NH₃ and NO oxidation, which favors the SCR reaction. These effects result in high NH₃-SCR activity and high selectivity for nitrogen with the tungsten containing catalysts.

Acknowledgments

The authors would like to acknowledge the Ministry of Science and Technology, PR China for financial support of Project 2010CB732304 and Science and Technology Department of Zhejiang Province Project 2011C31010. Moreover, we would also thank the State Key Lab of New Ceramics and Fine Process in Tsinghua University.

References

- Adamowska M, Krzto A, Najbar M, Costa P D, Djéga-Mariadassou G, 2008. DRIFT study of the interaction of NO and O₂ with the surface of Ce_{0.62}Zr_{0.38}O₂ as deNO_x catalyst. *Catalysis Today*, 137(2-4): 288–291.
- Barton D G, Shtein M, Wilson R D, Soled S L, Iglesia E, 1999a. Structure and electronic properties of solid acids based on tungsten oxide nanostructures. *Journal of Physical Chemistry B*, 103(4): 630–640.
- Barton D G, Soled S L, Meitzner G D, Fuentes G A, Iglesia E, 1999b. Structural and catalytic characterization of solid acids based on Zirconia modified by tungsten oxide. *Journal of Catalysis*, 181(1): 57–72.
- Bigey C, Hilaire L, Maire G, 2001. WO₃CeO₂ and Pd/WO₃CeO₂ as potential catalysts for reforming applications: I. physicochemical characterization study. *Journal of Catalysis*, 198(2): 208–222.
- Brandenberger S, Kröcher O, Tissler A, Althoff R, 2008. The state of the art in selective catalytic reduction of NO_x by ammonia using metal-exchanged Zeolite catalysts. *Catalysis Reviews Science and Engineering*, 50(4): 492–531.
- Busca G, Larrubia M A, Arrighi L, Ramis G, 2005. Catalytic abatement of NO_x: Chemical and mechanistic aspects. *Catalysis Today*, 107-108: 139–148.
- Busca G, Liett L, Ramis G, Berti F, 1998. Chemical and mechanistic aspects of the selective catalytic reduction of NO_x by ammonia over oxide catalysts: A review. *Applied Catalysis B: Environmental*, 18(1-2): 1–36.
- Casapu M, Kröcher O, Elsener M, 2009. Screening of doped MNO_x-CeO₂ catalysts for low-temperature NO-SCR. *Applied Catalysis B: Environmental*, 88(3-4): 413–419.
- Chen L, Li J H, Ge M F, Zhu R H, 2010. Enhanced activity of tungsten modified CeO₂/TiO₂ for selective catalytic reduction of NO_x with ammonia. *Catalysis Today*, 153(3-4): 77–83.
- Cormal A, Atienzar P, García H, Chane-Ching J, 2004. Hierarchically mesostructured doped CeO₂ with potential for solar-cell use. *Nature Material*, 3(6): 394–397.
- Cortés-Jacome M A, Angeles-Chavez C, López-Salinas E, Navarrete J, Toribio P, Toledo J A, 2007. Migration and oxidation of tungsten species at the origin of acidity and catalytic activity on WO₃-ZrO₂ catalysts. *Applied Catalysis A: General*, 318: 178–189.
- Dumesic J A, Topsøe N Y, Topsøe H, Chen Y, Slabick T, 1996. Kinetics of selective catalytic reduction of nitric oxide by ammonia over Vanadia/Titania. *Journal of Catalysis*, 163(2): 409–417.
- Gao X, Du X S, Cui L W, Fu Y C, Luo Z Y, Cen K F, 2010. A Ce-Cu-Ti oxide catalyst for the selective catalytic reduction of NO with NH₃. *Catalysis Communications*, 12(4): 255–258.
- Gao X, Jiang Y, Zhong Y, Luo Z Y, Cen K F, 2010. The activity and characterization of CeO₂-TiO₂ catalysts prepared by the sol-gel method for selective catalytic reduction of NO with NH₃. *Journal of Hazardous Materials*, 174(1-3): 734–739.
- Gregorio F D, Keller V, 2004. Activation and isomerization of hydrocarbons over WO₃/ZrO₂ catalysts I. Preparation, characterization, and X-ray photoelectron spectroscopy studies. *Journal of Catalysis*, 225(1): 45–55.
- Grossale A, Nova I, Tronconi E, 2008. Study of a Zeolite-based system as NH₃-SCR catalyst for diesel exhaust after treatment. *Catalysis Today*, 136(1-2): 18–27.
- Guisnet M, Magnoux P, 1997. Deactivation by coking of zeolite catalysts. Prevention of deactivation. Optimal conditions for regeneration. *Catalysis Today*, 36(4): 47–483.
- Huang S J, Chen F C, Liu S L, Zhu Q J, Zhu X X, Xin W J et al., 2007. The influence of preparation procedures and tungsten loading on the metathesis activity of ethene and 2-butene over supported WO₃ catalysts. *Journal of Molecular Catalysis A: Chemical*, 267(1-2): 224–233.
- Huang Y, Tong Z Q, Wu B, Zhang J F, 2008. Low temperature selective catalytic reduction of NO by ammonia over V₂O₅-CeO₂/TiO₂. *Journal of Fuel Chemistry and Technology*, 36(5): 616–620.
- Kim T, Burrows A, Kiely C J, Wachs I E, 2007. Molecular/electronic structure-surface acidity relationships of model-supported tungsten oxide catalysts. *Journal of Catalysis*, 246(2): 370–381.
- Klukowski D, Balle P, Geiger B, Wagloehner S, Kureti S, Kimerle B et al., 2009. On the mechanism of the SCR reaction on Fe/HBEA zeolite. *Applied Catalysis B: Environmental*, 93(1-2): 185–193.
- Li Y, Cheng H, Li D Y, Qin Y S, Xie Y M, Wang S D, 2008. WO₃/CeO₂-ZrO₂, a promising catalyst for selective catalytic reduction (SCR) of NO_x with NH₃ in diesel exhaust. *Chemical Communications*, (12): 1470–1472.
- Lin Q, Shimizu K, Satsuma A, 2009. Redox property of tungstated-zirconia analyzed by time resolved in situ UV-Vis spectroscopy. *Applied Catalysis A: General*, 365(1): 55–61.
- Liu H D, Wei L Q, Yue R L, Chen Y F, 2010. CrO_x-CeO₂ binary oxide as a superior catalyst for NO reduction with NH₃ at low temperature in presence of CO. *Catalysis Communications*, 11(9): 829–833.
- Liu Q Y, Liu Z Y, Li C Y, 2006. Adsorption and activation of NH₃ during selective catalytic reduction of NO by NH₃. *Chinese Journal of Catalysis*, 27(7): 636–646.
- Mamede A S, Payen E, Grange P, Poncelet G, Ion A, Alifanti M et al., 2004. Characterization of WO_x/CeO₂ catalysts and their reactivity in the isomerization of hexane. *Journal of Catalysis*, 223(1): 1–12.
- Martin C, Malet P, Solana G, Rives V, 1998. Structural analysis of silica-supported tungstates. *Journal of Physical Chemistry B*, 102(15): 2759–2768.
- Mineshige A, Taji T, Muroi Y, Kobune M, Fujii S, Nishi N et al., 2000. Oxygen chemical potential variation in ceria-based solid oxide fuel cells determined by Raman spectroscopy. *Solid State Ionics*, 135(1-4): 481–485.
- Onfroy T, Clet G, Houalla M, 2005. Acidity, Surface Structure, and Catalytic Performance of WO_x Supported on Monoclinic Zirconia. *Journal of Physical Chemistry B*, 109(8): 3345–3354.

- Qi G, Yang R T, Chang R, 2004. $\text{MNO}_x\text{-CeO}_2$ mixed oxides prepared by co-precipitation for selective catalytic reduction of NO with NH_3 at low temperatures. *Applied Catalysis B: Environmental* 51(2): 93–106.
- Ramis G, Busca G, Bregani F, Forzatti P, 1990. Fourier transform-infrared study of the adsorption and coadsorption of nitric oxide, nitrogen dioxide and ammonia on vanadia-titania and mechanism of selective catalytic reduction. *Applied Catalysis*, 64: 259–278.
- Ramis G, Yi L, Busca G, Turco M, Kotur E, Willey R J, 1995. Absorption, activation, and oxidation of ammonia over SCR catalysts. *Journal of Catalysis*, 157(2): 523–535.
- Reddy B M, Khan A, Yamada Y, Kobayashi T, Loridant S, Volta J, 2002. Surface characterization of $\text{CeO}_2/\text{SiO}_2$ and $\text{V}_2\text{O}_5/\text{CeO}_2/\text{SiO}_2$ catalysts by Raman, XPS, and other techniques. *Journal of Physical Chemistry B*, 106(42): 10964–10972.
- Reddy B M, Khan A, Yamada Y, Kobayashi T, Loridant S, Volta J, 2003. Structural characterization of CeO_2TiO_2 and $\text{V}_2\text{O}_5/\text{CeO}_2\text{TiO}_2$ catalysts by Raman and XPS techniques. *Journal of Physical Chemistry B*, 107(22): 5162–5176.
- Santesteban J G, Calabro D C, Borghard W S, Chang C D, Vartuli J C, Tsao Y P et al., 1999. H-spillover and SMSI effects in paraffin hydroisomerization over $\text{Pt}/\text{WO}_x/\text{ZrO}_2$ Bifunctional Catalysts. *Journal of Catalysis*, 183(2): 314–322.
- Schulz H, Stark W J, Maciejewski M, Pratsinis S E, Baiker A, 2003. Flame-made nanocrystalline ceria/zirconia doped with alumina or silica: structural properties and enhanced oxygen exchange capacity. *Journal of Materials Chemistry*, 13(12): 2979–2984.
- Schwidder M, Heikens S, Toni A D, Geisler S, Berndt M, Brückner A et al., 2008. The role of NO_2 in the selective catalytic reduction of nitrogen oxides over Fe-ZSM-5 catalysts: Active sites for the conversion of NO and of NO/NO_2 mixtures. *Journal of Catalysis*, 259(1): 96–103.
- Shan W P, Liu F D, He H, Shi X Y, Zhang C B, 2011. Novel cerium-tungsten mixed oxide catalyst for the selective catalytic reduction of NO_x with NH_3 . *Chemical Communications*, 47(28): 8046–8048.
- Si Z C, Weng D, Wu X D, Jiang Y, 2010. Roles of Lewis and Brønsted acid sites in NO reduction with ammonia on $\text{CeO}_2\text{-ZrO}_2\text{-NiO-SO}_4^{2-}$ catalyst. *Journal of Rare Earths*, 28(5): 727–731.
- Singh G P, Singh D P, 2011. Effect of WO_3 on structural and optical properties of $\text{CeO}_2\text{-PbO-B}_2\text{O}_3$ glasses. *Physica B: Condensed Matter*, 406(3): 640–644.
- Sjövall H, Blint R J, Olsson L, 2009. Detailed kinetic modeling of NH_3 SCR over Cu-ZSM-5. *Applied Catalysis B: Environmental*, 92(1-2): 138–153.
- Tsunekawa S, Fukuda T, Kasuya A, 2000. Blue shift in ultraviolet absorption spectra of monodisperse CeO_{2x} nanoparticles. *Journal of Applied Physics*, 87(3): 1318–1321.
- Vargas M A L, Casanova M, Trovarelli A, Busca G, 2007. An IR study of thermally stable $\text{V}_2\text{O}_5\text{-WO}_3\text{-TiO}_2$ SCR catalysts modified with silica and rare-earths (Ce, Tb, Er). *Applied Catalysis B: Environmental*, 75(3-4): 303–311.
- Vidal H, Kašpar J, Pijolat M, Colon G, Bernal S, Cordon A et al., 2000. Redox behavior of CeO_2ZrO_2 mixed oxides I. Influence of redox treatments on high surface area catalysts. *Applied Catalysis B: Environmental*, 27(1): 49–63.
- Wachs I E, Kim T, Ross E I, 2006. Catalysis science of the solid acidity of model supported tungsten oxide catalysts. *Catalysis Today*, 116(2): 162–168.
- Wang X H, Lu G Z, Guo Y, Xue Y Y, Jiang L Z, Guo Y L et al., 2007. Structure, thermal-stability and reducibility of Si-doped Ce-Zr-O solid solution. *Catalysis Today*, 126(3-4): 412–419.
- Xu W D, He H, Yu Y B, 2009. Deactivation of a Ce/TiO_2 catalyst by SO_2 in the selective catalytic reduction of NO by NH_3 . *Journal of Physical Chemistry C*, 113(11): 4426–4432.
- Yagoubi B, Hogarth C A, 1993. Effect of cerium oxide addition on the optical absorption edge of tungsten trioxide. *Journal of Materials Science*, 28(1): 239–242.
- Zhu J, Gao F, Dong L H, Yu W J, Qi L, Wang Z et al., 2010. Studies on surface structure of $\text{M}_x\text{O}_y/\text{MoO}_3/\text{CeO}_2$ system ($\text{M}=\text{Ni, Cu, Fe}$) and its influence on SCR of NO by NH_3 . *Applied Catalysis B: Environmental*, 95(1-2): 144–152.

JOURNAL OF ENVIRONMENTAL SCIENCES

Editors-in-chief

Hongxiao Tang

Associate Editors-in-chief

Nigel Bell Jiuhui Qu Shu Tao Po-Keung Wong Yahui Zhuang

Editorial board

R. M. Atlas University of Louisville USA	Alan Baker The University of Melbourne Australia	Nigel Bell Imperial College London United Kingdom	Tongbin Chen Chinese Academy of Sciences China
Maohong Fan University of Wyoming Wyoming, USA	Jingyun Fang Peking University China	Lam Kin-Che The Chinese University of Hong Kong, China	Pinjing He Tongji University China
Chihpin Huang "National" Chiao Tung University Taiwan, China	Jan Japenga Alterra Green World Research The Netherlands	David Jenkins University of California Berkeley USA	Guibin Jiang Chinese Academy of Sciences China
K. W. Kim Gwangju Institute of Science and Technology, Korea	Clark C. K. Liu University of Hawaii USA	Anton Moser Technical University Graz Austria	Alex L. Murray University of York Canada
Yi Qian Tsinghua University China	Jiuhui Qu Chinese Academy of Sciences China	Sheikh Raisuddin Hamdard University India	Ian Singleton University of Newcastle upon Tyne United Kingdom
Hongxiao Tang Chinese Academy of Sciences China	Shu Tao Peking University China	Yasutake Teraoka Kyushu University Japan	Chunxia Wang Chinese Academy of Sciences China
Rusong Wang Chinese Academy of Sciences China	Xuejun Wang Peking University China	Brian A. Whitton University of Durham United Kingdom	Po-Keung Wong The Chinese University of Hong Kong, China
Min Yang Chinese Academy of Sciences China	Zhifeng Yang Beijing Normal University China	Hanqing Yu University of Science and Technology of China	Zhongtang Yu Ohio State University USA
Yongping Zeng Chinese Academy of Sciences China	Qixing Zhou Chinese Academy of Sciences China	Lizhong Zhu Zhejiang University China	Yahui Zhuang Chinese Academy of Sciences China

Editorial office

Qingcai Feng (Executive Editor) Zixuan Wang (Editor) Suqin Liu (Editor) Zhengang Mao (Editor)
Christine J Watts (English Editor)

Journal of Environmental Sciences (Established in 1989)

Vol. 24 No. 7 2012

Supervised by	Chinese Academy of Sciences	Published by	Science Press, Beijing, China
Sponsored by	Research Center for Eco-Environmental Sciences, Chinese Academy of Sciences		Elsevier Limited, The Netherlands
Edited by	Editorial Office of Journal of Environmental Sciences (JES) P. O. Box 2871, Beijing 100085, China Tel: 86-10-62920553; http://www.jesc.ac.cn E-mail: jesc@263.net , jesc@rcees.ac.cn	Distributed by	Domestic Science Press, 16 Donghuangchenggen North Street, Beijing 100717, China Local Post Offices through China Foreign Elsevier Limited http://www.elsevier.com/locate/jes
Editor-in-chief	Hongxiao Tang	Printed by	Beijing Beilin Printing House, 100083, China
CN 11-2629/X	Domestic postcode: 2-580	Domestic price per issue	RMB ¥ 110.00

ISSN 1001-0742

