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Effects of calcination temperature on physicochemical property and activity of CuSO₄/TiO₂ ammonia-selective catalytic reduction catalysts

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ABSTRACT

CuSO₄/TiO₂ catalysts with high catalytic activity and excellent resistant to SO₂ and H₂O₃, were thought to be promising catalysts used in Selective catalytic reduction of nitrogen oxides by NH₃. The performance of catalysts is largely affected by calcination temperature. Here, effects of calcination temperature on physicochemical property and catalytic activity of CuSO₄/TiO₂ catalysts were investigated in depth. Catalyst samples calcined at different temperatures were prepared first and then physicochemical properties of the catalyst were characterized by N2 adsorption-desorption, X-ray diffraction, thermogravimetric analysis, Raman spectra, Fourier-transform infrared spectroscopy, X-ray photoelectron spectroscopy, temperature-programmed desorption of NH3, temperature-programmed reduction of H₂ and in situ diffuse reflectance infrared Fourier transform spectroscopy. Results revealed that high calcination temperature had three main effects on the catalyst. First, sintering and anatase transform into rutile with increase of calcination temperature, causing a decrement of specific surface area. Second, decomposition of CuSO₄ under higher calcination temperature, resulting in disappears of Brønsted acid sites (S-OH), which had an adverse effect on surface acidity. Third, CuO from the decomposition of CuSO4 changed surface reducibility of the catalyst and favored the process of NH3 oxidation to nitrogen oxides (NOx). Thus, catalytic activity of the catalyst calcined under high temperatures (≥600°C) decreased largely.

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Introduction

As an efficient method to abate nitrogen oxides (NO_x) from the combustion of fossil fuels, selective catalytic reduction of NO_x by NH₃ (NH₃-SCR) over catalysts have attracted much attention (Odenbrand, 2018; Chen et al., 2018b). Due to high NOx removal efficiency and excellent SO2 tolerance, vanadiumbased catalysts are the main commercial SCR catalysts used today (He et al., 2018; Chen et al., 2018a). However, some shortcomings of vanadium-based catalysts such as biological toxicity of vanadium and undesirable activity of SO2 oxidation still existed (Wang et al., 2019; Yu et al., 2019b). Thus, novel vanadium-free SCR catalysts are developed by many researchers (Yan et al., 2019; Cheng et al., 2019; Lian et al., 2019; Woo et al., 2019). Metal oxides and zeolite catalysts performed high activity in NH3-SCR reaction, but their low SO2 tolerance has been one of main obstacles for industrial applications (Gao et al., 2017; Hammershøi et al., 2018; Wijayanti et al., 2016).

From the view point of high SO₂ tolerance, metal sulfates were used in NH₃-SCR reaction by some researchers (Ma et al., 2011; Du et al., 2016; Yu et al., 2017). Fe₂(SO₄)₃/TiO₂ catalysts performed high activity with NO_x conversion higher than 90% in temperature range from 350 to 450°C; moreover, they owned a high resistant to water and SO₂ (Ma et al., 2011). Support Fe₂(SO₄)₃ and CuSO₄ on Ce-Ti mixed oxides catalysts displayed higher activity than vanadium-based catalysts in the temperature range from 250 to 350°C, in addition, the support metal sulfate catalysts also showed high resistances to SO₂ (Du et al., 2016). In our previous work, due to abundant acid sites (both Brønsted and Lewis acid sites) on catalyst surface, we found that CuSO₄/TiO₂ catalysts with 10 wt.% CuSO₄ performed high NH₃-SCR activity in temperature range from 280 to 380°C and the catalysts owned excellent tolerance to water and SO₂ (Yu et al., 2017). Metal sulfate catalysts have been thought to be promising SCR catalysts used in industry. However, compared with metal oxides and zeolite catalysts, the researches about metal sulfate catalysts are less and more attention should be paid.

Calcination temperature is an important factor for preparing catalysts. The calcination temperature can affect physicochemical properties of SCR catalysts, such as distribution of active sites, surface acidity and reducibility, thus impact the catalytic performance largely (Wu et al., 2016, 2019). The effects of calcination temperature on physicochemical properties and catalytic activity of CuSO₄/TiO₂ catalysts are still unknown.

In this work, CuSO₄/TiO₂ catalysts were calcined at 500, 600 and 700°C, respectively. The effects of calcination temperature on physicochemical properties of the catalyst were investigated in depth by N₂ adsorption-desorption, X-ray diffraction (XRD), thermogravimetric analysis (TGA), Raman spectra, Fourier-transform infrared spectroscopy (FTIR), X-ray photoelectron spectroscopy (XPS), temperature-programmed desorption of NH₃ (NH₃-TPD), temperature-programmed reduction of H₂ (H₂-TPR) and in situ diffuse reflectance infrared Fourier transform spectroscopy (in situ DRIFTS). The NH₃-SCR activity of all prepared catalysts was also tested.

1. Materials and methods

1.1. Catalysts preparation

 TiO_2 (anatase, Aladdin Industrial Corporation, USA) (6.0 g) was mixed with 1.04 g of $CuSO_4 \cdot 5H_2O$ (analytical reagent, Sinopharm Chemical Reagent Corporation, China) in an agate mortar and grinded to uniformity. Then mixed powder was calcined at $500^{\circ}C$ in air for 5 hr (named CuSTi). In order to determine the effect of calcination temperature on the catalyst, 2.0 g of CuSTi sample was calcined at 600 and $700^{\circ}C$ in air for 3 hr, named CuSTi-600 and CuSTi-700, respectively.

1.2. Activity test

The catalytic activity of the sample was tested on a fixed-bed quartz reactor (diameter 10 mm and height 600 mm). Catalyst sample (1.0 mL, 20–40 mesh) was placed in middle of the reactor. The mixture gas contained 700 ppmV NO, 700 ppmV NH₃, 5.0 vol.% H_2O and 4.0 vol.% O_2 in N_2 . Total gas flow rate was 1000 mL/min, corresponding to gas hourly space velocity (GHSV) = 60,000 hr⁻¹. Concentrations of NO, NO₂, N_2O and NH₃ in the gas were detected by a Fourier transform infrared spectrometer (FTIR) gas analyzer (IGS, Thermo Fisher, USA). NO_x conversion (x) was calculated from Eq. (1):

$$x = \frac{C_{NO_x,in} - C_{NO_x,out}}{C_{NO_v,in}} \times 100\%$$
 (1)

where, $C_{\text{NOx,in}}$ and $C_{\text{NOx,out}}$ mean the concentration of NO_x (NO + NO₂) in the inlet and outlet, respectively.

The oxidation of NH₃ by O_2 over each catalyst sample was also tested (the mixture gas contained 700 ppmV NH₃, 5.0 vol.% H₂O and 4.0% O₂ in N₂) and the conversion of NH₃ to NO_x (y) under NH₃ oxidation conditions was calculated from Eq. (2):

$$y = \frac{C_{\text{NO}_x,\text{out}}}{C_{\text{NH}_3,\text{in}}} \times 100\%$$
 (2)

where $C_{NH3,in}$ means the concentration of NH_3 in the inlet and $C_{NOx,out}$ means the concentration of NO_x ($NO + NO_2$) in the outlet under NH_3 oxidation conditions.

1.3. Catalyst characterizations

 N_2 adsorption-desorption isotherm of the catalyst sample was recorded by a surface area and pore size analyzer (NOVA, 2000e; Quantachrome Company, USA). Powder XRD of each sample was carried out on X'Pert Pro XRD diffractometer (EPSILON5, PANalytical B.V., Netherland), using Cu palladium and $K\alpha$ radiation. The mean crystallite size of anatase ($d_{\rm crys}$) was calculated by Scherrer Eq. (3) based on full width at half maximum (FWHM) of the (101) plane.

$$d_{\rm crys} = \frac{K\lambda}{\sigma \cos \theta} \tag{3}$$

where K is a constant which depends on the catalyst particles and is generally assumed to be 1, λ is the wavelength of the X-ray radiation, σ is the full width at half maxima (FWHM) of (101) plane, and θ is the position of (101) plane.

TGA was collected on an HCT-1 thermo analyzer (HCT-1, Beijing Hengjiu Company, China) under an air flowing (20 mL/ min). Laser Raman spectra were collected on laser Raman spectrometer (LabRAM Aramis, HORIBA Jobin Yvon, France) with Ar ion laser (514.5 nm). FTIR was recorded on Bruker Vertex 70 infrared spectrometer (Vertex 70, Bruker Corporation, Germany), 4 cm⁻¹ (32 scans). XPS was recorded on ESCALAB 250 spectrometer (ESCALAB 250, Thermo Fisher Scientific Company, USA) with Al Ka radiation. H2-TPR of the sample was recorded on ChemBET-3000 TPR-TPD chemisorption analyzer (ChemBET-3000, Quantachrome, USA) and an online mass spectrum (MS, DYCOR LC-D100, Ametek Company, USA) was used to record the concentration of H2 (m/ z = 2) and SO_2 (m/z = 64). NH₃-TPD was also recorded on the ChemBET-3000 TPR-TPD chemisorption analyzer. The signal of NH_3 (m/z = 16) was recorded by online MS.

In situ DRIFTS was performed on Bruker Vertex 70 infrared spectrometer (Vertex 70, Bruker Corporation, Germany) with mercury-cadmium-telluride (MCT) detector. Powder sample in the reaction cell (Horizon, Harrick Scientific, USA) was treated under N_2 at 450°C for 1 hr to remove the surface impurities. Then reaction cell was cooled to 300°C. Background spectrum was recorded. Afterwards, 2000 ppmV NH_3/N_2 was inlet to the reaction cell for 15 min and then purged with N_2 for another 15 min.

2. Results and discussion

2.1. Physicochemical property of materials

The specific surface area ($S_{\rm BET}$), total pore volume ($V_{\rm p}$) and average pore diameter ($D_{\rm p}$) of the catalyst calcined at different temperature are shown in Table 1. $S_{\rm BET}$ of CuSTi sample calcined at 500°C was 46.2 m²/g with $V_{\rm p}$ of 0.257 cm³/g and $D_{\rm p}$ of 11.1 nm. When the calcination temperature increased to 600°C, $S_{\rm BET}$ and $V_{\rm p}$ of CuSTi-600 sample decreased to 29.0 m²/g and 0.212 cm³/g, about 63% and 82% of CuSTi sample. Meanwhile, $D_{\rm p}$ of CuSTi-600 sample increased to 14.6 nm. The calcination temperature of 700°C had a significantly negative effect on the pore structure of the catalyst. $S_{\rm BET}$ of CuSTi-700 sample decreased sharply to 4.6 m²/g, only 10% of CuTiS sample and $V_{\rm p}$ decreased to 0.062, about 24% of CuTiS sample. However, $D_{\rm p}$ of CuSTi-700 sample increased to 27.1 nm, 2.4

Table 1 $-$ Textural property of the catalyst sample.						
Sample	S_{BET} (m ² /g)	V_p (cm ³ /g)	D _p (nm)	T _{acid} (mmol/g)		
CuSTi	46.2	0.257	11.1	3.65		
CuSTi-600	29.0	0.212	14.6	1.50		
CuSTi-700	4.6	0.062	27.1	/		

 $S_{\rm BET}$: specific surface area obtained at relative pressure (P/ $P_0)=0.05-0.30;\ V_p$: total pore volume obtained at P/P $_0=0.99;\ D_p$: Barrett-Joyner-Halenda (BJH) pore diameter calculated from the N_2 desorption branch; $T_{\rm acid}$: total amount of acid sites calculated by the integral area of NH $_3$ desorption peak from temperature-programmed desorption of NH $_3$ (NH $_3$ -TPD). CuSTi, CuSTi-600 and CuSTi-700 refer to fresh catalyst, catalysts calcined at 600°C and catalysts calcined at 700°C, respectively.

times of CuSTi sample. It could be concluded that higher calcination temperature could destroy the pore structure of catalysts and lead to the sharply decrease of specific surface area. The low surface area would result in decrement of active sites, which was harmful to catalytic activity.

The effects of calcination temperature on crystal structure of the catalyst are shown in Fig. 1. Diffraction peaks assigned to anatase TiO2 could be found on CuSTi sample. Besides, a small peak at $2\theta = 34.3^{\circ}$ which was assigned to CuSO₄ appeared, suggesting that CuSO₄ had a good dispersion on the carrier. The results of scanning transmission electron microscope (STEM)-mapping (Appendix A Fig. S1) also indicated that CuSO₄ had a good dispersion. As shown in Appendix A Fig. S1, S elements presented a good dispersion on the catalyst surface. After the calcination temperature reached 600°C, CuSTi-600 sample still showed diffraction peaks of anatase, but intensities of the peaks increased. According to Scherrer Equation, mean crystallite size (d_{crys} , based on FWHM of the peak at $2\theta = 25.3^{\circ}$) of anatase for CuSTi sample was 21.8 nm, similar to other researches (Miszczak and Pietrzyk, 2015). However, dcrvs of CuSTi-600 sample increased to 26.5 nm, suggesting that calcined at 600°C caused sintering of the carrier. The small peak assigned to CuSO₄ disappeared on CuSTi-600 sample, implying that part of CuSO₄ decomposed. After the catalyst was calcined at 700°C, the diffraction peaks changed significantly. Though the diffraction peaks assigned to anatase could be detected, intensities of the peaks decreased largely. Diffraction peaks of rutile were clear on CuSTi-700 sample, suggesting that rutile was the main crystal form of TiO2. Among the crystal forms of TiO2, anatase was not thermodynamically stable and would transform into rutile under high temperature (Miszczak and Pietrzyk, 2015). There were no peaks assigned to CuSO₄ and new peaks at $2\theta = 35.5^{\circ}$ and 48.7° which could be associated with CuO appeared, implying that CuSO₄ totally decomposed (Eq. (4)). The TGA results of CuSO₄/TiO₂ catalysts also prove this conclusion. As shown in Appendix A Fig. S2, the loss of mass start from 580°C and end at 790°C with a mass loss of 4.8%, which could be associated with decomposition of CuSO₄ on the sample (the theoretical

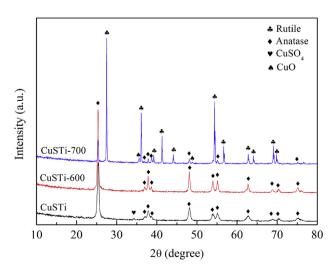


Fig. 1 - X-ray diffraction (XRD) of the catalyst calcined at different temperatures.

value was 5.0%). High calcination temperature would cause the sintering and anatase-to-rutile transformation of the carrier, destroying the pore structure of the catalyst, thus led to significant decrease of specific surface area.

$$CuSO_4 \rightarrow CuO + O_2 + SO_2 \tag{4}$$

The influence of calcination temperature on morphology of the catalyst was investigated by SEM and the photos are shown in Fig. 2. The particles on $CuSO_4/TiO_2$ catalyst calcined at $500^{\circ}C$ were sphere shape with uniform size and presented a good distribution. When the calcination temperature increased to $600^{\circ}C$, it could be found that agglomeration appeared and the size of the sphere particles increased largely, indicating that sintering of the catalyst should occur, in accordance with the result of XRD. The agglomeration and

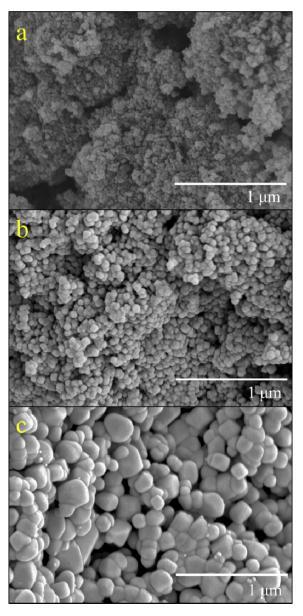


Fig. 2 – Scanning electron microscope (SEM) photos of the (a) CuSTi, (b) CuSTi-600 and (c) CuSTi-700 catalysts.

accumulation were more serious on CuSTi-700 sample and the particles were inhomogeneous, which could destroy the pore structure and cause the decrease of surface area and pore volume. Moreover, the serious agglomeration and accumulation of surface particles would hinder the mass transfer and diffusion of reactant and product gases (Xu et al., 2018; Zhao et al., 2018), which could lead to the decrease of catalytic activity.

Raman spectra of catalysts after normalization are shown in Fig. 3. CuSTi sample showed three main peaks at 391, 515, and 636 cm⁻¹ and a small peak at 792 cm⁻¹, which were the typical Raman peaks of TiO2 (Yao et al., 2017; Ceballos-Chuc et al., 2018). CuSTi sample also presented another small peak at 1094 cm $^{-1}$, which could be assigned to v_{asym} (SO) of sulfates (Ramis et al., 1996). The peaks assigned to TiO2 didn't change after the calcination temperature increased to 600°C. However, the peak at 1094 cm⁻¹ assigned to sulfate was almost disappear, implying part of sulfates decomposed at 600°C, consistent with the results of XRD and TGA. After CuSO₄/TiO₂ was calcined at 700°C, the peaks assigned to TiO₂ changed remarkably, suggesting high temperature had a large influence on the carrier, and the anatase-to-rutile transformation should occur (Nova et al., 2001). FTIR was also used to investigate the influence of calcination temperature on CuSO₄/TiO₂ catalyst and the results are presented in Appendix A Fig. S3. A broad peak in the range from 900 to 400 cm⁻¹ on CuSTi sample could be assigned to anatase TiO2 (Yu et al., 2019a). The peaks at 981, 1062, 1141 cm⁻¹ and a shoulder peak at 1220 cm⁻¹ could be assigned to surface sulfate and bisulfate (Yu et al., 2019a). With the increase of calcination temperature, the broad peak assigned to anatase TiO2 split into two peaks, suggesting that sintering and the anatase-torutile transformation of TiO2 occurred. The intensity of peaks at 981, 1062, 1141 and 1220 cm⁻¹ decreased significantly with the increase of calcination temperature, indicating sulfate and bisulfate would decompose under high temperature. From the results of Raman, FTIR and XRD, it could be concluded that higher calcination temperature could cause the decomposition of CuSO₄, sintering and the anatase-torutile transformation of TiO₂.

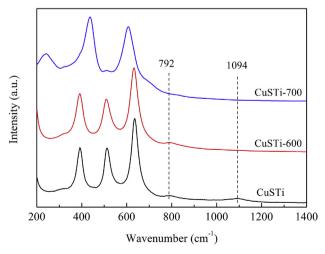


Fig. 3-Raman spectra of the catalyst calcined at different temperatures.

XPS spectra of S 2p and O 1s for the catalyst calcined at different temperatures are presented in Fig. 4. On CuSTi sample, three peaks centered at 166.5, 168.6 and 170.0 eV were found, which could be assigned to sulfite, sulfate and bisulfate, respectively (Yang et al., 2011; Huang et al., 2014). When the calcination temperature increased to 600°C, it could be found that the peak assigned to sulfite disappeared, suggesting that sulfite on the catalyst should decompose. No peaks assigned to sulfite or sulfate on CuSTi-700 sample could be found, implying that all CuSO₄ decomposed after calcined at 700°C, in accordance with the results of XRD and Raman spectra. There were three O species on CuSTi sample (Table 2), which were lattice oxygen (O_{α} , 530.0 eV), surface OH group (O_{β} , 531.6 eV) and adsorbed oxygen (O_{γ} , 533.4 eV) (Yu et al., 2017; Jian et al., 2019). After the catalyst calcined at 600°C, the concentration of surface OH group decreased from 45.1% to 35.0%. The concentration of surface OH group was only 16.1% on CuSTi-700 sample. It could be concluded that high calcination temperature would reduce the concentration of surface OH group on the catalyst. According to previous researches, metal sulfate could form S-OH on the catalyst surface. The decomposition of metal sulfate should be one of the main reasons for the concentration decrease of surface OH group. Because surface OH group could act as Brønsted acid sites (Yu et al., 2017), the high calcination temperature should damage the Brønsted acid sites of the catalyst.

Acid sites play an important role in NH₃-SCR reaction (Zha et al., 2018). NH₃-TPD was used to characterize the acid sites of each sample and the results can be found in Fig. 5. CuSTi sample presented three peaks at 175, 261 and 321°C, which should be assigned to weak, medium and strong acid sites, respectively. According to the amount of NH₃ desorption during NH₃-TPD experiment, the total amount of acid sites (T_{acid}) was calculated and it was 3.65 mmol/g for CuSTi sample. When the calcination temperature increased to 600°C, it could be found that the amount of NH₃ desorption decreased largely and T_{acid} was 1.50 mmol/g, about 41% of CuSTi sample. There was little amount of NH₃ desorption on CuSTi-700 sample, suggesting that the acid sites almost disappeared after the catalyst calcined at 700°C. High calcination

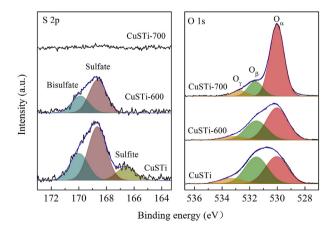


Fig. 4 – S 2p and O 1s X-ray photoelectron spectroscopy (XPS) spectra of the prepared catalysts. O_{α} : lattice oxygen; O_{β} : surface OH group; O_{γ} : adsorbed oxygen.

Table 2 - Content of oxygen species in synthesized catalysts (%).

Sample	Οα	Οβ	Ογ
CuSTi	45.6 (530.1)	45.1 (531.6)	9.3 (533.4)
CuSTi-600	58.5 (530.1)	35.0 (531.6)	6.5 (533.3)
CuSTi-700	77.6 (530.1)	16.1 (531.7)	6.3 (532.8)

Data in parentheses means the position of XPS peak (eV).

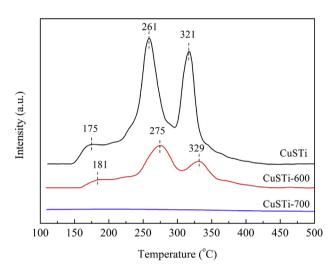
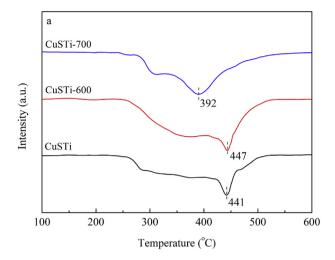


Fig. 5 - Temperature-programmed desorption of NH $_3$ (NH $_3$ -TPD) results of the prepared catalysts.

temperature would cause decomposition of $CuSO_4$ and the decrease of specific surface area, which should be the main reason for the amount decrease of acid sites.

The surface reducibility of catalyst has a significant effect on NH₃-SCR reaction (Cao et al., 2019; Leistner et al., 2018). The effect of calcination temperature on the surface reducibility was characterized by H2-TPR and the results are shown in Fig. 6a. In addition, the emission of SO₂ during H₂-TPR experiment was detected by MS and the results can be found in Fig. 6b. H₂ consumption started from 250°C and ended at 550°C on CuSTi sample. The signal of SO₂ was almost parallel to H₂ consumption. Compared with CuSTi sample, H₂ consumption of CuSTi-600 sample at low temperature decreased obviously, but the signal at high temperature was similar to CuSTi sample. The signal of SO2 also presented the same trend, implying part of sulfate and sulfite decomposed when the catalyst calcined at 600°C, consistent with the results of XPS. However, the H₂ consumption of CuSTi-700 was significantly different from CuSTi sample and there was no signal of SO₂ during H₂-TPR process. All CuSO₄ in the catalyst should decompose into CuO and the formation of CuO caused the change of surface reducibility on the catalyst calcined at

The results for in situ DRIFTS of $\rm NH_3$ over each sample under 300°C are shown in Fig. 7. Several bonds at 1319, 1388, 1442, 1601, 3182, 3280, 3351 cm⁻¹ and a wide band from 1645 to 1855 cm⁻¹ were found on $\rm CuSO_4/\rm TiO_2$ catalysts. The bond at 1442 cm⁻¹ and the wide band from 1645 to 1855 cm⁻¹ could be assigned to $\rm NH_4^+$ on Brønsted acid sites (B) (Yu et al., 2017; Ye



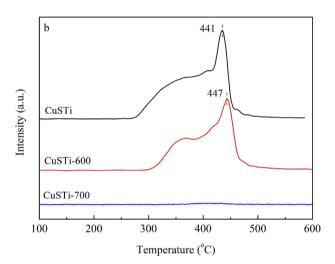


Fig. 6 — (a) Temperature-programmed reduction of H₂ (H₂-TPR) of the catalyst calcined at different temperatures and (b) emission of SO₂ during H₂-TPR process.

et al., 2019). Meanwhile, the bonds at 1319 and 1601 cm^{-1} could be assigned to NH₃ on Lewis acid sites (L) (Yu et al., 2017; Ye et al., 2019). Our previous work indicated that both Brønsted and Lewis acid sites participated in NH₃-SCR reaction (Yu et al., 2017). At high wavenumbers, the band at 3182 cm⁻¹ could be attributed to N-H of the coordinated ammonia on the catalyst surface and the bands at 3280 and 3351 cm⁻¹ could be attributed to ammonium ions (Ma et al., 2014; Fan et al., 2018). After the catalyst was calcined at 600°C, it could be found that the intensity of the bonds assigned to Brønsted and Lewis acid sites decreased largely, suggesting that the amount of acid sites reduced, in accordance with the results of NH₃-TPD. The decrease of specific surface area and the decomposition of sulfite or sulfate should be the main reasons for the reduction of acid sites. The amount ratio of Brønsted to Lewis acid sites on the sample was calculated by the ratio of the peak intensity at 1442 cm^{-1} to the peak intensity at 1601 cm^{-1} . The ratio was 1.10 for CuSTi sample and 0.59 for CuSTi-600 sample, suggesting that the influence of calcination temperature on Brønsted acid sites was larger than Lewis acid sites. The

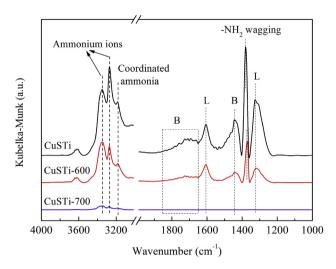


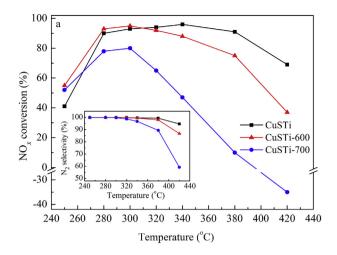
Fig. 7 – In situ diffuse reflectance infrared Fourier transform spectroscopy (DRIFTS) of NH₃ on CuSTi, CuSTi-600 and CuSTi-700 sample at 300°C. B: Brønsted acid sites; L: Lewis acid sites.

decomposition of sulfite or sulfate at high calcination temperature would decrease the amount of surface S–OH, which was the main Brønsted acid sites on sulfate catalyst (Yu et al., 2017). When the calcination temperature increased to 700°C, almost all bonds disappeared, suggesting that high calcination temperature would destroy the acid sites on the catalyst, in accordance with the results of NH₃-TPD. The decomposition of CuSO₄ and the anatase transform into rutile should be the main reasons for the disappears of acid sites.

2.2. Catalytic activity

Fig. 8a presents NO_x conversion of the catalyst calcined at different temperatures. N_2 selectivity was also calculated and the result can be found in the insert of Fig. 8a. CuSTi sample performed high activity during $280-380^{\circ}$ C with NO_x conversion higher than 90% and N_2 selectivity was more than 99% during $280-380^{\circ}$ C. After calcined at 600° C, the NO_x conversion of CuSTi-600 sample decreased significantly at temperature higher than 340° C, suggesting that the calcination temperature of 600° C could impact catalytic activity to a certain extent. However, when the calcination temperature reached 700° C, the activity of CuSTi-700 sample decreased sharply and NO_x conversion even dropped to a negative value at 420° C. In addition, N_2 selectivity decreased from 94.6% to 59.1% at 420° C, suggesting that high calcination temperature would have a negative effect on N_2 selectivity of the catalyst.

The negative value of NO_x conversion for CuSTi-700 sample at 420°C should mean that NH_3 was oxidized to NO_x by O_2 over the sample. In order to confirm this conclusion, the oxidation of NH_3 by O_2 over each sample was tested and the results can be found in Fig. 8b. It could be found that the calcination temperature had a large effect on NH_3 conversion over the catalyst. The oxidation of NH_3 over CuSTi sample under the temperature lower than 380°C was hard to occur and the NH_3 conversion was only 4.1% at 420°C. When the calcination temperature increased to 600°C, the NH_3 conversion at



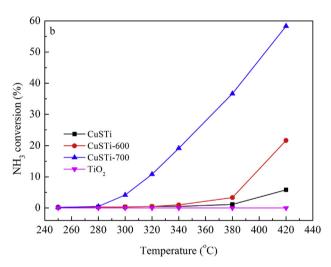
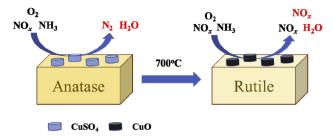


Fig. 8 – (a) NO_x conversion and (inset) N_2 selectivity of each sample and (b) NH_3 conversion over the catalyst under NH_3 oxidation conditions. Reaction conditions: 700 ppmV NO (absent for (b)), 700 ppmV NH_3 , 5.0 vol.% H_2O and 4.0 vol.% O_2 in N_2 , gas hourly space velocity 60,000 hr^{-1} .

temperature higher than 340°C increased significantly. After the catalyst was calcined at 700°C, it could be found that the NH₃ conversion increased sharply with temperature and reached 58.7% at 420°C. The oxidation of NH₃ on pure TiO₂ calcined at 700°C was also tested and the result can be found in Fig. 8b. It could be found that there was little NH₃ oxidized to NO_x on pure TiO₂. Thus, it can be concluded that NH₃ should be oxidized by O₂ over CuO of CuSTi-700 sample in NH₃-SCR reaction process and formed new NO_x at high temperature, which caused the sharply decrease of NO_x conversion (Scheme 1). According to other researches, Brønsted acid sites were not necessary for ammonia oxidation and Lewis acid sites were active for ammonia oxidation to nitrogen oxides (Si et al., 2010; Yu et al., 2014). The possible reaction routes for ammonia oxidation on Lewis acid sites are presented below.

$$NH_3(g) + Lewis acid sites \rightarrow NH_3(ads)$$
 (5)

$$NH_3(ads) \rightarrow NH_2(ads) + H^+ + e^-$$
 (6)



Scheme $1 - NH_3$ -selective catalytic reduction reaction over the catalyst calcined at different temperatures.

$$2NH2(ads) + O(ads) \rightarrow 2NH(ads) + H2O$$
 (7)

$$NH(ads) + O(ads) \rightarrow HNO(ads)$$
 (8)

$$2HNO(ads) + O(ads) \rightarrow H_2O + 2NO$$
 (9)

where (g) and (ads) refer to gaseous phase and adsorption state, respectively.

As shown in Fig. 7, with the increase of calcination temperature, the amount of Brønsted acid sites decreased largely due to the decomposition of CuSO_4 . Thus, the ammonia oxidation activity of the catalyst calcined at high temperature should be increased and considerable ammonia can't participate in $\text{NH}_3\text{-SCR}$ reaction, which caused a decrease of NO_x conversion.

3. Conclusions

The calcination temperature had a large effect on physicochemical properties and activity of $\text{CuSO}_4/\text{TiO}_2$ catalysts. Higher calcination temperature would cause sintering and the anatase transform into rutile, which destroyed the pore structure and thus largely decreased the specific surface area. The decomposition of CuSO_4 under high calcination temperature led to the decrease of Brønsted acid sites amount and had an adverse effect on surface acidity, which could cause the decrease of catalytic activity. Moreover, CuO from the decomposition of CuSO_4 under higher calcination temperature favored the oxidation of NH_3 , causing the significant decrease of SCR activity.

Conflict of interest

The authors declared that they have no conflicts of interest to this work.

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Appendix A. Supplementary data

Supplementary data to this article can be found online at https://doi.org/10.1016/j.jes.2020.01.010.

REFERENCES

- Cao, L., Wu, X.D., Xu, Y.F., Lin, Q.W., Hu, J.F., Chen, Y., et al., 2019. Ceria-modified WO₃-TiO₂-SiO₂ monolithic catalyst for high-temperature NH₃-SCR. Catal. Commun. 120, 55–58.
- Ceballos-Chuc, M.C., Ramos-Castillo, C.M., Alvarado-Gil, J.J., Oskam, G., Rodríguez-Gattorno, G., 2018. Influence of brookite impurities on the Raman spectrum of TiO₂ anatase nanocrystals. J Phys. Chem. C 122, 19921–19930.
- Chen, C.M., Gao, Y., Liu, S.T., Chen, J.M., Jia, W.B., 2018a. Review on the latest developments in modified vanadium-titanium-based SCR catalysts. Chin. J. Catal. 39, 1347—1365.
- Chen, Y.X., Li, C., Chen, J.X., Tang, X.F., 2018b. Self-prevention of well-defined-facet Fe₂O₃/MoO₃ against deposition of ammonium bisulfate in low-temperature NH₃-SCR. Environ. Sci. Technol. 52, 11796–11802.
- Cheng, J., Ye, Q., Zheng, C.K., Cheng, S.Y., Kang, T.F., Dai, H.X., 2019. Effect of ceria loading on Zr-pillared clay catalysts for selective catalytic reduction of NO with NH₃. New J. Chem. 43, 10850–10858.
- Du, X.S., Wang, X.M., Chen, Y.R., Gao, X., Zhang, L., 2016. Supported metal sulfates on Ce-TiO $_{\rm X}$ as catalysts for NH $_3$ -SCR of NO: High resistances to SO $_2$ and potassium. J. Ind. Eng. Chem. 36, 271–278.
- Fan, J., Ning, P., Song, Z.X., Liu, X., Wang, L.Y., Wang, J., et al., 2018. Mechanistic aspects of NH_3 -SCR reaction over CO_2/TiO_2 -Zr O_2 -S O_2^{4-} catalyst: In situ DRIFTS investigation. Chem. Eng. J. 334, 855–863.
- Gao, F.Y., Tang, X.L., Yi, H.H., Li, J.Y., Zhao, S.Z., Wang, J.G., et al., 2017. Promotional mechanisms of activity and SO_2 tolerance of Co- or Ni-doped MnO_x -CeO $_2$ catalysts for SCR of NO_x with NH_3 at low temperature. Chem. Eng. J 317, 20–31.
- Hammershøi, P.S., Jensen, A.D., Janssens, T.V.W., 2018. Impact of SO₂-poisoning over the lifetime of a Cu-CHA catalyst for NH₃-SCR. Appl. Catal. B-Environ. 238, 104–110.
- He, G.Z., Lian, Z.H., Yu, Y.B., Yang, Y., Liu, K., Shi, X.Y., et al., 2018. Polymeric vanadyl species determine the low-temperature activity of V-based catalysts for the SCR of NO_x with NH_3 . Sci. Adv. 4 eaau4637.
- Huang, H.L., Lan, Y., Shan, W.P., Qi, F.H., Xiong, S.C., Liao, Y., et al., 2014. Effect of sulfation on the selective catalytic reduction of NO with NH_3 over γ -Fe₂O₃. Catal. Lett. 144, 578–584
- Jian, Y.F., Yu, T.T., Jiang, Z.Y., Yu, Y.K., Douthwaite, M., Liu, J.Y., et al., 2019. In-depth understanding of the morphology effect of α -Fe₂O₃ on catalytic ethane destruction. ACS Appl. Mater. Interfaces 11, 11369–11383.
- Leistner, K., Kumar, A., Kamasamudram, K., Olsson, L., 2018. Mechanistic study of hydrothermally aged Cu/SSZ-13 catalysts for ammonia-SCR. Catal. Today 307, 55–64.

- Lian, Z.H., Shan, W.P., Wang, M., He, H., Feng, Q.C., 2019. The balance of acidity and redox capability over modified CeO_2 catalyst for the selective catalytic reduction of NO with NH₃. J. Environ. Sci. 79, 273–279.
- Ma, L., Li, J.H., Fu, L.X., 2011. Catalytic performance, characterization, and mechanism study of $Fe_2(SO_4)_3/TiO_2$ catalyst for selective catalytic reduction of NO_x by ammonia. J. Phys. Chem. C 115, 7603–7612.
- Ma, L., Cheng, Y.S., Cavataio, G., McCabe, R.W., Fu, L.X., Li, J.H., 2014. In situ DRIFTS and temperature-programmed technology study on NH₃-SCR of NOx over Cu-SSZ-13 and Cu-SAPO-34 catalysts. Appl. Catal. B-Environ. 156–157, 428–437.
- Miszczak, S., Pietrzyk, B., 2015. Anatase-rutile transformation of ${\rm TiO_2}$ sol-gel coatings deposited on different substrates. Ceram. Int. 41, 7461–7465.
- Nova, I., dall'Acqua, L., Lietti, L., Giamello, E., Forzatti, P., 2001. Study of thermal deactivation of a de-NOx commercial catalyst. Appl. Catal. B-Environ. 35, 31–42.
- Odenbrand, C.U.I., 2018. CaSO $_4$ deactivated V $_2$ O $_5$ -WO $_3$ /TiO $_2$ SCR catalyst for a diesel power plant. Characterization and simulation of the kinetics of the SCR reactions. Appl. Catal. B-Environ. 234, 365–377.
- Ramis, G., Yi, L., Busca, G., 1996. Ammonia activation over catalysts for the selective catalytic reduction of NO_x and the selective catalytic oxidation of NH_3 . An FT-IR study. Catal. Today 28, 373–380.
- Si, Z.C., Weng, D., Wu, X.D., Li, J., Li, G., 2010. Structure, acidity and activity of CuO_x/WO_x-ZrO₂ catalyst for selective catalytic reduction of NO by NH₃. J. Catal. 271, 43–51.
- Wang, C., Yu, F., Zhu, M.Y., Tang, C.G., Zhang, K., Zhao, D., et al., 2019. Highly selective catalytic reduction of NO_x by MnO_x - CeO_2 - Al_2O_3 catalysts prepared by self-propagating high-temperature synthesis. J. Environ. Sci. 75, 124–135.
- Wijayanti, K., Leistner, K., Chand, S., Kumar, A., Kamasamudram, K., Currier, N.W., et al., 2016. Deactivation of Cu-SSZ-13 by SO_2 exposure under SCR conditions. Catal. Sci. Technol. 6, 2565–2579.
- Woo, J.W., Bernin, D., Ahari, H., Shost, M., Zammit, M., Olsson, L., 2019. Understanding the mechanism of low temperature deactivation of Cu/SAPO-34 exposed to various amounts of water vapor in the NH₃-SCR reaction. Catal. Sci. Technol. 9, 2623–3636
- Wu, M.Z., Zhan, W.C., Guo, Y., Wang, Y.S., Guo, Y.L., Gong, X.Q., et al., 2016. Solvent-free selective oxidation of cyclohexane with molecular oxygen over manganese oxides: Effect of the calcination temperature. Chin. J. Catal. 37, 184–192.
- Wu, Y.H., Chu, B.X., Zhang, M., Yi, Y.N., Dong, L.H., Fan, M.G., et al., 2019. Influence of calcination temperature on the catalytic properties of $LaCu_{0.25}Co_{0.75}O_3$ catalysts in NO_x reduction. Appl. Surf. Sci. 481, 1277–1286.
- Xu, L.T., Niu, S.L., Lu, C.M., Zhang, Q., Li, J., 2018. Influence of calcination temperature on $Fe_{0.8}Mg_{0.2}O_z$ catalyst for selective catalytic reduction of NO_x with NH_3 . Fuel 219, 248–258.
- Yan, Z.D., Shi, X.Y., Yu, Y.B., He, H., 2019. Alkali resistance promotion of Ce-doped vanadium-titanic-based NH₃-SCR catalysts. J. Environ. Sci. 73, 155–161.
- Yang, S.J., Guo, Y.F., Yan, N.Q., Wu, D.Q., He, H.P., Qu, Z., et al., 2011. Nanosized cation-deficient Fe-Ti spinel: a novel magnetic sorbent for elemental mercury capture from flue gas. ACS Appl. Mater. Interfaces 3, 209–217.
- Yao, X.J., Zhao, R.D., Chen, L., Du, J., Tao, C.Y., Yang, F.M., et al., 2017. Selective catalytic reduction of NO_x by NH₃ over CeO₂ supported on TiO₂: Comparison of anatase, brookite, and rutile. Appl. Catal. B-Environ. 208, 82–93.
- Ye, D., Ren, X.Y., Qu, R.Y., Liu, S.J., Zheng, C.H., Gao, X., 2019. Designing SO₂-resistant cerium-based catalyst by modifying with Fe₂O₃ for the selective catalytic reduction of NO with NH₃. Mol. Catal. 462, 10–18.

- Yu, T., Wang, J., Huang, Y., Shen, M.Q., Li, W., Wang, J.Q., 2014. NH₃ oxidation mechanism over Cu/SAPO-34 catalysts prepared by different methods. ChemCatChem 6, 2074–2083.
- Yu, Y.K., Miao, J.F., Wang, J.X., He, C., Chen, J.S., 2017. Facile synthesis of CuSO₄/TiO₂ catalysts with superior activity and SO₂ tolerance for NH₃-SCR: physicochemical properties and reaction mechanism. Catal. Sci. Technol. 7, 1590–1601.
- Yu, Y.K., Chen, C.W., Ma, M.D., Douthwaite, M., He, C., Miao, J.F., et al., 2019a. SO₂ promoted in situ recovery of thermally deactivated Fe₂(SO₄)₃/TiO₂ NH₃-SCR catalysts: From experimental work to theoretical study. Chem. Eng. J. 361, 820–829.
- Yu, Y.K., Chen, C.W., He, C., Miao, J.F., Chen, J.S., 2019b. In situ growth synthesis of CuO@Cu-MOFs core-shell materials as novel low-temperature NH₃-SCR catalysts. ChemCatChem 11, 979—984
- Zha, K.W., Kang, L., Feng, C., Han, L.P., Li, H.R., Yan, T.T., et al., 2018. Improved $\rm NO_x$ reduction in the presence of alkali metals by using hollandite Mn-Ti oxide promoted Cu-SAPO-34 catalysts. Environ. Sci.: Nano 5, 1408–1419.
- Zhao, Y.J., Guo, Z.Y., Zhang, H.J., Peng, B., Xu, Y.X., Wang, Y., et al., 2018. Hydrogenation of diesters on copper catalyst anchored on ordered hierarchical porous silica: Pore size effect. J. Catal. 357, 223–237.