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Toluene decomposition performance and NOx by-product formation during a DBD-catalyst process

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ABSTRACT

Characteristics of toluene decomposition and formation of nitrogen oxide (NOx) by-products were investigated in a dielectric barrier discharge (DBD) reactor with/without catalyst at room temperature and atmospheric pressure. Four kinds of metal oxides, i.e., manganese oxide (MnOx), iron oxide (FeOx), cobalt oxide (CoOx) and copper oxide (CuO), supported on Al₂O₃/nickel foam, were used as catalysts. It was found that introducing catalysts could improve toluene removal efficiency, promote decomposition of by-product ozone and enhance CO₂ selectivity. In addition, NOx was suppressed with the decrease of specific energy density (SED) and the increase of humidity, gas flow rate and toluene concentration, or catalyst introduction. Among the four kinds of catalysts, the CuO catalyst showed the best performance in NOx suppression. The MnOx catalyst exhibited the lowest concentration of O₃ and highest CO₂ selectivity but the highest concentration of NOx. A possible pathway for NOx production in DBD was discussed. The contributions of oxygen active species and hydroxyl radicals are dominant in NOx suppression.

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Introduction

Emission of volatile organic compounds (VOCs) from various industrial processes is one of the most important sources of air pollution. VOCs are detrimental to both human health and the environment, either directly from their toxicity and malodorous nature or indirectly as ozone and smog precursors (Chen et al., 2009). Toluene, as one of the typical VOCs, is usually chosen as a probe contaminant for treatment because it is difficult to deal with due to the presence of a benzene ring and because it causes respiratory disorders and even carcinogenic and mutagenic effects. It also represents one of the most commonly detected VOCs in many industries, such as those producing lacquers, adhesives, and rubber, as well as in some printing and leather tanning processes (Angel et al., 2008)

Dielectric barrier discharge (DBD) has offered an innovative technology for removing toluene from waste gas characterized by low toluene concentration and high flow rate (Atten et al., 1987; Clements et al., 1989; Yamamoto et al., 1996; Snyder and Anderson, 1998; Pietsch, 2001; Kim et al., 2008; Mista and Kacprzyk, 2008). DBD, which contains at least one dielectric barrier, distributes microdischarges throughout the discharge area. It initiates chemical reactions by electron impact dissociation of organic molecules and ionization of the carrier gas (Magureanu et al., 2007). DBD produces a plasma region characterized by significant non-thermal properties. Electric energy is primarily used for the production of high energy electrons, leaving the bulk of gas at room temperature. The potential for energy-saving is a main advantage of DBD technology.

However, DBD alone usually cannot convert pollutants to ideal products. The main problem for VOC removal in a DBD system is the formation of toxic byproducts, such as CO, O₃, and NOx (Magureanu et al., 2007; Van Durme et

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al., 2008; Delagrange et al., 2006; Bo et al., 2009; Kalra et al., 2005). These would cause secondary pollution if not subjected to further treatment. Sometimes they are even more harmful than the original VOCs. Notable improvements in inhibition of their formation can be achieved through the plasma-catalysis process, which introduces a catalyst into the discharge zone. Metal oxides have been widely used as active components in the removal of organic compounds (Das and Parida, 2007; Guo et al., 2007; Sano et al., 2006; Karuppiah et al., 2012; Lahousse et al., 1998).

The byproduct NOx can affect reaction rate and even lead to deactivation of a catalyst. Van Durme reported that the deactivation of catalyst materials may be explained by the formation of HNO₃ in the plasma discharge (Van Durme et al., 2008). Subrahmanyam also found that NOx could form nitrate, which may cause the deactivation of catalysts (Subrahmanyam et al., 2006). In this study, the formation characteristics of NOx during toluene removal were investigated in a DBD-catalyst system at room temperature and atmospheric pressure. Four kinds of metal oxide catalysts, that is, manganese oxide (MnOx), iron oxide (FeOx), cobalt oxide (CoOx) and copper oxide (CuO), supported on Al_2O_3 /nickel foams, were placed in the discharge area of the reactor. Experimental tests were conducted to determine the effect of the catalysts on toluene oxidation. In addition, exhaust gas was experimentally simulated to determine the influences of humidity, gas flow rate and toluene concentration as well as catalysts on NOx formation.

1. Experimental

1.1. Experimental set-up

A schematic diagram of the experimental system is shown in Fig. 1. The apparatus mainly included a gas supply and regulation system, a DBD reactor with a power supply system and a gas analysis system. The initial toluene concentration ranged from 200 to 800 mg/m³. Gaseous toluene was obtained by controlling the air flow rate from a gas cylinder through pure toluene liquid (>99.5%) which was kept in a water bath ($T = 25 \pm 1^{\circ}$ C). After passing through a mixing chamber, the feed gas entered the DBD reactor. Gas flow rate was controlled to between 150 and 450 mL/min.

Toluene and other volatile organic compounds were analyzed by a gas chromatograph (GC-7890II, Tianmei, China) equipped with a hydrogen flame ionization detector (FID) and a DB-5MS capillary column (30 m \times 0.25 mm \times 0.25 mm) heated at 80°C. Humidity was measured by a humidity meter (Center 310, Shuangxu, Shanghai, China). The concentrations of CO and CO_2 were measured by a CO analyzer (TX2000, Oldham, France) and a CO_2 detector (GXH-3010E, Huayun, Beijing, China), respectively. The concentration of ozone was monitored by an ozone analyzer (DCS-1, Lida, Shanghai, China). NOx was detected with a NO analyzer (PGM-1140, Rae, USA) and a NO_2 analyzer (PGM-1150, Rae, USA).

1.2. DBD reactor and power supply system

A wire-plate DBD reactor was used in the reaction (Fig. 2). Two epoxy resin boards (200 mm × 45 mm × 0.8 mm) were used to form reactor walls and acted as dielectric barriers (the dielectric constant ε = 3.6). The high voltage electrode was made of brass wire (diameter 0.6 mm). The brass wire electrode was shaped into a spiral in order to increase energy density in the reaction volume. The wire-to-wire distance was 8 mm. Two grounded copper nets were fixed on the two sides of the middle epoxy resin board. The total length of the reactor was 200 mm, while the effective length was 150 mm. When no catalyst was used, the gap between the high-voltage electrode and the grounded electrode was 8 mm, resulting in a cross sectional area of 400 mm² and a reaction volume of 60 cm³ for the flow channel. To add an in-situ catalyst, it was supported by two pieces of nickel foam (150 mm \times 25 mm \times 2 mm) fixed on the grounded electrodes. This produced a gap of 6 mm, a cross-sectional area for the flow channel of 300 mm², a reaction volume of 45 cm³ and a catalyst volume of 15,000 mm³.

High voltage power was supplied by a booster (0–250 V) combined with a high voltage AC transformer (CTP-2000K, Suman, Nanjing, China) in series. In this experiment the frequency was controlled at 9.512 KHz. The applied voltage and current were measured by a high voltage probe (P6015, Tektronix, USA) and a digitizing oscilloscope (TDS1002, Tektronix, USA). The input power can be detected by the high voltage AC transformer directly. The discharge power (the power deposited to the reactor) was calculated from the applied voltage and the reactor current.

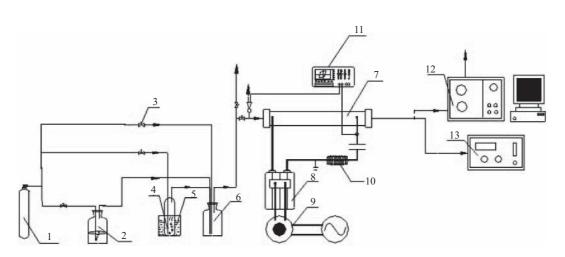


Fig. 1 – Schematic diagram of the experiment. (1) dry air cylinder, (2) bubbling for water vapor preparation, (3) mass flow controller, (4) bubbling for toluene preparation, (5) water bath for toluene generation, (6) buffer, (7) DBD reactor, (8) ac transformer, (9) booster, (10) resistance, (11) oscilloscope, (12) gas chromatogram, (13) ozone analyzer.

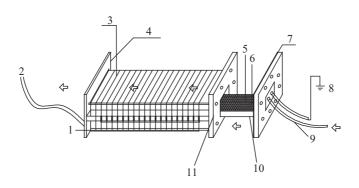


Fig. 2 – Structure of the DBD reactor with catalyst. (1) high voltage electrode, (2) polyethylene tube (gas out), (3) brass wire, (4) epoxy resin board (1 mm), (5) nickel foam, (6) copper net, (7) silicone pad, (8) ground electrode, (9) polyethylene tube (gas in), (10) epoxy resin board (2 mm), (11) drilled hole for a screw.

1.3. Catalyst

Four kinds of catalysts, i.e. MnOx, FeOx, CoOx and CuO supported on Al_2O_3 /nickel foam, were prepared by an impregnation method. The detailed process can be obtained in a previous article (Guo et al., 2006).

Toluene removal efficiency (η), specific energy density (SED, J/L), carbon dioxide selectivity S_{CO2} and energy efficiency (η_{E} , g/kWh) were calculated as follows:

$$\eta = \frac{[toluene]_{in} - [toluene]_{out}}{[toluene]_{in}} \times 100$$
(1)

$$SED = \frac{\text{discharge power (w)}}{\text{gas flow rate (L/min)}} \times 60$$
(2)

$$S_{CO_2}(\%) = \frac{[CO_2]}{[CO_2] + [CO]}$$
(3)

$$\eta_{E} = \frac{[toluene]_{removal}(g/m^{3}) \times gas \ flow \ rate \ \left(m^{3}/hr\right)}{input \ power \ (kW)}. \tag{4}$$

2. Results and discussion

2.1. Toluene removal performance with/without catalyst

2.1.1. Energy yield

The relationship between SED (J/L) and η_E (g/kWh) with/without catalyst presented in Fig. 3a indicates that SED is more dependent on E_Y in a DBD system than in a DBD-catalyst system. For example, the E_Y value was 0.21–1.12 g/kWh in the DBD system compared to 1.24–1.37 g/kWh in the DBD-FeOx catalyst system (SED: 140–320 J/L).

2.1.2. Toluene removal efficiency

The catalytic properties of various metal oxides for toluene oxidation were studied (Fig. 3b). It was found that catalysts promoted toluene decomposition significantly. The following toluene removal efficiency order was found: FeOx catalyst > MnOx catalyst > CuO catalyst > CoOx catalyst > no catalyst. The conversion was only 7.1% without catalyst compared with 45% with the FeOx catalyst (SED: 140 J/L).

2.1.3. Ozone formation

The production of ozone with/without catalyst is indicated in Fig. 3c. It can be seen that metal oxides promoted ozone decomposition. The concentration of O_3 was 34.6 mg/m³ with the MnOx catalyst, compared to 420 mg/m³ without catalyst at 320 J/L. By comparing the results presented in Fig. 3c, it can be concluded that the ozone concentration varied in the following order: MnOx catalyst < FeOx catalyst < CoOx catalyst < CuO catalyst < no catalyst. MnOx catalyst exhibited the best performance in terms of O_3 decomposition.

Plasma generates intermediate species having a sufficiently long lifetime to trigger surface reactions on a catalyst placed in the plasma reactor (Hammer et al., 2004). It is clear that most ozone is decomposed catalytically, forming molecular and highly active atomic oxygen (Guo et al., 2007; Magureanu et al., 2005; Delagrange et al., 2006). Delagrange reported a mechanism for O_3 decomposition with a MnO₂ catalyst as follows (Delagrange et al., 2006):

$$\begin{split} & O_3 \stackrel{\text{MINO}_2}{\longrightarrow} O(^3P) + {}^1\Delta O_2 \end{split} \tag{5} \\ & O_3 \stackrel{\text{MINO}_2}{\longrightarrow} O(^1D) + {}^3\sum O_2. \end{split}$$

2.1.4. Selectivity of CO₂

The selectivity of CO_2 decreased with increasing SED (Fig. 3d). The selectivity of CO_2 decreased from 80.56% to 48.84% when SED increased from 140 to 320 J/L without catalyst. The result was the same as in previous reports (Guo et al., 2006, 2010). But this trend was different than that reported in some papers (Magureanu et al., 2007; Karuppiah et al., 2012). Maybe differences in catalyst oxidation ability cause this discrepancy.

Fig. 3d also shows that CO_2 selectivity can be greatly enhanced by introducing catalysts. The CO_2 selectivity without catalyst was only 48.84% but it increased to 74.62% with the MnOx catalyst (SED: 320 J/L). Oxygen active species are formed in the decomposition of O_3 on a catalyst surface. Carbon monoxide and toluene (or intermediate reaction products) are oxidized by oxygen active species to CO_2 , which eventually lead to a higher CO_2 selectivity (Chavadej et al., 2007).

The MnOx catalyst exhibited the highest CO_2 selectivity (Fig. 3d) and lowest ozone concentration (Fig. 3c), which

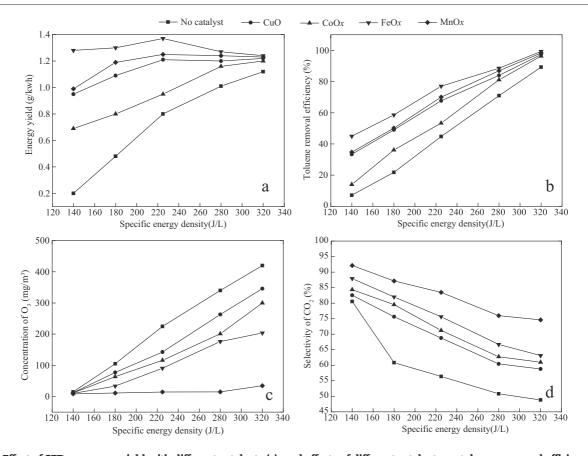


Fig. 3 – Effect of SED on energy yield with different catalysts (a), and effects of different catalysts on toluene removal efficiency (b), O₃ concentration (c), CO₂ selectivity (d). Gas flow rate: 450 m/min, initial concentration of toluene: 400 mg/m³, relative humidity (RH): 20%.

(2007).

proved that active species produced from O_3 decomposition oxidized CO into CO_2 . This is in accord with Sano's result, which concluded that a catalyst with a higher O_3 decomposition activity was related to a catalyst with a higher CO oxidation activity (Sano et al., 2006).

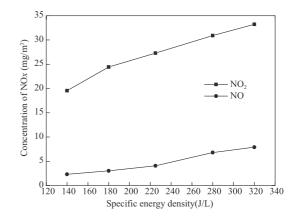
2.2. NOx performance under different conditions

2.2.1. Specific energy density effect

The production of the main NOx species (NO, NO₂) as a function of SED is shown in Fig. 4. The concentration of NO and NO₂ both increased with increasing SED. For example, the concentration of NO increased from 2.3 to 7.9 mg/m³, and for NO₂ it was 19.56– 33.23 mg/m³ with SED ranging from 140 to 320 J/L. This trend is consistent with the result by Cooray and Rahman (2005). The concentration of NO was much lower than that of NO₂. The concentration of NO₂ was 33.2 mg/m³ but for NO it was only 7.9 mg/m³ (SED: 320 J/L). Some active molecules and radicals such as O₃, OH and O produced in the discharge are effective in the oxidization of NO to NO₂, which results in less NO in the effluent.

2.2.2. Humidity effect

Since practical oxidation of VOCs in air usually occurs in the presence of water vapor, the effect of humidity on the formation of NOx was investigated as shown in Fig. 5a. The results suggested that the addition of water vapor caused a remarkable



suppression of NOx. Close to 65.0 mg/m³ of NO₂ was detected

under dry conditions at SED of 320 J/L, whereas, it decreased to

30.1 mg/m³ (relative humidity/RH: 20%) and 28.0 mg/m³ (RH:

40%). This is in agreement with the report of Van Durme et al.

Fig. 4 – Effect of SED (J/L) on NOx formation. Gas flow rate: 450 mL/min, initial concentration of toluene: 400 mg/m³, RH: 20%.

2.2.3. Effect of gas flow rate

Gas flow rate had a significant effect on NOx generation (Fig. 5b). When flow increased from 150 to 450 mL/min, the concentration of NO_2 decreased from 51.7 down to 33.23 mg/m³ and NO decreased from 15.8 to 7.9 mg/m³ (SED: 320 J/L). The increase of gas flow rate decreases the concentration of discharging electrons and the probability of collision rate with electrons. Therefore fewer N and O atoms are generated and NOx is depressed.

2.2.4. Effect of initial toluene concentration

Fig. 5c shows the concentration of NOx under different initial toluene concentrations. A better performance in suppressing NOx was identified when the toluene concentration was higher. NO₂ concentration decreased from 48.88 to 19.39 mg/m³ with the increase of toluene concentration from 200 to 800 mg/m³ (SED: 320 J/L).

More chemical species compete for radicals generated by a fixed amount of plasma energy under a higher toluene concentration. The detailed reactions were discussed by Guo et al. (2006). Moreover, detection of polymers containing N in the effluent supports the idea that N_2 can be broken, and radicals that are dissociated from toluene would compete for energy with N_2 (Subrahmanyam et al., 2006; Bo et al., 2007).

2.2.5. Effect of different catalysts

Catalysts had a positive effect on the suppression of NOx (Fig. 5d). The NO₂ concentration was 33.23 mg/m³ without catalyst, but it decreased to 16.6 mg/m³ with the CuO catalyst (SED: 320 J/L). And, the NO concentration without catalyst was nearly five times higher than that with the CuO catalyst. The concentration of NOx follows the trend: CuO catalyst < CoOx catalyst < FeOx catalyst < MnOx catalyst < no catalyst. The CuO catalyst shows excellent performance in the suppression of NOx. However, from the results of Fig. 3b, c, d, the CuO catalyst showed a weaker catalytic ability, for example lower toluene decomposition, the highest O₃ concentration and lowest CO₂ selectivity. But actually when we choose a catalyst, catalytic ability is the most important factor. This result in Fig. 5d could give some suggestions for NOx suppression.

The MnOx catalyst had the best ability in decomposing O_3 , but the concentration of NOx was the highest among the four kinds of catalysts. According to formula (5), with the MnOx catalyst most O_3 is decomposed to O_2 and O atoms, which may combine with N atoms and generate more NOx compared with other catalysts. On the other hand, these O atoms also are consumed in the decomposition of toluene. With catalyst, toluene removal efficiency is higher than without catalyst. The

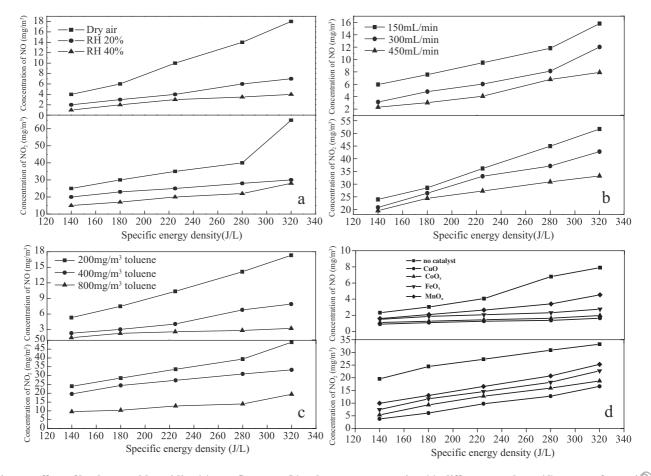


Fig. 5 – Effect of background humidity (a), gas flow rate (b), toluene concentration (c), different catalysts (d) on NOx formation. Gas flow rate: 450 mL/min, initial concentration of toluene: 400 mg/m³ when they are used in experiments.

competing reaction of O atoms with toluene causes a decrease in the NOx concentration.

Metal oxides have been reported to be highly effective for NOx removal in combination with non-thermal plasma (Rappé et al., 2004; Sun et al., 2005). This can be explained as follows. First, microdischarges are generated inside the catalyst pores, resulting in more discharge per volume and energy density of the discharge (Hensel et al., 2005; Holzer et al., 2005). Catalysts may induce a shift in the distribution of accelerated electrons (Van Durme et al., 2008). Then, partially oxidized hydrocarbons and peroxy radicals (RO₂) are generated, which react with NO and strongly influence NO₂ formation rates (Rappé et al., 2004). In addition, NO₂ reacts over the catalyst while partially oxidized hydrocarbons are consumed during selective catalytic reduction, producing CO₂, N₂ and H₂O. On the other side, adsorption is another reason for the lower amount of NOx detected with catalysts (Kwak et al., 2006).

2.3. Possible formation pathways of NOx

2.3.1. Nitrogen atom reaction pathway

Since in this study no nitrogen was contained in the pollutant, it could be assumed that NOx came from the air that was used as carrier gas. Electrons, ions, excited molecules and free radicals are formed during the DBD process. When fast electrons are absorbed in the carrier gas, they cause ionization and excitation of N₂, O_2 and H₂O in the carrier gas. Primary species and secondary electrons are formed. The G-values (molecules/100 eV) of the primary species are simplified as follows (Mätzing, 1989).

$$4.43N_2 \rightarrow 0.29N_2^* + 0.855N(^2D) + 0.295N(^2P) + 1.87N + 2.27N_2^+ + 0.69N^+ + 2.96e$$
(6)

 $5.3770_2 \rightarrow 0.0770_2^* + 2.250 {\binom{1}{D}} + 2.80 + 0.18 {\binom{0^*}{2}} + 2.070_2^+ + 1.230^+ + 3.3e \eqref{7}$

 $\begin{array}{l} 7.33H_20 {\rightarrow} 0.51H_2 + 0.460 {\binom{3}{P}} + 4.250H + 4.15H + 1.99H_20^+ + 0.01H_2^+ \\ \\ + 0.570H^+ + 0.67H^+ + 0.060^+ + 3.3e \end{array} \tag{8}$

 $N(^{2}D)$ and $N(^{2}P)$ as the first two electronically excited states of atomic nitrogen react with O₂ and H₂O, then NO is formed (Herron, 1999).

$$N(^{2}D) + O_{2} \rightarrow NO + O(^{3}P, ^{1}D)$$
 $k = 5.2 \times 10^{-12} \text{ cm}^{3}/(\text{mol} \cdot \text{sec})(298 \text{ K})$

(9)

$$N(^{2}D) + H_{2}O \rightarrow H_{2} + NO$$
 $k = 4.2 \times 10^{-11} \text{cm}^{3}/(\text{mol} \cdot \text{sec})(298 \text{K})$ (10)

$$N(^{2}P) + O_{2} \rightarrow O + NO$$
 $k = 2.5 \times 10^{-12} \text{cm}^{3}/(\text{mol} \cdot \text{sec})(298 \text{ K})$ (11)

2.3.2. Oxygen atom reaction pathway

NOx reactions involving O atom are as follows:

$$\mathbf{O} + \mathbf{NO} \rightarrow \mathbf{NO}_2 \quad k = 3.0 \times 10^{-11} \tag{12}$$

 $O + NO_2 \rightarrow NO_3 \quad k = 2.2 \times 10^{-11}$ (13)

$$0 + NO_2 \rightarrow O_2 + NOk = 9.7 \times 10^{-12}$$
 (14)

$$0 + NO_3 \rightarrow O_2 + NO_2 k = 1.7 \times 10^{-11}$$
 (15)

The rate constants are taken from the literature (Atkinson et al., 1992, 1997). The units of the reactions are: \sec^{-1} , $cm^3/(mol\cdotsec)$ and $cm^3/(mol\cdotsec)$ for first-, second- and third-order reactions, respectively. Kim indicated that the increase of oxygen partial pressure can reduce the formation of nitrogen oxide (Kim et al., 2008).

2.3.3. OH radical reaction pathway

Water mainly came from the oxidation of toluene with air and the water contained in the background gas.

$$C_7H_8 + 9O_2 \rightarrow 7CO_2(CO) + 4H_2O$$
 (16)

Fast electrons from the electron beam ionize H_2O and produce OH radical (Chae, 2003).

$$e + H_2 O \rightarrow e + O + O H$$
 (17)

$$\mathbf{e} + \mathbf{O}_2 \rightarrow \mathbf{e} + \mathbf{O}(^3\mathbf{P}) + \mathbf{O}(^1\mathbf{D}) \tag{18}$$

$$O(^{1}D) + H_{2}O \rightarrow OH \cdot + OH \cdot$$
(19)

OH radicals react with NO and NO_2 to form HNO_2 and HNO_3 . The reactions are as follows (Chmielewski et al., 2002):

$$OH \cdot + NO \rightarrow HNO_2 \qquad \qquad k = 3.2 \times 10^{-11} \tag{20}$$

$$DH \cdot + NO_2 \rightarrow HNO_3 \qquad k = 6.0 \times 10^{-11}$$
(21)

$$OH \cdot + NO_3 \rightarrow HO_2 + NO_2$$
 $k = 2.3 \times 10^{-11}$ (22)

$$OH + HNO_2 \rightarrow H_2O + NO_2$$
 $k = 4.9 \times 10^{-12}$ (23)

$$OH \cdot +HNO_3 \rightarrow H_2O + NO_3$$
 $k = 1.5 \times 10^{-13}$. (24)

Reactions (20) and (21) are critical for NOx removal. The results shown in Fig. 5a demonstrate the suppression of NOx with the addition of water.

2.3.4. O_3 reaction pathway

 O_3 , which is formed from the reaction of O_2 and O atoms, is involved in the NOx reduction reaction as follows (Chmielewski et al., 2002):

$$O_3 + NO \rightarrow NO_2 + O_2$$
 $k = 1.8 \times 10^{-14}$ (25)

$$O_3 + NO_2 \rightarrow NO_3 + O_2$$
 $k = 3.2 \times 10^{-17}$. (26)

Comparing the rate constant from reactions (12)-(15), (20)–(26), it can be seen that O atoms and OH radicals play a more important role than that of O_3 in this process.

3. Conclusions

(1) The concentration of NO₂ was much higher than that of NO since NO could be oxidized to NO₂ easily in a DBD reactor. (2)

Catalysts can suppress NOx production due to the fact that catalysts could affect some characteristics of the DBD and adsorb NOx. Long-lived species are useful to further decompose toluene on the surface of the catalyst. Among the four kinds of catalysts, the CuO catalyst showed the best performance in NOx suppression. (3) Oxygen active species and hydroxyl radicals are more important than ozone in suppressing NOx.

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